This document is the accepted manuscript version of a published work that appeared in final form after peer review and technical editing by the publisher. To access the final edited and published work see link to "View at Publisher (DOI)" at the DORA Empa web page (https://www.dora.lib4ri.ch/empa/).

Systematic Exploration of Biotransformation Reactions of Amine-containing Micropollutants in Activated Sludge

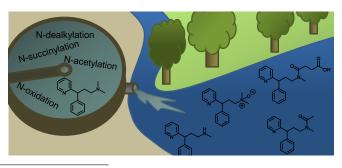
Rebekka Gulde,^{†,‡} Ulf Meier,[†] Emma L. Schymanski,[†] Hans-Peter E. Kohler,^{†,‡} Damian E. Helbling,[¶] Samuel Derrer,[†] Daniel Rentsch,[§] and Kathrin Fenner^{*,†,‡}

Eawag, Swiss Federal Institute of Aquatic Science and Technology, 8600 Dübendorf,
Switzerland, Department of Environmental Systems Science (D-USYS), ETH Zürich, 8092
Zürich, Switzerland, School of Civil and Environmental Engineering, Cornell University, Ithaca,
NY, USA, and EMPA, Swiss Federal Laboratories for Materials Science and Technology, 8600
Dübendorf, Switzerland

E-mail: kathrin.fenner@eawag.ch

Phone: +41 58 765 5085. Fax: +41 58 765 5802

Abstract



^{*}To whom correspondence should be addressed

2

[†]Eawag, Swiss Federal Institute of Aquatic Science and Technology, 8600 Dübendorf, Switzerland

[‡]Department of Environmental Systems Science (D-USYS), ETH Zürich, 8092 Zürich, Switzerland

[¶]School of Civil and Environmental Engineering, Cornell University, Ithaca, NY, USA

[§]EMPA, Swiss Federal Laboratories for Materials Science and Technology, 8600 Dübendorf, Switzerland

The main removal process for polar organic micropollutants during activated sludge treatment is biotransformation, which often leads to the formation of stable transformation products (TPs). As the analysis of TPs is challenging, the use of pathway prediction systems can help by generating a list of suspected TPs. To complete and refine pathway prediction, comprehensive biotransformation studies for compounds exhibiting pertinent functional groups under environmentally relevant conditions are needed. As many polar organic micropollutants present in wastewater contain one or several amine functional groups, we systematically explored amine biotransformation by conducting experiments with 19 compounds that contained 25 structurally diverse primary, secondary, and tertiary amine moieties. Identification of 144 TP candidates and structure elucidation of 101 of these resulted in a comprehensive view on initial amine biotransformation reactions. The reactions with the highest relevance were N-oxidation, N-dealkylation, N-acetylation, and N-succinylation. While many of the observed reactions were similar to those known for the mammalian metabolism of amine-containing xenobiotics, some N-acylation reactions were not previously described. In general, different reactions at the amine functional group occurred in parallel. Finally, recommendations on how these findings can be implemented to improve microbial pathway prediction of amine-containing micropollutants are given.

Introduction

8

10

11

12

13

14

15

16

17

18

19

Organic micropollutants (MPs) such as the active ingredients of pharmaceuticals, personal care products, and pesticides are conveyed by sanitary sewers or by surface runoff and storm sewers to wastewater treatment plants (WWTPs). There, the main removal process for polar organic MPs is microbial biotransformation in the activated sludge compartment. However, most MPs are not fully mineralized, but rather biotransformed to transformation products (TPs), which are then released with the WWTP effluents into the aquatic environment. Has been shown that some TPs have the potential to be as or even more toxic than their parent MPs, and thus contribute to the toxicological effects of anthropogenic chemicals on aquatic ecosystems. 6–10 To

comprehensively assess the risk of MPs, it is therefore desirable to also determine the occurrence and prevalence of their TPs in environmental water bodies. However, due to their presence at low concentrations in complex environmental matrices, the detection and identification of unknown TPs or MPs in general is challenging. Recently, screening for expected compounds in a so-called suspect screening approach has been successfully applied in different environmental matrices, *e.g.*, wastewater effluents, lake sediments, and surface waters. ^{11–15}

In order to generate a list of expected TPs for suspect screening, pathway prediction systems 35 are helpful. 16,17 Currently freely available systems for the prediction of microbial biotransforma-36 tion reactions for given parent MPs include PathPred, ¹⁸ CRAFT, ¹⁹ the OECD toolbox, ²⁰ and the 37 EAWAG-PPS. The EAWAG-PPS derives from the former University of Minnesota pathway prediction system (UM-PPS), 21,22 and is now hosted by Eawag, the Swiss Federal Institute of Aquatic Science and Technology. 23 The predictions of these tools are mainly based on a set of generalized biotransformation rules extracted from literature-reported, microbially mediated metabolic pathways and enzyme-catalyzed reactions. Biotransformation rules recognize functional groups present in the query compound structure and predict transformation reactions at these functional groups. 21 Currently, rule-based pathway prediction systems are generally sensitive, but of rather poor selectivity when validated with observed TPs from environmental simulation or monitoring studies. This means that the systems predict most of the observed TPs, but also many others. Therefore, the number of false positives increases rapidly, resulting in combinatorial explosion when the rules are applied iteratively to predict consecutive steps of biotransformation reactions. ²⁴ There are two potential reasons. First, the biotransformation rules corresponding to the different functional groups are based on too little microbial biotransformation data and are not refined to account for specific structural factors around the functional group. Second, the 51 majority of available microbial pathways are not specified on environmental conditions but are rather based on pure culture studies or ready biodegradability assays, where the compound of interest serves as sole or major carbon source. Thus, the derived biotransformation rules might include reactions that are unlikely to occur under environmentally relevant conditions, where the MPs that are present in trace concentrations are most likely turned over co-metabolically by mixed microbial communities. ²⁵

Comprehensive biotransformation studies conducted under environmentally relevant conditions 58 for a set of structurally diverse compounds could help overcoming these limitations. A systematic 59 set of initial transformation reactions observed for a given pertinent functional group can be 60 used to improve prediction by refining the specificity of existing biotransformation rules, both in 61 terms of structural features that affect the likelihood of certain types of reactions and by learning 62 about previously not implemented reactions that occur under relevant conditions. Such endeavors 63 have recently become much more accessible through the development of efficient workflows using 64 liquid-chromatography high-resolution tandem mass spectrometry (LC-HR-MS/MS) to screen for suspect and non-target TPs formed in biotransformation studies. ^{26,27} 66

Primary, secondary, and tertiary amine functional groups are highly abundant in wastewater-67 relevant MPs, such as the active ingredients of pharmaceuticals. 28 From the extensive exploration of the mammalian metabolism of xenobiotics, three main reactions that involve the amine functional groups are known, namely (i) N-oxidation, leading to the formation of N-oxides for tertiary amines, and to the formation of hydroxylamines for secondary and primary amines; (ii) 71 α-C-hydroxylation, leading to unstable hemiaminals that are either spontaneously cleaved to a carbonyl and an amine product by a so-called N-dealkylation reaction, oxidized to amides, or dehydrated to iminium species; and (iii) conjugation reactions such as N-acetylations for secondary and primary amines. The corresponding data were assembled by Testa et al. ^{29–33} who elaborated 75 on the corresponding enzymatic systems, mechanistic details, and structural and electronic fea-76 tures that influence the different types of reactions. All three main reactions have been observed 77 in microbial systems. 34 These and additional data on other functional groups suggest that micro-78 bial transformation reactions of MPs are similar to those observed in the mammalian metabolism of xenobiotics, even though individual major products for certain MPs might differ. ^{6,34–36} Overall, it can be hypothesized that the microbial metabolism of amine-containing MPs will be similar to 81 that of mammalian systems.

The objective of our research was to elucidate biotransformation pathways of a set of struc-83 turally diverse amine-containing compounds in an environmentally relevant microbial system to obtain a consistent set of initial microbial transformation reactions involving the amine functional group. The observed reactions should then be rationalized by comparison to mammalian systems and generalized to suggest biotransformation rules for improved prediction of microbial 87 biotransformation of amine-containing MPs. To do so, we performed systematic biotransformation experiments in activated sludge-seeded bioreactors under environmentally relevant conditions with 19 parent compounds that contain altogether 25 different primary, secondary, and tertiary 90 amine moieties. First generation TP candidates were identified by LC-HR-MS/MS-based ap-91 proaches, and the corresponding TP structures were elucidated where possible, with confidence levels assigned as proposed by Schymanski et al. (2014). 37 Reactions were assigned to observed parent-TP pairs and compared with the known mammalian biotransformation reactions of amines. Finally, generalized rules for biotransformation reactions were extracted for tertiary, secondary, and primary amines.

Materials and Methods

The following is a concise presentation of materials and methods; full details are given in the SI.

99 Micropollutant Selection

Environmentally relevant compounds (*e.g.*, pharmaceuticals, pesticides, and other wastewaterderived chemicals) were selected to cover a variety of primary, secondary, and tertiary amine
functional groups, and as few as possible other functional groups that were expected to be readily
biotransformed. The final set of 19 test compounds contained 25 amine moieties, including 13
tertiary, eight secondary, and four primary ones. All names and chemical structures are given in
Table 1. Additionally, five N-oxide TPs (of PHE, VEN, LID, DEP, and PAR) were also included
in the biotransformation experiments.

of Biotransformation Test System

The experimental setup for the biotransformation batch experiments was adopted from Helbling 108 et al. (2010). 27 Activated sludge (3 L) was sampled from the nitrification basin of a half municipal 109 and half industrial WWTP (ARA Neugut, Dübendorf, Switzerland). Reactors (100 mL amber 110 Schott bottles) were filled with 50 mL activated sludge and spiked with 60 µL compound solution 111 (100 mg/L in methanol:water 1:9) to a final concentration of 120 μ g/L for each individual 112 compound. The 19 amine-containing compounds were split into two groups, which were spiked 113 separately in two reactors. Additionally, each parent compound was spiked in individual reactors. 114 One reactor was spiked with a mixture containing the five N-oxide TPs. Samples (1.5 mL) were 115 taken at approximately 0 h, 2 h, 4 h, 8 h, 1 d, 1.5 d, 2 d, 3 d, and 4 d after the start of the 116 experiment and centrifuged. The supernatants (0.5 mL) were stored at 4 °C in the dark until 117 analysis for a maximal duration of 10 days. Additional sorption and abiotic control experiments 118 were conducted to test for abiotic processes. 119

120 Analytical Method

For chemical analysis, reversed-phase liquid chromatography coupled to a high-resolution quadrupole 121 Orbitrap mass spectrometer (Q Exactive, Thermo Scientific) was used. The analytical method 122 from Gulde et al. (2014)³⁸ was adjusted as follows: Three data-dependent MS2 scans were con-123 ducted after each full-scan in positive or positive-negative ionization switch mode. The dynamic 124 exclusion was set to 7 s for acquisition in positive mode and to 3 s for acquisition in positive-125 negative switch mode. Data-dependent MS2 were triggered on all selected parent compounds 126 and a list of candidate TP masses, which was assembled from previously found TPs in similar un-127 published experiments, EAWAG-PPS predictions, 23 metaprint2D predictions, 39 and by manually 128 applying a range of plausible atomic modifications. For all identified TP peaks, additional MS2 129 measurements were conducted later with a Q Exactive or Q Exactive Plus instrument (Thermo 130 Scientific). In individual runs for each TP, six targeted MS2 scans were triggered after each 131 full-scan in positive ionization-mode with normalized collision energies of 15, 30, 45, 60, 75, 90

133 (dimensionless).

ing

135

134 Transformation Product Identification by Suspect and Non-target Screen-

For TP identification, the underlying principles from Helbling et al (2010)²⁷ were applied using Compound Discoverer 1.0 (Thermo Scientific) for suspect and non-target screening and Sieve 137 2.2 (Thermo Scientific) for non-target screening. While we screened for a list of plausible TP 138 candidates with the suspect screening approach, we used the non-target screening approach to 139 discover additional TP candidate peaks. For both approaches, likely TP candidate peaks had 140 to fulfill the following criteria in order to be selected for further analysis: (i) intensity above a 141 set threshold; (ii) reasonable peak shape; (iii) presence in both treated reactors and absence or 142 significantly lower amounts in control reactors; (iv) TP-like time series pattern, i.e. increase or 143 increase and decrease over the time of the experiment; and (v) a reasonable chemical formula 144 derived from the exact mass and the isotopic pattern (Details are given in Chapter S1.4 in the 145 SI). Identified TP candidate peaks were then integrated manually using Xcalibur 2.2 (Thermo 146 Scientific) to yield time series of TP peak areas. Because the overall goal of this study was to 147 explore and interpret the initial biotransformation reactions of amine-containing compounds, we 148 focused our analysis on first generation TPs. Therefore, TPs were omitted from further analysis 149 that were clearly higher generation products according to their time series patterns or initial structural evidence. TPs formed by multiple, sequential reactions would only be kept for further analysis if they pointed towards a specific initial reaction for which no first generation TP was 152 observed.

54 Structure Elucidation

Structure elucidation was conducted either with the Compound Discoverer 1.0 software or manually based on the interpretation of: (i) the exact masses and the isotopic patterns of the positive

and, if available, negative MS spectra to propose chemical formulas; and (ii) the MS2 fragments and neutral losses of the positive and, if available, negative MS2 spectra to propose structural features. Afterwards, the interpretations were combined and further supported by the use of the MassFrontier 7.0 software (Thermo Scientific) to propose plausible TP structures. Apart from the confirmation of some TP structures through purchased or synthesized reference stan-dards, the outcome and confidence of the structural interpretation could vary significantly due to the availability of varying structural evidence. To transparently communicate the confidence in our structural interpretation, we assigned confidence levels as proposed by Schymanski et al. (2014), ³⁷ which ranged from Level 5 – exact mass, over Level 4 – unequivocal molecular formula, Level 3 – tentative candidates, and Level 2 – probable structure, to Level 1 – confirmed structure.

167 Assignment of Biotransformation Reactions

In order to interpret the results in terms of biotransformation reactions, it was necessary to assign reactions to observed parent-TP pairs. However, the certainty with which a biotransformation reaction can be attributed depends on the confidence in the TP structure and whether or not the formation of the respective TP from the corresponding parent compound can unambiguously be attributed to a plausible reaction. We therefore classified the attributed reactions as *certain*, *likely*, *possible*, and *unknown*.

To assess the importance of different reactions in terms of how much of the parent compounds was biotransformed via a specific reaction, we considered the frequency with which a given reaction was observed and the amount of the respective TPs formed. This analysis was carried out separately for tertiary, secondary and primary amines. To assess the *relative frequency* of a reaction, we counted all first-generation TPs that were certain or likely to be formed through that specific reaction and related them to the number of TPs that could potentially have been formed via the same reaction. To assess the amounts of TPs formed, we determined the detected maximal peak area of each TP and related it to the degree of biotransformation of the parent compound at that same time point (*i.e.*, change in the peak area of the biotransformed parent

compound between time zero and the time point of maximal peak area of the TP) to yield a so-called maximal relative amount. This procedure might be fairly uncertain due to the fact 184 that TPs and parent compounds might have different ionization efficiencies. However, since 185 the comparison of ionization efficiencies of 15 parent-TP pairs that were transformed by N-186 dealkylation and N-oxidation showed that the average difference in ionization efficiency was 1.5 187 (maximal: 3.1; Details are given in Chapter S10 in the SI) it was considered sufficient to yield 188 a rough estimate of the relative amount of TPs formed. To evaluate the overall importance of 189 a given type of reaction, we summed the maximal relative amounts of all first-generation TPs 190 that were certain or likely to be formed through the reaction of interest and divided it by the 191 number of TPs that could have been formed through this reaction, resulting in a so-called mean 192 maximal relative amount. It should be noted that the actual maximal relative amount of a TP 193 can be much higher than the observed TP areas suggest since the TP might itself be further 194 transformed to some extent. Therefore, we additionally characterized the time series patterns 195 as rising, rising and steady, or rising and falling to qualitatively assess the stability of the TPs. 196 However, we cannot fully exclude the occurrence of transformation reactions whose resulting TPs 197 are so quickly further transformed that we are not able to observe them. 198

Results and Discussion

In total, 144 TPs were identified for the 19 investigated parent amines. Of these, 43 were 200 omitted since they were likely higher generation TPs. For these, information on exact mass, 201 retention time, and proposed elemental formula are given in Table S11 in the SI. The structures 202 of the remaining 101 TPs were elucidated and confidence levels were assigned. As a result, the 203 structural confidence could be considered confirmed (Level 1) for 21 TPs, probable (Level 2b) for 204 nine TPs, tentative (Level 3) for 68 TPs, whereas for three TPs only an unequivocal molecular 205 formula (Level 4) could be assigned. The evidence for structure elucidation is given in Chapter 206 S4 in the SI and is electronically available in the MassBank database. 40 207

The biotransformation reactions that were assigned to the observed parent-TP pairs for reactions involving the amine functional group are listed in Table 1 together with their corresponding certainty, structural confidence level, *maximal relative amount*, degree of biotransformation of the parent compound at the time point where TP was maximal, and characterization of time series patterns. The same is given in Chapter S8 in the SI for the remaining observed reactions that did not involve an amine moiety. Related further information can be found in Chapter S6 in the SI, which contains maps of the biotransformation pathways for each parent compound, area time series of the parent compounds and their TPs, and the observed concentration time series for the parent compound in the abiotic control, the sorption control, and the biotransformation reactors.

For 18 parent amines (all except CLE), multiple biotransformation reactions were observed in parallel. No TPs were found for CLE, despite the presence of a secondary and primary amine and a decrease in parent concentration within the range of the other parent amines. For the remaining 18 parent compounds, between one and nine reactions that possibly involve the amine functional group were observed to occur in parallel. Besides those, additional reactions that did not involve the amine functional group were observed. This is in contrast to previous observations for amides, where for primary and secondary amides mostly only one type of reaction, namely hydrolysis, and only for tertiary amides multiple reactions were observed. ²⁶

The two principal initial oxidative reactions at the amine functional group that are known from mammalian metabolism of xenobiotics were also observed for the activated sludge system, namely (i) N-oxidation, and (ii) α -C-hydroxylation followed by N-dealkylation, oxidation to amide, or dehydration to iminium species. Additionally, several N-acylation reactions, namely N-formylation, N-acetylation, N-propionylation, N-malonylation, N-succinylation and presumably desaturation to N-fumarylated products were observed, of which only N-acetylation is known from the mammalian metabolism of xenobiotics. These reactions will be discussed in detail in the following sections. A schematic illustration of all reactions involving the amine functional group is given in Chapter S3 in the SI.

Additional observed reactions that did not involve the amine functional groups were O-demethylation, hydroxylation at groups other than the amine-N or the α -C, desaturation, oxidation of a hydroxylated product to a carbonyl or carboxylic acid product, dehydration of an aliphatic alcohol to an alkene, and combinations of several reactions including decarboxylation and β -oxidation. For some TPs, no plausible reactions leading to their formation could be assigned.

Table 1: Products and biotransformation reactions involving amine functional groups

Structure	Reaction	A *	B*	C*		Structure	Reaction	A *	B*	C*
	Pheniramine (PHE)						Pargyline (PAR)			
	r Heilitailille (FTL)					N⊕ N⊕	N-oxidation	Cp	76.6 (100)	rf
N N O O	N-oxidation	Cp	6.0 (73)	r		H N	N-demethylation	C ²	4.4 (66)	rf
			(. 0)				α-C-oxidation to amide or N- demethylation followed by N-	Ps	6.8 (96)	rf
N H	N-demethylation	Cp	8.4 (73)	r		r 1	formylation N-demethylation followed by either		` '	
						(H	N-hydroxylation, reaction at the terminal alkyne, or α-C-hydroxylation to hemiaminal	P ³	0.5 (85)	rf
	α-C-oxidation to formamide or N- demethylation followed by N- formylation	Ps	0.6 (73)	r		CI	Chlorocyclizine (CLC)		
N-	Venlafaxine (VEN)					CI				
⊕0,N—			F 4		I		N-demethylation	Cp	0.2	r
OH OH	N-oxidation	Cp	5.1 (64)	r		N NH			(97)	•
HN-OH	N-demethylation	Cp	3.0 (64)	r		0N-N-N-N-N-N-N-N-N-N-N-N-N-N-N-N-N-N-	Pyrilamine (PYR)			
\rightarrow \text{N} \rightar	Lidocaine (LID)					0-(N-)N-)N-)N-)N-)N-)N-)N-)N-)N-)N-)N-)N-)N	N-oxidation	L ³	16.2 (96)	rs
⊖ _O H H	N-oxidation	Cp	19.1 (47)	r		—\v\⊕ 0Θ			(00)	
	N-deethylation	Ср	12.3	r) — NH	N-demethylation	C ²	1.8 (96)	r
H	N-dealkylation (deamination)		(77)	<u>'</u>		HN—N=	N-debenzylation	C ²	0.5 (96)	rs
но	followed by reduction of the aldehyde product to an alcohol	L ³	0.3 (75)	rs			N. J. H. Leffer (Leaves and Allen		(30)	
XX	Spiroxamine (SPI)					N-W-	N-dealkylation (deamination) followed by oxidation of aldehyde product to a carboxylic acid	L ³	0.6 (96)	rs
X J H	N-deethylation	C ²	0.3 (91)	rf		HÓ Na				
X NH	N-depropylation	C ²	0.3 (91)	rf	$\ $		α-C-oxidation to amide or N- demethylation followed by N- formylation	P ^s	0.3 (93)	rf
	Deprenyl (DEP)					CI——N	N,N-dimethyl-p-chloroanilin	e (D	CA)	
__\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\\	N-oxidation	Cp	69.4 (99)	rf		[c	N-oxidation, α-C-hydroxylation to hemiaminal, or hydroxylation	P ³	1.9 (98)	f
+0	α-C-hydroxylation to a hemiaminal or reaction at the terminal alkyne	P ³	0.5 (98)	rf		CI——NH	N-demethylation	Cp	20.9 (86)	f
NH_	N-demethylation	Cp	5.6 (77)	rf		CI—N=0	α-C-oxidation to formamide or N- demethylation followed by N- formylation	P ²	11.8 (98)	rf
—————————————————————————————————————	N-depropargylation	Cp	13.7 (100)	r			1[(4-chlorophenyl)phenylmethyl]pip	erazir	ne
	α-C-oxidation to amide or N- demethylation followed by N- formylation	P ³	0.6 (90)	rf		N NH	(CPP)			
	N-depropargylation followed by N- formylation	L ³	0.1 (100)	rs		CI	N application	L ³	0.9	
	N-demethylation followed by N-acetylation	L ³	0.2 (99)	rf			N-acetylation	L	(99)	rs
* A: certainty of assigned	d reaction as certain (C), likely (L), ar	d n	oooible	. / [2/ T I	Supercarinta indicata etr		^p an	_1 S	_

^{*}A: certainty of assigned reaction as certain (C), likely (L), and possible (P). Superscripts indicate structural confidence levels, whereby pand represent TP structures with Level 1 that were confirmed by purchased or synthesized reference standards, respectively; B: maximal relative amount in % and, in brackets, degree of of biotransformation of the parent compound at the time point where the TP is maximal, in %; and C: characterization of time series pattern as rising (r), rising and steady (rs), and rising and falling (rf)

Table 1: continued

Structure	Reaction	A *	В*	C*	* Structure	Reaction	A *	В*	C*
CINO	N-succinylation	L ³	10.3 (97)	rs	O N OH	N-succinylation	L ³	9.2 (22)	rf
NH NH	Ortho-chorophenylpiperazine	e (O	CP)		OH OH	N-fumarylation or N-succinylation followed by desaturation	P^3	8.7 (33)	r
NH +O	N-hydroxylation or α-C- hydroxylation to hemiaminal	P^3	0.5 (75)	rf		Siloned by decadaration		(00)	
NH OH	N-dealkylation of 3° and 2° amine, followed by oxidation of aldehyde product to carboxylic acid	L ³	1.1 (93)	rs	HN OH	N-demethylvenlafaxine (N	VE)		
NH -2H +0	either $\alpha\text{-C-oxidation}$ to lactam or N-hydroxylation followed by oxidation to nitrone	P ³	0.4 (95)	rf	HO-V	N-dealkylation (deamination) followed by oxidation of the aldehyde product to carboxylic acid	L ³	0.6 (32)	rs
NH TO	α-C-oxidation to lactam	L ³	0.5 (93)	rs		and dehydration to alkene either hydroxylation followed by oxidation to a carbonyl,			
N N	N-formylation	L ³	2.5 (75)	rf		hydroxylation in combination with desaturation, α-C-oxidation to amide, combination of N-	P^3	1.0 (32)	r
CI N-V	N-acetylation	L ³	5.5 (85)	rf	: [demethylation and N-formylation, or N-hydroxylation followed by oxidation to nitrone			
CI N-OH	N-malonylation	L ³	1.0 (95)	r	NH -2H	either α-C-oxidation to amide or N- hydroxylation and subsequent	P ³	0.3	
CI NO HO	N-succinylation	Cp	3.1 (91)	rf	ОН	reaction to nitrone or combination of N-demethylation and N-formylation	P	(32)	r
CI HO	N-fumarylation or N-succinylation followed by desaturation	P ³	2.4 (95)	r	N O O	combination of N-demethylation and N-acetylation and dehydration to alkene	L ³	0.5 (32)	r
	N-demethylpheniramine (N	NPE	i)		F ₃ C H,	Fluoxetine (FLU)			
	N-dealkylation (deamination)	L ³	0.5 (33)	rs	F ₃ C NC ₃ H ₆ O	N-acetylation or combination of N-demethylation and N-propionylation	P ³	0.1 (32)	r
ОН	N-dealkylation (deamination) followed by oxidation of aldehyde product to carboxylic acid	L ³	0.8 (14)	rf	HO N	N-succinylation	L³	0.2 (96)	rf
	α -C-oxidation	L ³	0.6 (39)	rs	F HN	Feniramine (FEN)			
O NH	α-C-oxidation to formamide or combination of N-formylation and N-demethylation or vice versa	P ³	1.8 (40)	r	F—F H ₂ N	N-deethylation	Cp	2.4 (84)	r
	N-formylation	Cs	3.3 (40)	rs	F HN OH	combination of N-deethylation and N-hydroxylation	L³	0.7 (84)	rs
	N-acetylation	L ³	1.3 (33)	r		either hydroxylation followed by oxidation to carbonyl, combination of hydroxylation and desaturation, or N-hydroxylation followed by oxidation to nitrone	P ³	2.2 (84)	r
H, H	N-acetylation followed by N- demethylation or vice versa	P ³	0.9 (40)	r	F—F HN	α-C-oxidation to acetamide or combination of N-deethylation and N-acetylation	P ³	0.3 (84)	rs

^{*}A: certainty of assigned reaction as *certain* (C), *likely* (L), and *possible* (P). Superscripts indicate structural confidence levels, whereby pands represent TP structures with Level 1 that were confirmed by purchased or synthesized reference standards, respectively; B: maximal relative amount in % and, in brackets, degree of of biotransformation of the parent compound at the time point where the TP is maximal, in %; and C: characterization of time series pattern as rising (r), rising and steady (rs), and rising and falling (rf)

Table 1: continued

Structure	Reaction	A* B* C*		Structure	Reaction		В*	C*	
H OH CI	Clenisopenterol (CLE)				HN-V	N-acetylation	L ³	54.8 (88)	rs
O_ HN NH ₂	Primaquine (PRI)				HN	N-propionylation	L ³	1.8 (85)	r
NH ₂	N-dealkylation of 2° amine	C ²	5.5 (100)	rf	но				
N OH	N-dealkylation (deamination) of 1° amine and succesive oxidation of aldehyde product to carboxylic acid	L ³	64.7 (99)	rf	HN	N-succinylation	L ³	0.3 (48)	rf
OH OH	N-formylation of 2° amine in addition to N-dealkylation (deamination) of 1° amine followed by oxidation to carboxylic acid,	L ³	0.2 (99)	rf	HN	N-fumarylation or N-succinylation followed by desaturation	P ^s	0.3 (85)	r
N N N N N N N N N N N N N N N N N N N	N-acetylation of 1° amine	L ³	4.0 (99)	rf	F ₂ C NH ₂	N-demethylfluoxetine (NI	FL)		
	desaturation or oxidation of 2° amine to imine in addition to N-dealkylation (deamination) of 1° amine followed by oxidation to carboxylic acid,	P ³	5.3 (100)	rf		N-acetylation	L ³	2.4	rf
NH ₂	Mexiletine (MEX)				F ₃ C	,		(100)	
	desaturation or oxidation of primary amine to imine	P ³	0.7 (85)	r	F ₃ C	N-propionylation	P ³	2.2 (100)	rf
HN	N-formylation	L ³	0.5 (81)	rf	F ₃ C HOOO	N-succinylation	L ³	(96)	rf

^{*} A: certainty of assigned reaction as *certain* (C), *likely* (L), and *possible* (P). Superscripts indicate structural confidence levels, whereby pand represent TP structures with Level 1 that were confirmed by purchased or synthesized reference standards, respectively; **B**: maximal relative amount in % and, in brackets, degree of of biotransformation of the parent compound at the time point where the TP is maximal, in %; and **C**: characterization of time series pattern as rising (r), rising and steady (rs), and rising and falling (rf)

Biotransformation Reactions

N-oxidation

N-oxidation reactions involve the direct introduction of an oxygen atom on the nitrogen of the
amine functional group, resulting in N-oxide products for tertiary amines and hydroxylamine products for secondary and primary amines. This reaction is known from microbial ³⁴ and mammalian
metabolism of amines. For the mammalian metabolism of xenobiotics the following findings have
been reported: ^{29,31,41} N-oxidation reactions occur for a broad range of amines, such as aliphatic
amines, azaheterocycles, pyridine derivatives, and aniline derivatives. Whereas tertiary amines
are preferably oxidized by flavin-containing monooxygenases (FMOs), primary amines tend to be

transformed by cytochrome P450 (CYPs) enzymes. For secondary amines, both enzyme systems
are reported to be active. Furthermore, for compounds with two tertiary amine moieties, the
N-oxidation reaction is reported to occur preferably at the more basic amine moiety. In addition,
N-hydroxylated products are known to be further transformed to products such as nitrones (for
secondary amines) and nitroso, nitro, and oxime products (for primary amines). Both N-oxides
and hydroxylamines can be reduced and backtransformed to the original parent amine.

As can be seen in Table 1, six (PHE, VEN, LID, DEP, PAR, PYR) out of 13 tertiary amines were *likely* or *certainly* transformed to N-oxides, and one more amine (DCA) was *possibly* transformed by that same reaction. We thus observed the N-oxidation reaction to occur for a broad range of aliphatic tertiary amines with variable N-substituents and possibly also for aromatic tertiary amines. We additionally found evidence of regioselectivity towards the more basic amine moiety (PHE, PYR). Noteworthy, the *maximal relative amounts* of four N-oxide TPs (LID, DEP, PAR, PYR) were rather high at over 15%, especially for the two methylpropargyl amines DEP and PAR with values of 69.4% and 76.6%, respectively. This indicates that amines substituted with a propargyl moiety might be particularly amenable to the N-oxidation reaction. Overall, N-oxidation reactions were abundant for tertiary amines and thus constitute a major microbial biotransformation pathway for these compounds.

To test if N-oxidized products can be reduced back to tertiary amines, we spiked an additional biotransformation reactor with a mixture of five N-oxide standards (PENO of PHE, VENO of VEN, LINO of LID, DENO of DEP, PANO of PAR). Indeed, we observed the formation of the respective amines for PENO, VENO, and LINO, which clearly indicates that the backtransformation of N-oxide TPs to tertiary amines is possible. The corresponding time series are shown in Chapter S7 in the SI. Interestingly, no clear formation of the parent amines was observed for the N-oxides of DEP and PAR. This suggests that the lack of backtransformation of the methylpropargyl N-oxides could explain their high occurrence in the parent amine reactors. We additionally observed the formation of the N-dealkylated products for all N-oxides, which were formed in a similar or even higher amount (NPE, NVE) than when formed from the respective tertiary amines. It remains

unclear whether the formation of the N-dealkylated product is caused by a direct reaction from the
N-oxide to the N-dealkylated product, or whether it is formed in two steps, namely the reduction
of the N-oxide to the amine followed by the N-dealkylation of the amine. Similar processes
were observed for the N-oxide of N,N-dimethylaniline in mammalian systems. ⁴¹ A hydroxylamine
product was *likely* observed for FEN and *possibly* for PAR and OCP. Additionally, four nitrone
products (OCP, 2xNVE, FEN) were identified as *possible*. The confidence of the corresponding
structural assignment is low due to a number of other possible structures that would also match
the available structural evidence.

α-C-hydroxylation and subsequent N-dealkylation, oxidation to amides, or dehydration to iminium species

From both mammalian and microbial metabolism of xenobiotics, the hydroxylation of C-atoms 287 in α -position to N-atoms of amine functional groups is known to be catalyzed by cytochrome 288 P450 (CYPs) enzymes. 29,31,34 This reaction, which requires at least one hydrogen atom at the 289 α-carbon, results in hemiaminal intermediates, which are usually unstable. For most compounds, 290 these hemiaminals are spontaneously cleaved yielding amine and carbonyl products. This reaction 291 is called N-dealkylation. However, two less common alternative reactions of hemiaminal interme-292 diates have also been reported. ^{29,31,41} On the one hand, oxidation of the hemiaminal intermediates 293 to amides, which requires two hydrogen atoms at the α -carbon, is possible, especially for cyclic 294 amines. On the other hand, dehydration of the hemiaminal intermediate may lead to the forma-295 tion of an iminium species. The N-dealkylation reaction cleaves the substrate into two products 296 and since it is more likely to detect the product that bears the major structural unit, the reaction 297 is sometimes also named deamination if the carbonyl product bears the major structural unit. 298 In mammalian systems, N-dealkylation reactions occur for a variety of amines such as aliphatic, 299 aromatic, and cyclic amines. The readiness of the reaction decreases from tertiary over secondary to primary amines. Sterically less hindered N-substituents (methyl and ethyl) are more readily 301 cleaved than longer and α -branched N-substituents. For the carbonyl product it is known that further biotransformation reactions are likely to occur, such as the oxidation to a carboxylic acid moiety in case of aldehydes or the reduction to an alcohol moiety. ^{29,31,41}

Four hemiaminals were *possibly* observed (DEP, PAR, DCA, OCP). However, also other structures such as N-oxides, hydroxylamines, or hydroxylated TPs match the structural evidence. Since hemiaminals are usually unstable and a certain stability is only known for less basic amines, it is possible that only some of the less basic aliphatic amines DEP and PAR and the aniline-derivative DCA actually formed hemiaminals.

For 11 of the 13 tertiary amine moieties (the exceptions being the inner amines of the 310 piperazine moieties of CLC, CPP), one or two products of N-dealkylation were identified. N-311 demethylation occurred for all seven tertiary methylamines (PHE, VEN, DEP, PAR, CLC, PYR, 312 DCA). Other N-substituents for which dealkylation was observed were ethyl (LID, SPI), propyl 313 (SPI), propargyl (DEP), and p-methoxybenzyl (PYR). We additionally found oxidative ring open-314 ing (OCP) and cleavage of diethylamine and demethylamine in the case of LID and PYR, respec-315 tively, to yield an alcohol (LID) or carboxylic acid product (PYR) as the major structural unit. 316 The maximal relative amounts for four of these TPs were higher than 5%, with a distinctly high 317 value for the N-demethylated product of DCA of 20.9%. For the eight secondary amines, certain 318 or likely N-dealkylation was observed five times with cleavage of the following N-substituents: 319 piperazine ring (OCP), major structural unit (NPE, NVE), ethyl (FEN), and an α -branched ω amino alkyl chain (PRI). Additionally, possible demethylation was observed for three compounds 321 (NPE, NVE, FLU). For the three primary amines, only one deamination product was observed 322 (PRI). However, this TP was formed with an outstandingly high maximal relative amount of 323 64.7%. 324

Overall, our observations regarding the N-dealkylation reactions concurred with findings from mammalian systems with respect to the frequent occurrence of N-dealkylation for a range of tertiary, secondary, and primary alkyl, alkyl-aryl, and cyclic amines. Again, in agreement with mammalian systems, we also observed the sterically less hindered substituents to be more readily cleaved than more sterically hindered moieties. ^{29,31,41}

325

326

327

328

The further oxidation of hemiaminal intermediates to amide products was likely observed for 330 the two aliphatic amines OCP and NPE, where the α -carbon of the N-substituent bearing the 331 major structural unit was oxidized. Besides that, eight formamide or acetamide products were 332 observed (PHE, DEP, PAR, PYR, DCA, NPE, NVE, FEN) that could possibly be formed through 333 α-C-oxidation. However, since these products have been observed for methyl- and ethylamines, 334 they could also have been formed through N-demethylation and subsequent N-formylation, or N-335 deethylation followed by N-acetylation, respectively, as described in the next section. The fact that 336 only two amide TPs were observed where the α -C-oxidation occurred at the longer N-substituents 337 suggests that either α -C-oxidation occurs more readily for methyl and ethyl substituents or their 338 formation is mediated via the two-step reaction path. Findings from mammalian systems indicate 339 that for α -C-oxidation reactions to compete with spontaneous cleavage of the hemiaminal a 340 certain stability of the hemiaminal intermediates is required. This is given for less basic amines 341 such as aniline-derivatives, or for cyclic amines exhibiting a ring-chain tautomerism, but less likely 342 for the methyl- and ethylamines for which we observed amide formation. Therefore, and due to 343 several arguments discussed in the following section, it is likely that the observed formamides 344 and acetamides are mostly formed via sequential N-dealkylation and N-acylation. Overall, the 345 formation of amides through α -C-oxidation reactions seems to be a minor pathway compared to N-dealkylation. The third alternative reaction of hemiaminals, i.e., the formation of iminium species, was possibly observed twice (PRI, MEX). However, other structures would also match the available structural evidence.

N-acylation of secondary and primary amines

For the mammalian metabolism of xenobiotics, N-acetylation reactions are mainly known for primary aryl amines and are catalyzed by N-acetyltransferase enzymes that use acetyl-coenzyme
A as an acetyl donor. This reaction was also observed to occur for an even broader substrate spectrum, including aliphatic amines, in microbial systems. 42–44 Other N-acylation reactions such as N-formylation, N-propionylation, N-malonylation, and N-succinylation were not yet known for

the metabolism of xenobiotics in mammalian systems but were observed in other contexts such as the modification of lysine residues of proteins. 45 Similarly, N-acetylation and N-succinvlation 357 of lysine residues were also reported for bacteria. 46 358

For 9 of the 12 primary and secondary amines (the exceptions being both amine moieties of 359 CLE and the secondary amine of PRI), an N-acetylation reaction was likely (CPP, OCP, NPE, 360 NVE, PRI, MEX, NFL) or possibly (FLU, FEN) observed. Additionally, the N-demethylated 361 product of the tertiary amine of DEP underwent N-acetylation. The N-acetylated product of 362 MEX was formed in a remarkably high maximal relative amount of 54.8%. This, in addition to 363 the frequent occurrence of the N-acetylation reaction for primary and secondary amines, indicates 364 the importance of this reaction. 365

367

371

Formamide products, indicating the occurrence of an N-formylation reaction, were observed 366 12 times. Five of these TPs (DEP, OCP, NPE, PRI, MEX) were formed certainly or likely via N-formylation. For the remaining seven TPs (PHE, DEP, PAR, PYR, DCA, NPE, NVE), two 368 reaction paths are possible, as mentioned in the previous section, namely N-demethylation followed 369 by N-formylation or an α -C-oxidation of a methylamine to a formamide product. Beyond the 370 arguments supporting the two-step reaction pathway already mentioned in the previous section, we further observed the following: for five of the seven TPs in question, the intermediate Ndemethylation product was formed; the other N-formylated TPs demonstrate that N-formylation of secondary and primary amines is a common reaction; and two TPs of DEP are likely formed by a combination of N-dealkylation and N-acylation. We further inspected the time series pattern 375 of the N-dealkylated and formamide products for evidence of a sequential reaction (i.e., transient 376 patterns for the dealkylated products and an initial slow increase of the N-acylation products). 377 No consistent pattern that pointed towards one of the two reaction paths was found. Overall, we 378 nevertheless conclude that formamide products of methylamine parents were more likely formed 379 via combination of N-demethylation and N-formylation. 380

Further, we identified TPs that pointed towards possible (FLU, NFL) or likely (MEX) N-381 propionylation reactions, one likely N-malonylation reaction (OCP), and certain or likely N-382

succinylation reactions for four out of eight secondary amines (CPP, OCP, NPE, FLU) and two
out of four primary amines (MEX, NFL). The N-succinylated TP of OCP was confirmed by a
reference standard, which considerably increased our confidence in the structural assignments
of all N-succinylated products. The two products of CPP and NPE were formed in *maximal*relative amounts of 10.3% and 9.2%, respectively. These amounts and the occurrence of the
N-succinylation reaction for half of the secondary and primary amines indicates the importance
of this reaction, which, to the best of our knowledge, has only been observed as microbial
biotransformation reaction of a xenobiotic once before. 47

We also observed three N-fumarylated products (OCP, NPE, MEX), with the TP of MEX confirmed by a reference standard. These products can be formed either through direct N-fumarylation or by desaturation of the N-succinylated product. The latter reaction path is supported by the following: for all N-fumarylated TPs also an N-succinylation TP was observed; the time series patterns of the N-fumarylated product in comparison to those of the N-succinylated products point towards a sequential reaction path (see Chapter S6 in the SI); and the desaturation of succinate to fumarate is a common process in the metabolism of cells. Therefore, it is more likely that the N-fumarylated products were formed through a desaturation of the N-succinylated amines.

It should also be noted that the N-acylation products with short chains could be formed from an N-acylation reaction with a longer substituent that is subsequently cleaved, rather than by direct N-acylation. However, the frequency of detection, the time series patterns, and the similarity to reactions observed in mammalian systems suggests that especially the N-acetylation, N-formylation, and N-succinylation TPs were formed directly.

Overall, N-acetylation and N-succinylation and to a lesser extent N-formylation reactions were highly relevant for secondary and primary amines. N-propionylated, N-malonylated and N-fumarylated products were also observed. Except for N-acetylation, these N-acylation reactions represent the greatest difference between what we observed for microbial biotransformation relative to what is known from mammalian systems, in which the latter type of transformation

has not been reported for xenobiotics so far. However, the biochemical reactions themselves are known from other contexts for both microbial as well as mammalian systems. The respective N-acyltransferase enzymes are likely to transform the amines co-metabolically. 412

It should further be noted that the products of N-acylation reactions of primary and secondary 413 amines are secondary and tertiary amides. It is known from both mammalian 32 and microbial 414 metabolism²⁶ that amides can be hydrolysed to form a carboxylic acid and an amine product. 415 In a previous study that explored biotransformation pathways of amide-containing compounds, it 416 was observed that amide hydrolysis mostly occurred for primary and secondary amides while for 417 tertiary amides other reactions such as N-dealkylation were observed. ²⁶ In our context this means 418 that especially primary amines that undergo N-acylation reactions to form secondary amides are 419 likely to be back-transformed through hydrolysis.

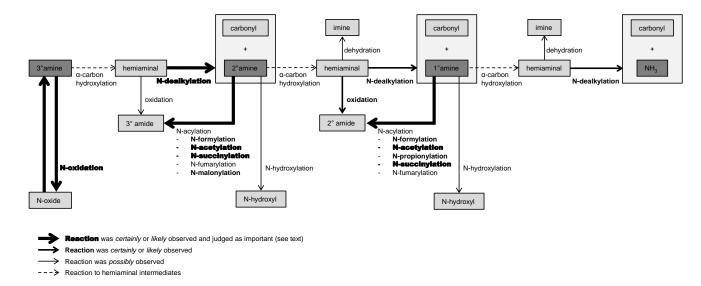


Figure 1: Overview of all observed reactions involving tertiary, secondary, and primary amine functional groups.

Further biotransformation reactions

411

420

421

Additionally to reactions at the amine functional group, we found the following transformation 422 reactions (Table 1 or Chapter S8 in the SI): O-demethylation (VEN, PYR, NVE), which was 423 observed for three of four methylarylamines; desaturation (LID, PRI, MEX); dehydration of an

Table 2: Relative importance of reactions involving amine functional groups

	ter	tiary a	mines	secondary amines			pri	mines	
	frequ	ency	mean maximal	frequ	ency	mean maximal	frequ	ency	mean maximal
	abs.	rel. [%]	relative amount [%]	abs.	rel. [%]	relative amount [%]	abs.	rel. [%]	relative amount [%]
N-oxidation	6 /13	46.2	14.8	0/8	0.0	-	0/4	0.0	-
N-dealkylation°	12 / 27	44.4	2.6	3 / 13	23.1	0.6	0/3	0.0	-
- N-demethylation	7/7	100.0	6.3	0/3	0.0	-	0/3	0.0	-
- N-deethylation	2/2	100.0	6.3	1/1	100.0	2.4	0/3	0.0	-
- N-dealkylation of longer substituents #	3 / 18	16.7	0.8	2/9	22.2	0.7	0/3	0.0	-
α-C-oxidation to amide *	0/16	0.0	-	2/7	28.6	0.16	0/1	0.0	-
N-formylation	0/0	-	_	2/8	25.0	0.7	1/4	25.0	0.1
N-acetylation	0/0	-	<u>-</u>	3/8	37.5	1	3/4	75.0	15.3
N-propionylation	0/0	-	-	0/8	0.0	-	1/4	25.0	0.5
N-malonylation	0/0	-	-	1/8	12.5	0.1	0/4	0.0	-
N-succinylation	0/0	-	-	4/8	50.0	2.9	2/4	50.0	0.6

[°] the following three specific N-dealkylation reactions are a subset of this general one. # including direct deamination products, i.e., carbonyl products, but not further transformed products such as carboxylic acids or alcohols. * not including the formation of formamides / acetamides from methylamines / ethylamines. Frequency is given as absolute (abs.) or relative (rel.) value. Reactions that were judged important (see text) are highlighted in bold.

aliphatic alcohol to an alkene (2x NVE); and hydroxylation (2x VEN, LID, 2x SPI, DEP, PAR, CLC, 425 DCA, NVE). Additionally, the following reactions likely occurred after hydroxylation: TPs with 426 an atomic modification of -2H+O that could have been formed through a combination of either 427 hydroxylation and subsequent oxidation to a carbonyl product or hydroxylation and independent 428 desaturation (2x VEN, SPI, 2x NVE, FEN); hydroxylation followed by oxidation to a carboxylic 429 acid product (LID); and TPs for which a combination of several of the discussed reactions as 430 well as decarboxylation reactions or β -oxidations are possible (VEN, 2x SPI, PRI). With this, we 431 possibly found for 12 of 19 test compounds additional biotransformation reactions in parallel to 432 transformations involving the amine functional groups. Furthermore, although seven TPs were 433 successfully assigned for SPI and PRI, for six TPs of the two parent amines it was not possible 434 to assign any reaction since the structural evidence was insufficient to propose a meaningful 435 structure. 436

137 Metabolic logic

Overall, we found that biotransformation reactions of amine-containing xenobiotics were indeed fairly similar in mammalian and microbial systems. This was certainly true for all transformations that were initiated by oxidative reactions. Additionally, these reactions also showed dependencies, where observable, on electronic and steric features that were consistent with those known from mammalian systems. However, we also observed a number of N-acylation reactions that have not been previously described for the metabolism of xenobiotics in mammalian systems, and which are most likely co-metabolically catalyzed by N-acyltransferases expressed for the regulation of proteins.

Figure 1 illustrates all observed reactions for the tertiary, secondary, and primary amine moieties. Observed reactions that did not concern the amine functional group are not displayed. To assess and compare the importance of the different reactions in terms of how much of the amines were biotransformed through each type of reaction, the following values were calculated based on all first-generation products that were formed *certainly* or *likely* through each type of reaction: the *relative frequency*, *i.e.*, the number of observed products relative to the number of theoretically possible products, and the *mean maximal relative amounts* of all TPs formed, taking into account again the number of theoretically possible products. The values are given in Table 2. Reactions with either a *relative frequency* >30% or a *mean maximal relative amount* >1% were judged as important compared to the spectrum of possible reactions and are illustrated in bold in Table 2 and with a thick arrow in Figure 1.

As can be seen in Figure 1 and Table 2, tertiary amines were predominantly biotransformed through N-oxidation and N-dealkylation reactions. N-acetylation and N-succinylation reactions were most important for secondary and primary amines. A proper evaluation of N-oxidation of primary and secondary amines to form hydroxylamines and subsequent nitrone products was impeded by the fact that structural assignment remained ambiguous. For all types of amines studied, diverse other, mainly oxidative reactions that did not involve the amine functional group were observed to proceed in parallel.

164 Implications for pathway prediction

With respect to improving tools for pathway prediction, particularly the Eawag-PPS system, this 465 study results in a number of suggestions concerning the rules for biotransformation of amine 466 functional groups. While the N-dealkylation reaction is already implemented in the Eawag-PPS 467 system and is triggered without any further specification for all types of amines, the other amine 468 reactions that we observed have so far not been implemented. While some of these reactions 469 are specific for certain types of amines, i.e., tertiary, secondary, or primary amines, the different 470 biotransformation reactions at the amine functional group were generally observed to occur in 471 parallel. No structural features could be identified that would promote certain reactions to 472 occur more readily than others. Therefore, rather than reducing combinatorial explosion, the 473 implementation of new rules will lead to an increase of predicted products for amine-containing 474 compounds. 475

One rather crude, but, in the light of still limited data, tangible way to guide interpretation of 476 the outcomes of pathway prediction towards the more important products is to assign likelihoods 477 to reactions. In the Eawag-PPS, one of five levels of aerobic likelihood (very likely, likely, neutral, 478 unlikely, and very unlikely) is assigned to each biotransformation rule to express expert knowledge 479 about the likelihood of the reaction occurring under aerobic conditions. ^{22,48} These likelihood levels could be used to express the differences in metabolic relevance of the observed amine biotransformation reactions. When doing so, the likelihoods for reactions at the amine functional 482 group should be consistent with the likelihoods of reaction occurring at other functional groups. 483 The fact that reactions involving the amine functional group were observed to occur in parallel 484 with O-demethylation reactions of aromatic ethers and miscellaneous C-hydroxylation reactions, 485 which are both implemented as neutral in the Eawag-PPS, suggests that they should also be 486 implemented as neutral. This is further supported by the hydrolysis reaction for primary and 487 secondary amides being implemented as likely, which is known to be preferred over (i) reactions 488 at the amine functional group for atenolol⁴⁹, and (ii) oxidative N-dealkylation reactions from 489 data collected for amides. 26 Therefore, we suggest the implementation of reactions involving the

amine functional group with a likelihood of neutral. In order to prevent overrepresentation of less likely reactions, we suggest the implementation of only those reactions that were judged to 492 be important. For the Eawag-PPS this would result in the implementation of the following new 493 rules: reversible N-oxidation for tertiary amines, and N-acylation reactions such as N-acetylation 494 and N-succinylation for secondary and primary amines. Further, the already implemented N-495 dealkylation rule (bt0063), which is triggered for all N-substituents of an amine moiety and 496 predicts the corresponding N-dealkylated and deaminated products, should be reassigned from 497 likely to neutral. The latter rule could be further refined based on the observation that N-498 dealkylation was found to be less prevalent for secondary and primary amines and for the cleavage 499 of sterically hindered substituents of tertiary amines. This could be represented by implementing 500 the N-dealkylation rule with a variable aerobic likelihood, ^{21,22} so that N-dealkylation reactions of 501 tertiary amines are assigned as neutral for methyl and ethyl, whereas N-dealkylation of secondary 502 and primary amines and tertiary amines that involve bond cleavage of larger substituents are 503 assigned as unlikely. Updating of rules in Eawag-PPS as suggested, would not only increase 504 the sensitivity of the prediction for amine-containing compounds, but would also gear predictions 505 towards more specifically predicting microbial biotransformation under environmentally relevant 506 conditions.

While implementation of fixed likelihood levels is the currently most tangible way of guiding interpretation of predicted pathways, more refined ways of learning from our data are currently being developed. Recently, Eawag-PPS has been redesigned and reimplemented into a new system called enviPath. The new system was specifically designed to facilitate entry of micropollutant biotransformation data, *i.e.*, biotransformation reactions and half-lives, and storage of meta-data, e.g., on environmental conditions. It also supports learning more sophisticated relative reasoning models on the data. We plan to add the data reported in this study into the enviPath system and also would like to encourage others to enter own data on biotransformation of micropollutants in activated sludge communities. Ultimately, the collection of these data should enable training pathway prediction engines that are more specific to activated sludge systems.

508

509

510

511

512

513

514

515

516

517

Environmental relevance

521

525

526

527

528

529

530

531

532

533

535

536

537

538

539

540

541

542

The finding that microbial and mammalian transformation reactions were in parts fairly simi-519 lar suggests that predicted as well as experimental transformation data from the mammalian 520 metabolism of pharmaceuticals can be used for suspect screening approaches targeted at TPs in the environment. Additionally, the well explored reactions of the mammalian system can support 522 the interpretation of biotransformation reactions in less explored microbial systems as it was done 523 in the present study. 524

Additionally, if we assume that our findings in activated sludge-seeded bioreactors can be extrapolated to full-scale WWTPs, we can expect that amine-containing compounds, which do not contain any other functional groups that are biotransformed more readily, undergo several reactions in parallel. Other studies exploring more structurally complex amine-containing MPs than we did also reported the products of transformation at the amine functional group to be found alongside products formed through reactions at other functional groups. 51-53 Since we found that very few of the resulting TPs were formed in major amounts, it is reasonable to assume that a complex mixture of the parent MPs and a number of low-level TPs will be present after activated sludge treatment unless the compound contains some other readily transformed functional group. Hence, the potential resulting (eco-)toxicological effects depend on the relative toxicity of the remaining parent compounds and the individual TPs. In cases where the amine moiety is involved in inducing a toxic effect of the parent compound, biotransformation at this site might result in less toxic TPs. This was shown for the N-acetylated TP of the herbicide glyphosate 42 as well as for the N-oxidized TPs of the antibiotics clarithromycin⁵⁴ and danofloxacin⁵⁵. However, since we know that N-oxide TPs as well as N-acylated TPs can be back-transformed to the parent amines - also reported for MPs holding other functional groups 56,57 - their toxic effect could be regained in the environment. Therefore, it is important to also consider the individual TPs and their potential reactions when assessing the risk of MPs.

Acknowledgements

We thank Jennifer Schollee for fruitful discussions. Additionally, we thank the operators and staff of the WWTP ARA Neugut for providing activated sludge samples. Funding for this project was provided by the Swiss National Science Foundation (project 200021_134677) and the NMR hardware was partially granted by the Swiss National Science Foundation (SNFS, grant No. 150638).

Supporting Information

Details on materials and methods, identified but not elucidated TPs, schematic illustration of amine reactions, structure elucidation evidence for all TPs, protocol of synthesis and NMR data of synthesized reference TPs, concentration time series of the parent compounds, pathways maps, time series patterns of TPs, time series pattern of N-oxide reactor, summary of TPs that were formed by reactions not involving the amine functional group, comparison of EAWAG-PPS predictions with experimental results, and comparison of ionization efficiency of TP and parent. This material is available free of charge via the Internet at http://pubs.acs.org.

557 References

- (1) Joss, A.; Keller, E.; Alder, A. C.; Göbel, A.; McArdell, C. S.; Ternes, T. A.; Siegrist, H. Removal of pharmaceuticals and fragrances in biological wastewater treatment. *Water Res.* **2005**, *39*, 3139–3152.
- oulton, R. L.; Kohn, T.; Cwiertny, D. M. Pharmaceuticals and personal care products in effluent matrices: A survey of transformation and removal during wastewater treatment and implications for wastewater management. *J. Environ. Monit.* **2010**, *12*, 1956–1978.
- 564 (3) Evgenidou, E. N.; Konstantinou, I. K.; Lambropoulou, D. A. Occurrence and removal of

- transformation products of PPCPs and illicit drugs in wastewaters: A review. *Sci. Total Environ.* **2015**, *505*, 905 926.
- the Biological Wastewater Treatment. *Environ. Sci. Technol.* **2014**, *48*, 244–254.
- (5) Rubirola, A.; Llorca, M.; Rodriguez-Mozaz, S.; Casas, N.; Rodriguez-Roda, I.; Barcelo, D.;

 Buttiglieri, G. Characterization of metoprolol biodegradation and its transformation products

 generated in activated sludge batch experiments and in full scale WWTPs. *Water Res.* **2014**,

 63, 21 32.
- 573 (6) Berkner, S.; Claudia, T. Biodegradability and transformation of human pharmaceutical ac-574 tive ingredients in environmentally relevant test systems. *Environmental Science and Pollu-*575 *tion Research* **2014**, *21*, 9461–9467.
- (7) Haddad, T.; Baginska, E.; Kuemmerer, K. Transformation products of antibiotic and cytostatic drugs in the aquatic cycle that result from effluent treatment and abiotic/biotic
 reactions in the environment: An increasing challenge calling for higher emphasis on measures at the beginning of the pipe. *Water Res.* **2015**, *72*, 75 126.
- (8) Toolaram, A. P.; Kuemmerer, K.; Schneider, M. Environmental risk assessment of anti-cancer drugs and their transformation products: A focus on their genotoxicity characterization-state of knowledge and short comings. *Mutat. Res.-Rev. Mutat.* **2014**, 760, 18 35.
- (9) Celiz, M. D.; Tso, J.; Aga, D. S. Pharmaceutical metabolites in the environment: Analytical challenges and ecological risks. *Environ. Toxicol. Chem.* **2009**, *28*, 2473–2484.
- 586 (10) Boxall, A. B. A.; Sinclair, C. J.; Fenner, K.; Kolpin, D.; Maund, S. J. When synthetic chemicals degrade in the environment. *Environ. Sci. Technol.* **2004**, *38*, 368A–375A.

- ⁵⁸⁸ (11) Pico, Y.; Barcelo, D. Transformation products of emerging contaminants in the environment and high-resolution mass spectrometry: a new horizon. *Anal. Bioanal. Chem.* **2015**, *407*, 6257–6273.
- 591 (12) Schymanski, E. L. et al. Non-target screening with high-resolution mass spectrometry: crit-592 ical review using a collaborative trial on water analysis. *Anal. Bioanal. Chem.* **2015**, *407*, 593 6237–6255.
- ⁵⁹⁴ (13) Moschet, C.; Piazzoli, A.; Singer, H.; Hollender, J. Alleviating the Reference Standard Dilemma Using a Systematic Exact Mass Suspect Screening Approach with Liquid Chromatography-High Resolution Mass Spectrometry. *Anal. Chem.* **2013**, *85*, 10312–10320.
- ⁵⁹⁷ (14) Chiaia-Hernandez, A. C.; Schymanski, E.; Praveen, K.; Singer, H.; Hollender, J. Suspect and nontarget screening approaches to identify organic contaminant records in lake sediments.

 ⁵⁹⁸ Anal. Bioanal. Chem. **2014**, 406, 7323–7335.
- (15) Hug, C.; Ulrich, N.; Schulze, T.; Brack, W.; Krauss, M. Identification of novel micropollutants in wastewater by a combination of suspect and nontarget screening. *Environ. Pollut.*2014, 184, 25 32.
- (16) Kern, S.; Fenner, K.; Singer, H. P.; Schwarzenbach, R. P.; Hollender, J. Identification of
 Transformation Products of Organic Contaminants in Natural Waters by Computer-Aided
 Prediction and High-Resolution Mass Spectrometry. *Environ. Sci. Technol.* 2009, *43*, 7039–7046.
- 607 (17) Bletsou, A. A.; Jeon, J.; Hollender, J.; Archontaki, E.; Thomaidis, N. S. Targeted and non-608 targeted liquid chromatography-mass spectrometric workflows for identification of trans-609 formation products of emerging pollutants in the aquatic environment. *TrAC-Trend. Anal.* 610 *Chem.* **2015**, *66*, 32 – 44.
- 611 (18) Moriya, Y.; Shigemizu, D.; Hattori, M.; Tokimatsu, T.; Kotera, M.; Goto, S.; Kanehisa, M.

- PathPred: an enzyme-catalyzed metabolic pathway prediction server. *Nucleic Acids Res.* **2010**, *38*, W138–W143.
- 614 (19) The CRAFT website; http://www.molecular-networks.com/products/craft.
- 615 (20) The OECD QSAR Toolbox website; http://www.oecd.org/chemicalsafety/risk-616 assessment/theoecdqsartoolbox.htm.
- 617 (21) Gao, J.; Ellis, L. B. M.; Wackett, L. P. The University of Minnesota Pathway Prediction
 618 System: multi-level prediction and visualization. *Nucleic Acids Res.* **2011**, *39*, W406–W411.
- 619 (22) Ellis, L.; Wackett, L. Use of the University of Minnesota Biocatalysis/Biodegradation
 620 Database for study of microbial degradation. *Microb. Inform. Exp.* **2012**, *2*, 1.
- 621 (23) EAWAG-BBD Pathway Prediction System website; http://eawag-bbd.ethz.ch/predict/.
- fermatics **2008**, *24*, 2079–2085. Ellis, L.; Wackett, L. Data-driven extraction of relative reasoning rules to limit combinatorial explosion in biodegradation pathway prediction. *Bioinformatics* **2008**, *24*, 2079–2085.
- 625 (25) Fischer, K.; Majewsky, M. Cometabolic degradation of organic wastewater micropollutants 626 by activated sludge and sludge-inherent microorganisms. *Appl. Microbiol. Biotechnol.* **2014**, 627 98, 6583–6597.
- 628 (26) Helbling, D. E.; Hollender, J.; Kohler, H. P. E.; Fenner, K. Structure-Based Interpretation 629 of Biotransformation Pathways of Amide-Containing Compounds in Sludge-Seeded Biore-630 actors. *Environ. Sci. Technol.* **2010**, *44*, 6628–6635.
- (27) Helbling, D. E.; Hollender, J.; Kohler, H. P. E.; Singer, H.; Fenner, K. High-Throughput
 Identification of Microbial Transformation Products of Organic Micropollutants. *Environ.* Sci. Technol. 2010, 44, 6621–6627.
- 634 (28) Manallack, D. T. The acid-base profile of a contemporary set of drugs: implications for drug discovery. SAR QSAR Environ. Res. **2009**, 20, 611–655.

- (29) Testa, B. In *Biochemistry of Redox Reactions*; Testa, B., Caldwell, J., Eds.; Academic Press, 1995.
- (30) Testa, B.; Krämer, S. D. The Biochemistry of Drug Metabolism An Introduction: Part 1.

 Principles and Overview. *Chem. Biodiversity* **2006**, *3*, 1053–1101.
- 640 (31) Testa, B.; Krämer, S. D. The Biochemistry of Drug Metabolism An Introduction: Part 2.

 Redox Reactions and Their Enzymes. *Chem. Biodiversity* **2007**, *4*, 257–405.
- 642 (32) Testa, B.; Krämer, S. D. The Biochemistry of Drug Metabolism An Introduction: Part 3.

 Reactions of Hydrolysis and Their Enzymes. *Chem. Biodiversity* **2007**, *4*, 2031–2122.
- 644 (33) Testa, B.; Krämer, S. D. The Biochemistry of Drug Metabolism An Introduction: Part 4.

 Reactions of Conjugation and Their Enzymes. *Chem. Biodiversity* **2008**, *5*, 2171–2336.
- 646 (34) Azerad, R. Microbial models for drug metabolism. *Adv. Biochem. Eng. Biotechnol.* **1999**, 63, 169–218.
- (35) Smith, R. V.; Rosazza, J. P. Microbial models of mammalian metabolism. *J. Pharm. Sci.* **1975**, *64*, 1737–1759.
- 650 (36) Kern, S.; Baumgartner, R.; Helbling, D. E.; Hollender, J.; Singer, H.; Loos, M. J.;
 651 Schwarzenbach, R. P.; Fenner, K. A tiered procedure for assessing the formation of bio652 transformation products of pharmaceuticals and biocides during activated sludge treatment.
 653

 J. Environ. Monit. 2010, 12, 2100–2111.
- (37) Schymanski, E. L.; Jeon, J.; Gulde, R.; Fenner, K.; Ruff, M.; Singer, H. P.; Hollender, J.
 Identifying Small Molecules via High Resolution Mass Spectrometry: Communicating Confidence. *Environ. Sci. Technol.* 2014, 48, 2097–2098.
- (38) Gulde, R.; Helbling, D. E.; Scheidegger, A.; Fenner, K. pH-Dependent Biotransformation
 of Ionizable Organic Micropollutants in Activated Sludge. *Environ. Sci. Technol.* 2014, 48,
 13760–13768.

- 660 (39) MetaPrint2D website; http://www-metaprint2d.ch.cam.ac.uk/metaprint2d-react.
- 661 (40) The MassBank website; www.massbank.eu.
- 662 (41) Rose, J.; Castagnoli, N. I. The metabolism of tertiary amines. *Med. Res. Rev.* **1983**, *3*, 73–88.
- Glyphosate Tolerance Gene. Science 2004, 304, 1151–1154.
- (43) Clark, A. M.; Evans, S. L.; Hufford, C. D.; McChesney, J. D. Microbial N-acetylation of
 Primaquine by Two Streptomyces Species: Time Course Studies and HPLC Analyses. *J. Nat. Prod.* 1982, 45, 574–581.
- (44) Asha, S.; Vidyavathi, M. Cunninghamella A microbial model for drug metabolism studies
 A review. *Biotechnol. Adv.* 2009, 27, 16–29.
- 672 (45) Choudhary, C.; Weinert, B. T.; Nishida, Y.; Verdin, E.; Mann, M. The growing landscape 673 of lysine acetylation links metabolism and cell signalling. *Nat. Rev. Mol. Cell Biol.* **2014**, 674 *15*, 536–550.
- 675 (46) Kosono, S.; Tamura, M.; Suzuki, S.; Kawamura, Y.; Yoshida, A.; Nishiyama, M.;
 676 Yoshida, M. Changes in the acetylome and succinylome of Bacillus subtilis in response
 677 to carbon source. *PLoS ONE* **2015**, *10*.
- 678 (47) Terzic, S.; Senta, I.; Matosic, M.; Ahel, M. Identification of biotransformation prod-679 ucts of macrolide and fluoroquinolone antimicrobials in membrane bioreactor treatment by 680 ultrahigh-performance liquid chromatography/quadrupole time-of-flight mass spectrometry. 681 *Anal. Bioanal. Chem.* **2011**, *401*, 353–363.
- 682 (48) Ellis, L. B. M.; Gao, J.; Fenner, K.; Wackett, L. P. The University of Minnesota pathway 683 prediction system: predicting metabolic logic. *Nucleic Acids Res.* **2008**, *36*, W427–W432.

- (49) Helbling, D. E.; Johnson, D. R.; Honti, M.; Fenner, K. Micropollutant biotransformation
 kinetics associate with WWTP process parameters and microbial community characteristics.
 Environ. Sci. Technol. 2012, 46, 10579–10588.
- (50) Wicker, J.; Lorsbach, T.; Gütlein, M.; Schmid, E.; Latino, D.; Kramer, S.; Fenner, K. enviPath The environmental contaminant biotransformation pathway resource. *Nucleic Acids Res.* **2015**, gkv1229.
- (51) Beretsou, V. G.; Psoma, A. K.; Gago-Ferrero, P.; Aalizadeh, R.; Fenner, K.; Thomaidis, N. S.
 Identification of biotransformation products of citalopram formed in activated sludge. Water
 Res. 2016, submitted.
- (52) Kosjek, T.; Negreira, N.; de Alda, M. L.; Barcelo, D. Aerobic activated sludge transformation
 of methotrexate: identification of biotransformation products. *Chemosphere* 2015, 119,
 S42–S50.
- 696 (53) Mardal, M.; Meyer, M. R. Studies on the microbial biotransformation of the novel psy-697 choactive substance methylenedioxypyrovalerone (MDPV) in wastewater by means of liquid 698 chromatography-high resolution mass spectrometry/mass spectrometry. *Sci. Total Environ.* 699 **2014**, *493*, 588–595.
- 100 (54) Lange, F.; Cornelissen, S.; Kubac, D.; Sein, M. M.; von Sonntag, J.; Hannich, C. B.; Golloch, A.; Heipieper, H. J.; Möder, M.; von Sonntag, C. Degradation of macrolide antibiotics by ozone: A mechanistic case study with clarithromycin. *Chemosphere* **2006**, *65*, 17 23.
- 703 (55) Rusch, M.; Kauschat, A.; Spielmeyer, A.; Römpp, A.; Hausmann, H.; Zorn, H.; Ham704 scher, G. Biotransformation of the Antibiotic Danofloxacin by Xylaria longipes Leads to an
 705 Efficient Reduction of Its Antibacterial Activity. *J. Agric. Food Chem.* **2015**, *63*, 6897–6904.
- 706 (56) Qu, S.; Kolodziej, E. P.; Long, S. A.; Gloer, J. B.; Patterson, E. V.; Baltrusaitis, J.;

 707 Jones, G. D.; Benchetler, P. V.; Cole, E. A.; Kimbrough, K. C.; Tarnoff, M. D.;

- Cwiertny, D. M. Product-to-Parent Reversion of Trenbolone: Unrecognized Risks for Endocrine Disruption. *Science* **2013**, *342*, 347–351.
- (57) Bonvin, F.; Omlin, J.; Rutler, R.; Schweizer, W. B.; J., A. P.; Strathmann, T. J.; McNeill, K.;
 Kohn, T. Direct Photolysis of Human Metabolites of the Antibiotic Sulfamethoxazole: Evidence for Abiotic Back-Transformation. *Environ. Sci. Technol.* 2013, 47, 6746–6755.