This document is the accepted manuscript version of the following article: Liu, W., Yang, H., Folberth, C., Müller, C., Ciais, P., Abbaspour, K. C., & Schulin, R. (2018). Achieving high crop yields with low nitrogen emissions in global agricultural input intensification. Environmental Science and Technology, 52(23), 13782-13791. https://doi.org/10.1021/acs.est.8b03610

Achieving high crop yields with low nitrogen emissions in global agricultural input intensification

Wenfeng Liu^{1,2,*}, Hong Yang^{1,3}, Christian Folberth⁴, Christoph Müller⁵, Philippe Ciais², Karim C. Abbaspour¹, Rainer Schulin⁶

¹Eawag, Swiss Federal Institute of Aquatic Science and Technology, Ueberlandstrasse 133, CH-8600 Duebendorf, Switzerland.

²Laboratoire des Sciences du Climat et de l'Environnement, LSCE/IPSL, CEA-CNRS-UVSQ,

Université Paris-Saclay, F-91191 Gif-sur-Yvette, France.

³Department of Environmental Sciences, MGU, University of Basel, Petersplatz 1, CH-4003

Basel, Switzerland.

⁴Ecosystem Services and Management Program, International Institute for Applied Systems

Analysis (IIASA), Schlossplatz 1, A-2361 Laxenburg, Austria.

⁵Potsdam Institute for Climate Impact Research, 14473 Potsdam, Germany.

⁶ETH Zürich, Institute of Terrestrial Ecosystems, Universitätstr. 16, CH-8092 Zürich, Switzerland.

*Corresponding author. Email: wenfeng.liu@eawag.ch & wfliu2012@gmail.com

Abstract

1

2

3

4

5

6

7

8

9

10

11

12

13

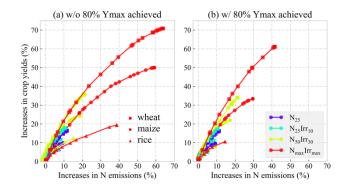
14

15

16

17

Increasing demand for food is driving a worldwide trend of agricultural input intensification. However, there is no comprehensive knowledge about the interrelations between potential yield gains and environmental trade-offs that would enable the identification of regions where input-driven intensification could achieve higher yields, yet with minimal environmental impacts. We explore ways of enhancing global yields, while avoiding significant nitrogen (N) emissions (Ne) by exploring a range of N and irrigation management scenarios. The simulated responses of yields and Ne to increased N inputs (Nin) and irrigation show high spatial variations due to differences in current agricultural inputs and agro-climatic conditions. Nitrogen use efficiency (NUE) of yield gains is negatively correlated with incremental Ne due to N_{in} additions. Avoiding further intensification in regions where high fractions of climatic yield potentials, ≥80%, are already achieved is key to maintain good NUE. Depending on the intensification scenarios, relative increases in N_e could be reduced by 0.3–29.6% of the baseline N_e with this intensification strategy as compared to indiscriminate further intensification, at the cost of a loss of yield increases by 0.2–16.7% of the baseline yields. In addition, irrigation water requirements and N_{in} would dramatically decrease by considering this intensification strategy.



18 TOC Art

1. Introduction

19

20

21

22

23

24

25

26

27

28

29

30

31

32

33

34

35

36

37

38

39

With a continuously growing global population, shifts to more animal-based diets, and possibly increasing competition in agricultural land use between food and biofuel crops, food production has to increase substantially ¹⁻³. It was estimated that global food production needs to be doubled by 2050 if no changes occur on the demand side⁴. There are in principle two different strategies to increasing food production: expansion of croplands and intensification⁵. For the first option, however, suitable land resources are very limited and their conversion to cropland is increasingly constrained by other land-use purposes. In many places, croplands are actually shrinking due to encroachment by urban development^{6,7}. The option of cropland expansion is only feasible in underdeveloped areas and often at high environmental and climate change costs such as greenhouse gas emissions and biodiversity loss from forest clearing⁸. For the majority of countries and regions, intensification of crop production on existing croplands is the only way to meet the increasing food demand. Concerning limited resources and severe environmental impacts, modern agriculture is facing tremendous challenges⁹. A major task faced by scientists, farmers, and policy makers is to find ways to increase global food production while keeping the environmental costs at a tolerable level 10-13. Since the Green Revolution in the 1960s, crop yields have been increased, in particular through the breeding of more productive crop varieties¹⁴ and the intensification of land management with increasing fertilizer inputs and irrigation¹⁵. However, input intensification has also put substantial pressure on the environment due to unbalanced input-output agricultural systems. The mismatch between inputs and outputs not only decreased resource use efficiency¹⁶⁻¹⁹, but also caused serious environmental problems²⁰⁻²⁴. In particular, increasing inputs of nitrogen (N) fertilizers have led to high N emissions to the environment^{25,26}, causing severe eutrophication and drinking water quality deterioration²⁷, as well as air pollution by N₂O and NH₃ emissions²⁸. Hence, there is an ever more pressing need to develop pathways towards input intensification without further compromising environmental health and quality^{4,11,29,30}. Given that the yields of major crops have recently been stagnated or even decreased in many regions of the world, where high agricultural inputs were applied³¹⁻³⁴, it seems that the benefits of additional inputs, particularly fertilizers, are saturating in these regions, suggesting that further input intensification will not help to increase yields. However, the scope for improved efficiency varies among different cropping systems³⁵. It is essential to identify regions where a high yield return can still be obtained at low environmental impacts and to explore how input intensification can be achieved in these regions most efficiently. Previous studies have shown a high potential to further increase crop production through agricultural intensification, and concluded that this is key to increasing global food production to meet the demand³⁶⁻⁴¹. However, these studies have been limited in several aspects, including: a) focusing on one type of input only, either N³⁷ or irrigation³⁶; b) no explicit consideration of environmental impacts^{39,40}; and c) simply using N surpluses to indicate environmental impacts of N pollution without explicitly quantifying different forms of N emissions⁴¹. Here we comprehensively address the agricultural challenges elaborated above by using the global agronomic model PEPIC⁴², Python-based Environmental Policy Integrated Climate (EPIC), to explore the benefits and trade-offs of further global agricultural input intensification

40

41

42

43

44

45

46

47

48

49

50

51

52

53

54

55

56

57

58

59

in terms of increased crop yields and associated N emissions (N_e). We used food production units (FPUs) as a spatial unit. The FPUs are based on river basins and economic regions as introduced by Cai and Rosegrant⁴³ and modified by Kummu et al.⁴⁴ Focusing on N inputs (N_{in}) and irrigation water (Irr), we considered four input intensification scenarios representing low, high and max levels of intensification (two scenarios for the low level) in addition to a baseline scenario with no additional intensification. Three major cereal crops (maize, rice, and wheat) were included in the analysis. In addition to looking into the differences in yields (ΔY), differences in N_e (ΔN_e), and differences in input requirements (ΔN_{in} and ΔIrr) between the intensification and the baseline scenarios, we explored the relationships between ΔY and ΔN_e in response to ΔN_{in} to identify the priority regions for intensification. We also assessed the N use efficiency (NUE) of yield gains as $\Delta Y/\Delta N_{in}$ and N emission intensity of new inputs as $\Delta N_e/\Delta N_{in}$.

2. Materials and methods

2.1. Simulation model and input data

The PEPIC model⁴² was used to simulate crop growth at a daily time step and the associated nutrient dynamics globally at a spatial resolution of 30 arc minutes (about 50 km at the equator). PEPIC has been used for investigating global crop—water relations of maize⁴² and assessing global N and phosphorus (P) emissions from the cultivation of major crops^{12,26}. In addition, it also shows comparable performance in representing yield variability at country level along with 13 other global crop models participating in the Agricultural Model Intercomparison and Improvement Project (AgMIP)^{45,46}. Inputs for PEPIC include longitude, latitude, elevation,

slope, soil properties (e.g. layer depth, pH, bulk density, organic carbon content), climate data (precipitation, temperature, solar radiation, relative humidity, wind speed), land cover, and crop management information. Land cover data are derived from the MIRCA2000 datasets, which provide rainfed and irrigated cultivation areas for 26 crops⁴⁷. As for management data, planting date, harvesting date, fertilizer, and irrigation are required. Planting and harvesting dates were obtained from the Center for Sustainability and the Global Environment (SAGE)⁴⁸. Crop-specific fertilizer and manure application data for the baseline simulation (including N and P) were downloaded from the EarthStat dataset (http://www.earthstat.org/), which were based on Mueller et al.⁴⁹ and West et al.⁵⁰ Information on model performance and other input data is provided in the Supporting Information.

2.2. Management practices and intensification scenarios

In this study, we considered four intensification scenarios— N_{25} , $N_{25}Irr_{10}$, $N_{50}Irr_{30}$ and $N_{max}Irr_{max}$ —in addition to the baseline scenario (Table 1). A sufficient amount of P was applied automatically in each scenario to eliminate the effects of P deficit on plant growth following Folberth et al.⁵¹ N_{in} in the baseline ($N_{in-base}$) was determined by the minimum value between actual N_{in} (N_{in-a}) and automatic N_{in} ($N_{in-auto}$), where N application is based on crop N requirements, for the baseline yield simulation. Here, N_{in-a} was based on the EarthStat dataset, while $N_{in-auto}$ was estimated by PEPIC with automatic N fertilization without N limitation at a trigger value of 10% N stress⁵² under the baseline irrigation cropland condition based on the MIRCA2000 dataset. Such setting excluded the impacts of over N application on N_e in some regions^{26,53}, since this study focused on the effects of input additions. The automatic N

fertilization used here can determine the proper amount of Nin at right time based on soil N concentration and crop N requirements⁵⁴. It deserves to note that the automatic N fertilization may lead to optimistic bias on NUE simulated and that actual management practices are often less targeted (e.g. timing). Irrigation was applied automatically without water limitation for the irrigated cultivation at a trigger value of 10% water stress⁵⁵. Increased N_{in} and conversion of rainfed to irrigated cropland relative to the baseline were considered as intensification scenarios in this study. For N_{in}, we first used PEPIC to determine the maximum N_{in} (N_{in-max}) by using the automatic fertilization schedule setting different N_{in} caps considering all croplands with full irrigation. We found that when N_{in} cap was set to 400 kg N per ha, maximum crop yields (Ymax) could be achieved for the three crops (Table S1). Therefore, actual N_{in} estimated from this condition was used as maximum Nin. Here, Ymax was estimated by forcing the model with sufficient water and N fertilizer without changing the response curves of yields to N and water. This definition of Ymax is consistent with Lassaletta et al. 18,56, Mueller et al. 57, and Mogollón et al. 58, but different from the definition of van Ittersum et al. 59, that also considered changes in the response curves. Combination of this fertilizer and full irrigation was applied in the N_{max}Irr_{max} scenario. Then, the difference between N_{in-max} and $N_{in-base}$, i.e. ΔN_{in-max} , was calculated for each grid cell.

120 Insert Table 1 here

103

104

105

106

107

108

109

110

111

112

113

114

115

116

117

118

119

121

122

123

In N_{25} , N_{in} was increased by 25% of ΔN_{in-max} relative to the baseline, while in $N_{25}Irr_{10}$, 10% of the rainfed cropland was converted to irrigated land in addition to a 25% of ΔN_{in-max} increase in N_{in} . In $N_{50}Irr_{30}$, 30% of the rainfed land was converted to irrigated land in addition to a 50% of

ΔN_{in-max} increase in N_{in}. These levels of intensification were based on Mueller et al.⁴¹, who found that jointly increasing N_{in} by 30% and irrigated lands by 25% would reach productivity levels that represent 75% of the attainable yields of the year 2000. Therefore, we considered an increasing intensification level with an ascending order of baseline, N₂₅, N₂₅Irr₁₀, N₅₀Irr₃₀, and $N_{max}Irr_{max}$. Although global constant percentage values of ΔN_{in-max} and rainfed land were used here to increase N_{in} and irrigation, the actual increases in N_{in} and irrigation were different due to the large differences in N_{in} and rainfed land area under the baseline condition (Figs. S1 and S2) and the input requirements under the N_{max}Irr_{max} scenario. It should be noted that it may not be practical to set the high- and/or max-level input intensification scenarios everywhere across the world. This holds particularly for many poor countries in Africa, as they are currently in lack of infrastructure and/or enough water and fertilizer to reach such intensified agricultural input requirements. However, the scenarios considered here reflect general mechanisms and challenges under the agricultural intensification. Thus, they provide useful information on exploring sustainable pathways towards future agricultural development. They can also be used in land use socio-economic models to prioritize agricultural subsidies and better agricultural practices for increasing N_{in} or N recycling in regions where yields are severely limited by this nutrient.

2.3. Definition of target variables and data analysis

124

125

126

127

128

129

130

131

132

133

134

135

136

137

138

139

140

141

142

143

144

Four model outputs were considered in the analysis, i.e. irrigation water (Irr [mm]), N inputs (N_{in} [kg N per ha]), yields (Y [t per ha]), and N emissions (N_e [kg N per ha]). In this study, N_e refers to total N emissions, including N emissions to the aquatic and atmospheric environments,

estimated by Eq. 2 in the Supporting Information. Crop growth was simulated separately under rainfed and irrigated conditions. Therefore, there are two response curves of yields and Ne to N_{in} for irrigated and rainfed cultivations. The combined outputs of each variable were calculated using the area-weighted average of irrigated and rainfed outputs based on the MIRCA2000 dataset of crop-specific fractions of irrigated and rainfed land for each crop in each grid cell⁴⁷. The combined results were aggregated to FPUs, continental, and global levels. Subsequently, the differences of each variable between the intensification scenarios and the baseline were calculated, i.e. differences in Irr (Δ Irr), differences in N_{in} (ΔN_{in}), differences in Y (ΔY), and differences in N_e (ΔN_e). For FPUs with $\Delta Y < 0.05$ t per ha, outputs under the intensification scenarios were treated as the same values of the baseline to exclude possible errors due to minor responses. After this treatment, ΔIrr , ΔN_{in} , ΔY , and ΔN_e were re-evaluated. The NUE of yield gains, defined as ratio $\Delta Y/\Delta N_{in}$, and N emission intensity of new inputs, defined as ratio $\Delta N_e/\Delta N_{in}$, were calculated to explore the responses of yields and N_e to N_{in} additions under different intensification scenarios. We constructed frontier lines of cumulative ΔY to cumulative ΔN_e by intensifying the FPUs one by one. We first focused on the FPUs with low N emission intensity of new inputs and high NUE of yield gains and then on the FPUs with high N emission intensity of new inputs and low NUE of yield gains. It means that we logically intensified the FPUs with an ascending order of N emission intensity of new inputs and a descending order of NUE of yield gains. In addition to intensifying all FPUs to high/max levels, we investigated the impacts of stopping intensification in FPUs where a target yield, defined by a given fraction of Ymax, is achieved.

145

146

147

148

149

150

151

152

153

154

155

156

157

158

159

160

161

162

163

164

Different fractions (70%, 75%, 80%, 85%, 90%, and 95%) of Ymax were tested. If a given yield target was already achieved in an FPU under scenarios of baseline, N₂₅, N₂₅Irr₁₀, N₅₀Irr₃₀ scenarios, we stopped further intensification in that FPU.

3. Results

166

167

168

169

170

171

172

173

174

175

176

177

178

179

180

181

182

183

184

185

186

3.1. Yield benefits

The simulated responses of crop yields to increased N_{in} and irrigation showed high spatial variations, with very large effects in some regions, but only minor ones in other areas, mainly reflecting current limitations (Fig. 1). These benefits also showed different spatial patterns among the three crops. Areas with high ΔY for maize and rice were found to be concentrated mainly in the southern parts of Africa and South America, where the baseline yields were quite low and differences in yields between the N_{max}Irr_{max} and baseline scenarios were very high (Fig. S3). In addition to these regions, parts of East Europe and the Middle East also showed large ΔY for wheat. Wheat presented smaller yield improvements compared to maize and rice. For other regions, particularly China and India, only small ΔY were predicted for the three crops. Globally, the average ΔY in the four intensification scenarios ranged between 10 and 50% of baseline yields for maize, 7 and 19% for rice, and 16 and 71% for wheat (Table 2). Yields responded mainly to increased N_{in} and less to intensified irrigation. Increased N_{in} were found to substantially increase irrigation water use efficiency (defined as yield per unit of applied irrigation water), while the effects of irrigation additions on NUE (defined as yield per unit of applied N_{in}) were quite small at the continental level (Tables S2 and S3). Meanwhile, to achieve high yields, significant increases in the inputs of these resources were required (Table 2), with

 Δ Irr between 0 and 250% of baseline irrigation water and ΔN_{in} between 12 and 138% of baseline N_{in} , depending on scenarios and crops.

189 Insert Fig. 1 here

190 Insert Table 2 here

3.2. Nitrogen emissions

The simulations predicted that yield benefits from N_{in} and irrigation intensification would result in substantial ΔN_e around the world, with increases varying between 9 and 59% of the baseline N_e for maize, 7 and 39% for rice, and from 12 to 64% for wheat among the four scenarios (Table 2). Increases in N_e was particularly high in Africa, Oceania, and South America for maize and rice, and in Africa, Europe, Oceania, and South America for wheat (Fig. 2 and Fig. S4). The overall geographical distribution patterns were similar to those for the respective ΔY , but there were also some differences. Large ΔN_e were predicted in Southeast Asia for all three crops and in the eastern parts of the USA for wheat, while yields were not expected to increase significantly in these regions in response to N_{in} and irrigation additions. While increasing N_{in} always increased N_e , increased irrigation could reduce N_e in some regions relative to the baseline scenario. That is because the plants that grow better under irrigation cultivation take up more N_e .

204 Insert Fig. 2 here

3.3. Relationship between yield increases and nitrogen emissions

The NUE of yield gains, expressed in terms of the ratio $\Delta Y/\Delta N_{in}$, and the N emission intensity of new inputs, expressed as the ratio $\Delta N_e/\Delta N_{in}$, present opposite spatial distribution patterns at

the FPUs level (Figs. S5 and S6). The Southeast Asia and the northern parts of South America showed high N emission intensity of new inputs for maize and rice. The NUE of yield gains showed clear negative linear relationships to the N emission intensity of new inputs for all the three crops, especially for maize and rice under all four intensification scenarios, and for wheat under the N_{max}Irr_{max} scenario (Fig. 3). The position of individual FPU along the relationships between NUE of yield gains and N emission intensity of new inputs was closely related to the magnitude of N_{in} additions. In contrast, the increases in irrigation land had little effect on the balance between relative benefits on yield vs. N_e, as these relationships showed little differences between the scenarios without and with irrigation addition, for instance between N₂₅ and N₂₅Irr₁₀. The negative relationships shown in Fig. 3 indicate that there is a win–win situation with higher yields and low N_e if further input intensification concentrates on regions where additional inputs have the highest NUE of yield gains, as these are also the regions with the smallest additional N_e.

Insert Fig. 3 here

In most FPUs with a high N emission intensify of new inputs and a low NUE of yield gains (Figs. S5 and S6), yields under each intensification scenario have already obtained a high fraction (e.g. 80%) of Ymax under the former scenario considering an order of baseline, N₂₅, N₂₅Irr₁₀, N₅₀Irr₃₀ and N_{max}Irr_{max} (Fig. 3). This means that significant N_e can be avoided by stopping further increases in N_{in} and irrigation applications in such FPUs. This condition was found to be especially common for rice cultivation, where about 79% of the croplands presently under rice cultivation already reaches 80% of Ymax in the baseline scenario (Table S4). For

wheat and maize, the respective percentages were 35% and 26%. The regions showing low yield gap were mainly located in East Asia, central Europe and the eastern USA for maize, eastern and southeastern Asia for rice, and China and India for wheat (Fig. S7). In the N₅₀Irr₃₀ scenario, the fraction of croplands cannot achieve 80% of the Ymax is still big for maize and wheat.

3.4. Frontier lines for intensifying croplands on a global scale

Based on the negative relationship between NUE of yield gains and N emission intensity of new inputs as shown in Fig. 3, we constructed the distributions of cumulative ΔY against cumulative ΔN_e (Fig. 4), sorting the FPUs by an ascending order in N emission intensity of new inputs and meanwhile a descending order in NUE of yield gains. Starting with FPUs low in ΔN_e and high in ΔY which are the best regions for intensification, these lines decrease in slope as ΔY decreases, while ΔN_e increases. This is particularly notable for scenarios $N_{50}Irr_{30}$ and $N_{max}Irr_{max}$. We observed that initial increases in yields can be achieved by lower-input systems at less environmental burden (e.g. yellow line steeper than red line for rice and initially for wheat and maize). Furthermore, the frontier lines show a large difference between the $N_{max}Irr_{max}$ and the other scenarios, indicating the high N pollution versus small yield benefits of further intensifying crop production when the level of intensification is already high.

Insert Fig. 4 here

An important strategy for limiting ΔN_e with minimal compromises on yield increases is to avoid further intensification in FPUs, where high fractions of Ymax have been achieved, which we here define as a threshold of 80% (Fig. 4). With this condition, simulated ΔN_e were reduced

from 6.4-40.4 kg N per ha to 6.2-20.2 kg N per ha (this reduction representing 0.3-29.6% of the baseline N_e) for maize; from 5.4–28.7 kg N per ha to 3.7–10.7 kg N per ha (the reduction representing 2.2–24.3% of the baseline N_e) for rice; and from 5.1–27.7 kg N per ha to 4.9–18.1 kg N per ha (the reduction representing 0.4–22.3% of the baseline N_e) for wheat, depending on the scenarios. Expected ΔY were reduced by only 0.0–0.9 t per ha (corresponding to 0.2–16.7% of the baseline yields) for maize; 0.1–0.4 t per ha (corresponding to 1.8–8.8% of the baseline yields) for rice; and 0.0-0.3 t per ha (corresponding to 0.2-9.8% of the baseline yields) for wheat (Tables 2 and 3). Furthermore, the results also show that much less N and irrigation water resources are required to achieve the same increases in yield with a strategy in which further intensification is limited to cropland where the yield gap is still comparatively large than with indiscriminate further intensification. In addition to restricting intensification to FPUs with yield levels of <80% Ymax, we performed the analogous analysis setting the restriction fractions at 70%, 75%, 85%, 90%, and 95% of Ymax (Fig. S8). Similar trends for ΔY and ΔN_e were found for these fractions as with the fraction of 80%. However, the trade-offs between Y achievements and increases in N_e varied by using the different yield target levels.

265 Insert Table 3 here

4. Discussion

250

251

252

253

254

255

256

257

258

259

260

261

262

263

264

266

267

268

269

270

To detail the different responses of yields and N_e to N_{in} and water additions, the response curves of yields and N_e to N_{in} under irrigated and rainfed cultivations are displayed for eight continental-climate regions in Fig. 5. The climate classification is based on the Köppen-Geiger climate map⁶⁰. We observed very different responses in different regions, consistent with the

EPIC simulations from Balkovič et al.⁶¹ The differences between the irrigated and rainfed cultivations are particularly high in the arid regions, where the irrigation water requirements are generally high. These response curves from PEPIC highlight the large spatial variability of yield benefits and N_e trade-offs under intensification scenarios on a global scale. Comparison of the PEPIC response curves with those from other crop models would help to estimate the uncertainty on our results arising from the use of a single model.

271

272

273

274

275

276

277

278

279

280

281

282

283

284

285

286

287

288

289

290

291

Insert Fig. 5 here

The high spatial heterogeneity of the responses of yields and N_e to increased N_{in} and irrigation inputs is mainly explained by differences in these inputs between under the baseline scenario and under the N_{max}Irr_{max} scenario. Regions with a high level of baseline N_{in} (Fig. S1) generally respond with low NUE of yield gains but high N emission intensity of new inputs to further intensification. This is in agreement with previous studies. When Nin is already high, the yield benefits of additional N_{in} are low¹⁸, while the potential of N_e tends to be high¹⁴. Maximum N requirements for achieving Ymax presented high variation (Fig. S1) and were mainly determined by climate condition, particularly the potential heat units (Fig. S9), as both of them show quite similar spatial distribution. Such patterns of maximum N_{in} also partially explain the heterogeneity of yield responses to N_{in}. Compared to N_{in}, most regions showed relatively low ΔY in response to more irrigation in all three crops, with significant variations across regions (Fig. S10), which is true as most cultivation regions of these three crops were not found limited by irrigation⁴⁹. Especially, yield response is more sensitive to N_{in} additions than to irrigation in Africa and South America, where the baseline Nin was generally very low. The different responses of yields to irrigation additions are mainly due to different fractions of irrigated lands to total cultivated lands (Fig. S2) and the differences in irrigation requirements under different scenarios, e.g. the baseline management as shown in Fig. S11. A low irrigation requirement indicates that growing season precipitation is generally sufficient for crop growth (Fig. S12), and so there are low benefits from increased irrigation. On the other hand, there is only limited opportunity to further increasing yields by taking additional land under irrigation where a large fraction of the cropland is already under irrigation. On land where yields are already close to what can be achieved with the maximum N and irrigation water inputs, further increasing the intensity can produce only minor benefits, while the environmental impacts and the consumption of resources become more significant. Such regions need to be identified and excluded from further intensification. Based on our simulations, it was not possible to double the production of the three crops on the current cropland area even if there was no limit on N fertilizer application and irrigation, especially in rice production (Table 2). Similar findings were reported by Mueller et al.⁴⁹, who estimated production potentials close to 100% of maximum attainable yields through nutrients and irrigation management. Therefore, other measures need to be considered to further increase yields, such as breeding more productive varieties⁶², change of cropping intensities⁶³, and better allocation on a global scale of crops to cultivated land⁴⁰, among others. Combining these measures with better N and irrigation management can be expected to further increase Ymax and strengthen the responses of yields to Nin and irrigation in currently high yield regions, as elaborated in Mogollón et al.⁵⁸ Besides, for a better agricultural development, extensification,

292

293

294

295

296

297

298

299

300

301

302

303

304

305

306

307

308

309

310

311

which focuses on reducing external inputs, could also be considered in the already intensified regions to alleviate environmental impacts with minor yield reduction⁶⁴. Nevertheless, here we focused on the three major food crops. It is worth noting that other types of crops (particularly fruits and vegetable) may hold the potential to double the production⁶⁵. For instance, apple yields in China could increase from 16.5 t per ha to 37 t per ha through better management⁶⁶. Hence, a comprehensive assessment including different management options and incorporating more crops should be conducted in further research. Agricultural inputs cannot be increased without limit and they are also associated with significant environmental impacts. While there is sufficient N in the atmosphere for practically unlimited industrial production of mineral N fertilizer, the energy required for it is likely to set an upper limit. Already about 2% of the world's energy use is for the production of reactive N⁶⁷. Moreover, there is still a long way to go for many developing countries, mainly in Africa, to afford and distribute enough nutrients for their croplands. In comparison to N fertilizer, freshwater is a much more limited resource⁶⁸ and its spatiotemporal distribution is already very uneven for natural reasons⁶⁹. That is why we set lower levels of irrigated land expansion. While we found that by expanding the irrigated areas, maize and wheat yields could be increased substantially in the western USA and western and central Asia (Fig. S10), the available water resources there set a rather low upper limit to this option⁷⁰. Further expansion of irrigation agriculture hence bears the risk of worsening this problem. Therefore, it will be vital to increase the efficiency of irrigation^{71,72} and fertilizer¹⁹ applications. As shown here, one strategy to achieve this is to concentrate intensification efforts in regions with currently low yields and

313

314

315

316

317

318

319

320

321

322

323

324

325

326

327

328

329

330

331

332

high yield potentials, while avoiding further intensification in regions where yields are already very high. This strategy could simultaneously conserve agricultural resources and also substantially reduce additional N pollution. Mueller et al. 41 also explored trade-offs between excess N (differences between Nin and N in crop harvest) and crop production using frontier lines. The difference between our results and their study is that we used frontier lines to explore the benefits of avoiding further 'useless' intensification in regions where yields are already close to their maximum levels, i.e. 70-95% of Ymax (Fig. 4 and Fig. S8). The ΔN_e may be reduced if setting a low yield target. But then the increases in yields will also be lower. Thus, it is important to find a level at which these effects are in a desirable balance. However, setting such a target level is subject to local and regional policies, social and economic trade-offs, access to N fertilizers, capacity to improve agronomical practices, as well as consideration of regional environmental vulnerabilities. Due to significant variations of agricultural performance among different countries in terms of trade-offs between yield achievements and environmental impacts because of different natural conditions (e.g. climate and soil) and agronomic practices (e.g. different NUE), much more efforts should be made to determine a reasonable yield target level on local scale^{57,73}. Due to enormous disparities in capacity and socioeconomic conditions among different countries, it may not be possible to intensify agricultural inputs to such a reasonable yield target level in the short term in economically week regions, e.g. many poor countries in Africa. Therefore, in addition to setting a critical target level, it is also important to make sure that the high yield improvements without significant environmental costs could be practically achieved for these

334

335

336

337

338

339

340

341

342

343

344

345

346

347

348

349

350

351

352

353

economic models to address benefits and trade-offs between agricultural intensification and food trade-dependence^{53,73}. This is beyond the focus of the current study, but it certainly deserves an in-depth investigation. Our previous analysis showed that the PEPIC model realistically captured large-scale N_e due to N inputs, but with high uncertainties from model parameters²⁶. In this study, we document the variable responses of yields and Ne to Nin and water additions in different regions across the world. More detailed follow up studies to significantly reduce uncertainties should focus on evaluating the PEPIC response against data from field experiments with nutrient and water treatments 74,75, and on comparing PEPIC response curves with other crop models, such as planned in AgMIP⁷⁶. We did not consider the impacts of climate change and adaptation strategies⁷⁷, which are also important factors affecting future food security, although their uncertainties are also high^{55,78,79}. As the purpose of this study is to investigate the yield potentials of input intensification, associated environmental trade-offs, and their regional differences, we only considered four intensification scenarios (representing low, high, and maximum levels) in order to reduce the computation time. A full range of N and irrigation intensification scenarios would help to identify the optimal intensification level. Overcoming the limitations specified above was beyond the scope of this study and will be subject to future research. Supporting Information: Details on model description and model performance. Details on nitrogen inputs, crop yield responses, nitrogen emission responses, and percentage of potential

regions. Answering this question requires detailed socio-economic information and complex

355

356

357

358

359

360

361

362

363

364

365

366

367

368

369

370

371

372

373

374

- 376 yields achieved under difference scenarios. Details on irrigated land, potential heat units,
- 377 irrigation water requirements, and growing season precipitation under baseline. Details on
- 378 yield-nitrogen responses, water use efficiency, nitrogen use efficiency at the continental level.
- 379 Author contributions: W.L. and H.Y. designed research. W.L. performed research and
- analysed data. All authors participated in the interpretation of results and the writing and
- 381 editing process.
- 382 **Acknowledgments:** This study was supported by funding from the Swiss Federal Institute
- of Aquatic Science and Technology (Eawag) and the World Food System Center at ETH Zürich,
- 384 Switzerland. W.L. acknowledges the support received from the Early Postdoctoral Mobility
- 385 Fellowship awarded by Swiss National Science Foundation (P2EZP2 175096). C.F. was
- 386 supported by European Research Council Synergy grant ERC-2013-SynG-610028 Imbalance-P.
- 387 C.M. acknowledges financial support from the MACMIT project (01LN1317A) funded
- through the German Federal Ministry of Education and Research (BMBF).
- 389 **Competing interests:** The authors declare that they have no competing financial interests.
- 390 **References**
- 391 1. Bodirsky, B. L.; Popp, A.; Lotze-Campen, H.; Dietrich, J. P.; Rolinski, S.; Weindl, I.;
- 392 Schmitz, C.; Muller, C.; Bonsch, M.; Humpenoder, F.; Biewald, A.; Stevanovic, M., Reactive
- 393 nitrogen requirements to feed the world in 2050 and potential to mitigate nitrogen pollution.
- 394 *Nat. Commun.* **2014,** *5*, 3858.
- 395 2. Bodirsky, B. L.; Rolinski, S.; Biewald, A.; Weindl, I.; Popp, A.; Lotze-Campen, H.,
- Global food demand scenarios for the 21st century. *PLoS One* **2015**, *10*, (11), e0139201.

- 397 3. Lotze-Campen, H.; von Lampe, M.; Kyle, P.; Fujimori, S.; Havlik, P.; van Meijl, H.;
- 398 Hasegawa, T.; Popp, A.; Schmitz, C.; Tabeau, A.; Valin, H.; Willenbockel, D.; Wise, M.,
- 399 Impacts of increased bioenergy demand on global food markets: an AgMIP economic model
- 400 intercomparison. Agr. Econ. **2014**, 45, (1), 103-116.
- 401 4. Tilman, D.; Balzer, C.; Hill, J.; Befort, B. L., Global food demand and the sustainable
- 402 intensification of agriculture. Proc. Natl. Acad. Sci. U.S.A. 2011, 108, (50), 20260-20264.
- 403 5. Matson, P. A.; Parton, W. J.; Power, A. G.; Swift, M. J., Agricultural intensification and
- 404 ecosystem properties. *Science* **1997**, 277, (5325), 504-509.
- 405 6. Deng, J. S.; Qiu, L. F.; Wang, K.; Yang, H.; Shi, Y. A. Y., An integrated analysis of
- 406 urbanization-triggered cropland loss trajectory and implications for sustainable land
- 407 management. Cities **2011**, 28, (2), 127-137.
- 408 7. Lambin, E. F.; Meyfroidt, P., Global land use change, economic globalization, and the
- 409 looming land scarcity. *Proc. Natl. Acad. Sci. U.S.A.* **2011**, *108*, (9), 3465-3472.
- 8. Foley, J. A.; Ramankutty, N.; Brauman, K. A.; Cassidy, E. S.; Gerber, J. S.; Johnston, M.;
- Mueller, N. D.; O'Connell, C.; Ray, D. K.; West, P. C.; Balzer, C.; Bennett, E. M.; Carpenter,
- S. R.; Hill, J.; Monfreda, C.; Polasky, S.; Rockstrom, J.; Sheehan, J.; Siebert, S.; Tilman, D.;
- 413 Zaks, D. P. M., Solutions for a cultivated planet. *Nature* **2011**, *478*, (7369), 337-342.
- 9. Godfray, H. C. J.; Beddington, J. R.; Crute, I. R.; Haddad, L.; Lawrence, D.; Muir, J. F.;
- Pretty, J.; Robinson, S.; Thomas, S. M.; Toulmin, C., Food Security: The challenge of feeding
- 416 9 billion people. *Science* **2010**, *327*, (5967), 812-818.
- 10. Foley, J. A.; DeFries, R.; Asner, G. P.; Barford, C.; Bonan, G.; Carpenter, S. R.; Chapin, F.

- 418 S.; Coe, M. T.; Daily, G. C.; Gibbs, H. K.; Helkowski, J. H.; Holloway, T.; Howard, E. A.;
- Kucharik, C. J.; Monfreda, C.; Patz, J. A.; Prentice, I. C.; Ramankutty, N.; Snyder, P. K.,
- 420 Global consequences of land use. *Science* **2005**, *309*, (5734), 570-574.
- 421 11. Godfray, H. C. J.; Garnett, T., Food security and sustainable intensification. *Phil. Trans.*
- 422 *R. Soc. B* **2014,** *369*, (1639), 20120273.
- 423 12. Liu, W.; Yang, H.; Ciais, P.; Stamm, C.; Zhao, X.; Williams, J. R.; Abbaspour, K. C.;
- 424 Schulin, R., Integrative crop-soil-management modeling to assess global phosphorus losses
- from major crop cultivations. Global Biogeochem. Cycles 2018, 32, (7), 1074-1086.
- 426 13. Sayer, J.; Cassman, K. G., Agricultural innovation to protect the environment. *Proc. Natl.*
- 427 Acad. Sci. U.S.A. 2013, 110, (21), 8345-8348.
- 428 14. Bodirsky, B. L.; Muller, C., Robust relationship between yields and nitrogen inputs
- indicates three ways to reduce nitrogen pollution. *Environ. Res. Lett.* **2014**, *9*, (11), 111005.
- 430 15. Tilman, D.; Fargione, J.; Wolff, B.; D'Antonio, C.; Dobson, A.; Howarth, R.; Schindler,
- D.; Schlesinger, W. H.; Simberloff, D.; Swackhamer, D., Forecasting agriculturally driven
- 432 global environmental change. *Science* **2001**, 292, (5515), 281-284.
- 433 16. Conant, R. T.; Berdanier, A. B.; Grace, P. R., Patterns and trends in nitrogen use and
- nitrogen recovery efficiency in world agriculture. Global Biogeochem. Cycles 2013, 27, (2),
- 435 558-566.
- 436 17. Cui, Z. L.; Wang, G. L.; Yue, S. C.; Wu, L.; Zhang, W. F.; Zhang, F. S.; Chen, X. P.,
- 437 Closing the N-use efficiency gap to achieve food and environmental security. Environ. Sci.
- 438 *Technol.* **2014,** *48*, (10), 5780-5787.

- 439 18. Lassaletta, L.; Billen, G.; Grizzetti, B.; Anglade, J.; Garnier, J., 50 year trends in nitrogen
- 440 use efficiency of world cropping systems: the relationship between yield and nitrogen input to
- 441 cropland. Environ. Res. Lett. 2014, 9, (10), 105011.
- 19. Zhang, X.; Davidson, E. A.; Mauzerall, D. L.; Searchinger, T. D.; Dumas, P.; Shen, Y.,
- Managing nitrogen for sustainable development. *Nature* **2015**, *528*, (7580), 51-59.
- 20. Clark, C. M.; Tilman, D., Loss of plant species after chronic low-level nitrogen
- deposition to prairie grasslands. *Nature* **2008**, *451*, (7179), 712-715.
- 21. Liu, J.; You, L.; Amini, M.; Obersteiner, M.; Herrero, M.; Zehnder, A. J.; Yang, H., A
- high-resolution assessment on global nitrogen flows in cropland. *Proc. Natl. Acad. Sci. U.S.A.*
- 448 **2010,** *107*, (17), 8035-40.
- 22. Liu, X.; Zhang, Y.; Han, W.; Tang, A.; Shen, J.; Cui, Z.; Vitousek, P.; Erisman, J. W.;
- Goulding, K.; Christie, P.; Fangmeier, A.; Zhang, F., Enhanced nitrogen deposition over China.
- 451 *Nature* **2013**, *494*, (7438), 459-62.
- 452 23. Schlesinger, W. H., On the fate of anthropogenic nitrogen. Proc. Natl. Acad. Sci. U.S.A.
- **2009**, *106*, (1), 203-208.
- 454 24. Steffen, W.; Richardson, K.; Rockstrom, J.; Cornell, S. E.; Fetzer, I.; Bennett, E. M.;
- Biggs, R.; Carpenter, S. R.; de Vries, W.; de Wit, C. A.; Folke, C.; Gerten, D.; Heinke, J.;
- 456 Mace, G. M.; Persson, L. M.; Ramanathan, V.; Reyers, B.; Sorlin, S., Planetary boundaries:
- Guiding human development on a changing planet. Science 2015, 347, (6223), 1259855.
- 458 25. Babbin, A. R.; Ward, B. B., Controls on nitrogen loss processes in Chesapeake Bay
- 459 sediments. Environ. Sci. Technol. 2013, 47, (9), 4189-96.

- 460 26. Liu, W.; Yang, H.; Liu, J.; Azevedo, L. B.; Wang, X.; Xu, Z.; Abbaspour, K. C.; Schulin,
- 461 R., Global assessment of nitrogen losses and trade-offs with yields from major crop
- 462 cultivations. Sci. Total Environ. 2016, 572, 526-537.
- 463 27. Liu, W.; Antonelli, M.; Liu, X.; Yang, H., Towards improvement of grey water footprint
- assessment: With an illustration for global maize cultivation. *J. Cleaner Prod.* **2017**, *147*, 1-9.
- 28. Cui, Z. L.; Yue, S. C.; Wang, G. L.; Zhang, F. S.; Chen, X. P., In-season root-zone N
- 466 management for mitigating greenhouse gas emission and reactive N losses in intensive wheat
- 467 production. Environ. Sci. Technol. 2013, 47, (11), 6015-6022.
- 468 29. Garnett, T.; Appleby, M. C.; Balmford, A.; Bateman, I. J.; Benton, T. G.; Bloomer, P.;
- Burlingame, B.; Dawkins, M.; Dolan, L.; Fraser, D.; Herrero, M.; Hoffmann, I.; Smith, P.;
- Thornton, P. K.; Toulmin, C.; Vermeulen, S. J.; Godfray, H. C., Agriculture. Sustainable
- intensification in agriculture: premises and policies. *Science* **2013**, *341*, (6141), 33-4.
- 30. Pretty, J., Agricultural sustainability: concepts, principles and evidence. *Phil. Trans. R.*
- 473 *Soc. B* **2008**, *363*, (1491), 447-465.
- 474 31. Iizumi, T.; Yokozawa, M.; Sakurai, G.; Travasso, M. I.; Romanernkov, V.; Oettli, P.;
- Newby, T.; Ishigooka, Y.; Furuya, J., Historical changes in global yields: major cereal and
- 476 legume crops from 1982 to 2006. *Global Ecol. Biogeogr.* **2014**, *23*, (3), 346-357.
- 477 32. Ray, D. K.; Mueller, N. D.; West, P. C.; Foley, J. A., Yield trends are insufficient to
- double global crop production by 2050. *PLoS One* **2013**, *8*, (6), e66428.
- 33. Ray, D. K.; Ramankutty, N.; Mueller, N. D.; West, P. C.; Foley, J. A., Recent patterns of
- 480 crop yield growth and stagnation. *Nat. Commun.* **2012**, *3*, 5989.

- 481 34. Wei, X.; Zhang, Z.; Shi, P. J.; Wang, P.; Chen, Y.; Song, X.; Tao, F. L., Is yield increase
- sufficient to achieve food security in China? *PLoS One* **2015**, *10*, (2), e0116430.
- 483 35. Carberry, P. S.; Liang, W. L.; Twomlow, S.; Holzworth, D. P.; Dimes, J. P.; McClelland,
- 484 T.; Huth, N. I.; Chen, F.; Hochman, Z.; Keating, B. A., Scope for improved eco-efficiency
- varies among diverse cropping systems. Proc. Natl. Acad. Sci. U.S.A. 2013, 110, (21),
- 486 8381-8386.
- 487 36. Brauman, K. A.; Siebert, S.; Foley, J. A., Improvements in crop water productivity
- 488 increase water sustainability and food security—a global analysis. *Environ Res Lett* **2013,** 8,
- 489 (2), 024030.
- 490 37. Chen, X.; Cui, Z.; Fan, M.; Vitousek, P.; Zhao, M.; Ma, W.; Wang, Z.; Zhang, W.; Yan,
- 491 X.; Yang, J.; Deng, X.; Gao, Q.; Zhang, Q.; Guo, S.; Ren, J.; Li, S.; Ye, Y.; Wang, Z.; Huang,
- 492 J.; Tang, Q.; Sun, Y.; Peng, X.; Zhang, J.; He, M.; Zhu, Y.; Xue, J.; Wang, G.; Wu, L.; An, N.;
- Wu, L.; Ma, L.; Zhang, W.; Zhang, F., Producing more grain with lower environmental costs.
- 494 Nature **2014**, 514, (7523), 486-9.
- 495 38. Chen, X. P.; Cui, Z. L.; Vitousek, P. M.; Cassman, K. G.; Matson, P. A.; Bai, J. S.; Meng,
- 496 Q. F.; Hou, P.; Yue, S. C.; Romheld, V.; Zhang, F. S., Integrated soil-crop system management
- 497 for food security. *Proc. Natl. Acad. Sci. U.S.A.* **2011,** *108*, (16), 6399-6404.
- 498 39. Johnston, M.; Licker, R.; Foley, J.; Holloway, T.; Mueller, N. D.; Barford, C.; Kucharik,
- 499 C., Closing the gap: global potential for increasing biofuel production through agricultural
- 500 intensification. *Environ. Res. Lett.* **2011**, *6*, (3), 034028.
- 40. Mauser, W.; Klepper, G.; Zabel, F.; Delzeit, R.; Hank, T.; Putzenlechner, B.; Calzadilla,

- A., Global biomass production potentials exceed expected future demand without the need for
- 503 cropland expansion. Nat. Commun. 2015, 6, 8946.
- 41. Mueller, N. D.; West, P. C.; Gerber, J. S.; MacDonald, G. K.; Polasky, S.; Foley, J. A., A
- tradeoff frontier for global nitrogen use and cereal production. *Environ Res Lett* **2014**, 9, (5),
- 506 054002.
- 507 42. Liu, W. F.; Yang, H.; Folberth, C.; Wang, X. Y.; Luo, Q. Y.; Schulin, R., Global
- 508 investigation of impacts of PET methods on simulating crop-water relations for maize. Agr.
- 509 Forest Meteorol. **2016**, 221, 164-175.
- 510 43. Cai, X. M.; Rosegrant, M. W., Global water demand and supply projections part 1. A
- 511 modeling approach. Water Int. **2002**, 27, (2), 159-169.
- 512 44. Kummu, M.; Ward, P. J.; de Moel, H.; Varis, O., Is physical water scarcity a new
- 513 phenomenon? Global assessment of water shortage over the last two millennia. Environ. Res.
- 514 *Lett.* **2010,** *5*, (3), 034006.
- 45. Müller, C.; Elliott, J.; Chryssanthacopoulos, J.; Arneth, A.; Balkovic, J.; Ciais, P.; Deryng,
- D.; Folberth, C.; Glotter, M.; Hoek, S.; Iizumi, T.; Izaurralde, R. C.; Jones, C.; Khabarov, N.;
- Lawrence, P.; Liu, W.; Olin, S.; Pugh, T. A. M.; Ray, D.; Reddy, A.; Rosenzweig, C.; Ruane,
- 518 A. C.; Sakurai, G.; Schmid, E.; Skalsky, R.; Song, C. X.; Wang, X.; de Wit, A.; Yang, H.,
- Global gridded crop model evaluation: benchmarking, skills, deficiencies and implications.
- 520 Geosci. Model Dev. **2017**, 10, 1403-1422.
- 521 46. Porwollik, V.; Müller, C.; Elliott, J.; Chryssanthacopoulos, J.; Iizumi, T.; Ray, D. K.;
- Ruane, A. C.; Arneth, A.; Balkovič, J.; Ciais, P.; Deryng, D.; Folberth, C.; Izaurralde, R. C.;

- Jones, C. D.; Khabarov, N.; Lawrence, P. J.; Liu, W.; Pugh, T. A. M.; Reddy, A.; Sakurai, G.;
- 524 Schmid, E.; Wang, X.; de Wit, A.; Wu, X., Spatial and temporal uncertainty of crop yield
- 525 aggregations. Eur. J. Agron. **2017**, 88, 10-21.
- 526 47. Portmann, F. T.; Siebert, S.; Doll, P., MIRCA2000-Global monthly irrigated and rainfed
- 527 crop areas around the year 2000: A new high-resolution data set for agricultural and
- 528 hydrological modeling. Global Biogeochem. Cycles 2010, 24.
- 529 48. Sacks, W. J.; Deryng, D.; Foley, J. A.; Ramankutty, N., Crop planting dates: an analysis
- of global patterns. *Global Ecol. Biogeogr.* **2010**, *19*, (5), 607-620.
- 49. Mueller, N. D.; Gerber, J. S.; Johnston, M.; Ray, D. K.; Ramankutty, N.; Foley, J. A.,
- Closing yield gaps through nutrient and water management. *Nature* **2012**, 490, (7419),
- 533 254-257.
- 50. West, P. C.; Gerber, J. S.; Engstrom, P. M.; Mueller, N. D.; Brauman, K. A.; Carlson, K.
- 535 M.; Cassidy, E. S.; Johnston, M.; MacDonald, G. K.; Ray, D. K.; Siebert, S., Leverage points
- for improving global food security and the environment. *Science* **2014**, *345*, (6194), 325-328.
- 51. Folberth, C.; Yang, H.; Gaiser, T.; Liu, J.; Wang, X.; Williams, J.; Schulin, R., Effects of
- 538 ecological and conventional agricultural intensification practices on maize yields in
- sub-Saharan Africa under potential climate change. *Environ. Res. Lett.* **2014,** *9*, (4), 044004.
- 540 52. Balkovič, J.; van der Velde, M.; Skalský, R.; Xiong, W.; Folberth, C.; Khabarov, N.;
- 541 Smirnov, A.; Mueller, N. D.; Obersteiner, M., Global wheat production potentials and
- 542 management flexibility under the representative concentration pathways. Global Planet.
- 543 *Change* **2014,** *122*, 107-121.

- 53. Liu, W.; Yang, H.; Liu, Y.; Kummu, M.; Hoekstra, A. Y.; Liu, J.; Schulin, R., Water
- 545 resources conservation and nitrogen pollution reduction under global food trade and
- agricultural intensification. Sci. Total Environ. 2018, 633, 1591-1601.
- 54. Williams, J. R., The EPIC model. In Computer Models of Watershed hydrology, Singh, V.
- P., Ed. Water Resources Publications: Highlands Ranch, Colo, 1995.
- 55. Rosenzweig, C.; Elliott, J.; Deryng, D.; Ruane, A. C.; Muller, C.; Arneth, A.; Boote, K. J.;
- Folberth, C.; Glotter, M.; Khabarov, N.; Neumann, K.; Piontek, F.; Pugh, T. A. M.; Schmid, E.;
- 551 Stehfest, E.; Yang, H.; Jones, J. W., Assessing agricultural risks of climate change in the 21st
- century in a global gridded crop model intercomparison. Proc. Natl. Acad. Sci. U.S.A. 2014,
- 553 111, (9), 3268-3273.
- 56. Lassaletta, L.; Billen, G.; Garnier, J.; Bouwman, L.; Velazquez, E.; Mueller, N. D.;
- Gerber, J. S., Nitrogen use in the global food system: past trends and future trajectories of
- agronomic performance, pollution, trade, and dietary demand. Environ. Res. Lett. 2016, 11,
- 557 (9).
- 558 57. Mueller, N. D.; Lassaletta, L.; Runck, B. C.; Billen, G.; Garnier, J.; Gerber, J. S.,
- Declining spatial efficiency of global cropland nitrogen allocation. Global Biogeochem.
- 560 *Cycles* **2017**, *31*, (2), 245-257.
- 58. Mogollón, J. M.; Lassaletta, L.; Beusen, A. H. W.; van Grinsven, H. J. M.; Westhoek, H.;
- Bouwman, A. F., Assessing future reactive nitrogen inputs into global croplands based on the
- shared socioeconomic pathways. *Environ. Res. Lett.* **2018**, *13*, (4).
- 59. van Ittersum, M. K.; Cassman, K. G.; Grassini, P.; Wolf, J.; Tittonell, P.; Hochman, Z.,

- Yield gap analysis with local to global relevance—A review. Field Crop Res. 2013, 143, 4-17.
- 566 60. Peel, M. C.; Finlayson, B. L.; McMahon, T. A., Updated world map of the
- Koppen-Geiger climate classification. *Hydrol. Earth Syst. Sci.* **2007**, *11*, (5), 1633-1644.
- 568 61. Balkovič, J.; van der Velde, M.; Schmid, E.; Skalský, R.; Khabarov, N.; Obersteiner, M.;
- Sturmer, B.; Xiong, W., Pan-European crop modelling with EPIC: Implementation, up-scaling
- and regional crop yield validation. Agr. Syst. 2013, 120, 61-75.
- 571 62. Khush, G. S., Green revolution: the way forward. Nat. Rev. Genet. 2001, 2, (10),
- 572 815-822.
- 573 63. Challinor, A. J.; Parkes, B.; Ramirez-Villegas, J., Crop yield response to climate change
- varies with cropping intensity. *Global Change Biol.* **2015**, *21*, (4), 1679-1688.
- 575 64. van Grinsven, H. J. M.; Erisman, J. W.; de Vries, W.; Westhoek, H., Potential of
- extensification of European agriculture for a more sustainable food system, focusing on
- 577 nitrogen. Environ. Res. Lett. 2015, 10, (2), 025002.
- 578 65. Ebert, A., Potential of underutilized traditional vegetables and legume crops to contribute
- 579 to food and nutritional security, income and more sustainable production systems.
- 580 *Sustainability* **2014,** *6*, (1), 319-335.
- 581 66. Wang, N.; Wolf, J.; Zhang, F., Towards sustainable intensification of apple production in
- 582 China Yield gaps and nutrient use efficiency in apple farming systems. Journal of
- 583 *Integrative Agriculture* **2016,** *15*, (4), 716-725.
- 584 67. Sutton, M., A.; Bleeker, A.; Howard, C. M.; Bekunda, M.; Grizzetti, B.; de Vries, W.; van
- Grinsven, H. J., M.; Abrol, Y. P.; Adhya, T. K.; Billen, G.; Davidson, E. A.; Datta, A.; Diaz, R.;

- Erisman, J. W.; Liu, X., J.; Oenema, O.; Palm, C.; Raghuram, N.; Reis, S.; Scholz, R. W.;
- Sims, T.; Westhoek, H.; Zhang, F. S., Our Nutrient World: the challenge to produce more food
- and energy with less pollution. Centre for Ecology and Hydrology (CEH): 2013.
- 589 68. Liu, W.; Antonelli, M.; Kummu, M.; Zhao, X.; Wu, P.; Liu, J.; Zhuo, L.; Yang, H.,
- 590 Savings and losses of global water resources in food-related virtual water trade. WIREs Water
- 591 **2018**, e1320.
- 592 69. Oki, T.; Kanae, S., Global hydrological cycles and world water resources. Science 2006,
- 593 *313*, (5790), 1068-1072.
- 594 70. Elliott, J.; Deryng, D.; Mueller, C.; Frieler, K.; Konzmann, M.; Gerten, D.; Glotter, M.;
- Florke, M.; Wada, Y.; Best, N.; Eisner, S.; Fekete, B. M.; Folberth, C.; Foster, I.; Gosling, S.
- N.; Haddeland, I.; Khabarov, N.; Ludwig, F.; Masaki, Y.; Olin, S.; Rosenzweig, C.; Ruane, A.
- 597 C.; Satoh, Y.; Schmid, E.; Stacke, T.; Tang, Q. H.; Wisser, D., Constraints and potentials of
- future irrigation water availability on agricultural production under climate change. Proc. Natl.
- 599 Acad. Sci. U.S.A. **2014**, 111, (9), 3239-3244.
- 71. Jagermeyr, J.; Gerten, D.; Heinke, J.; Schaphoff, S.; Kummu, M.; Lucht, W., Water
- savings potentials of irrigation systems: global simulation of processes and linkages. *Hydrol*.
- 602 Earth Syst. Sci. 2015, 19, (7), 3073-3091.
- 72. Jagermeyr, J.; Gerten, D.; Schaphoff, S.; Heinke, J.; Lucht, W.; Rockstrom, J., Integrated
- 604 crop water management might sustainably halve the global food gap. Environ. Res. Lett. 2016,
- 605 11, (2), 025002.
- 73. Zhang, X., BIOGEOCHEMISTRY A plan for efficient use of nitrogen fertilizers. *Nature*

- **2017,** *543*, (7645), 322-323.
- 608 74. Li, Z.; Xu, M.; Zhang, H.; Zhang, W.; Gao, J., Grain yield trends of different food crops
- under long-term fertilization in China. Scientia Agricultura Sinica 2009, 42, (7), 2407-2414.
- 610 75. Stewart, W. M.; Dibb, D. W.; Johnston, A. E.; Smyth, T. J., The contribution of
- 611 commercial fertilizer nutrients to food production. Agron. J. 2005, 97, (1), 1-6.
- 76. Elliott, J.; Müller, C.; Deryng, D.; Chryssanthacopoulos, J.; Boote, K.; Büchner, M.;
- 613 Foster, I.; Glotter, M.; Heinke, J.; Iizumi, T., The Global Gridded Crop Model
- Intercomparison: data and modeling protocols for Phase 1 (v1. 0). Geosci. Model Dev. 2015, 8,
- 615 (2), 261-277.
- 77. Sinha, E.; Michalak, A. M.; Balaji, V., Eutrophication will increase during the 21st
- century as a result of precipitation changes. *Science* **2017**, *357*, (6349), 405-408.
- 78. Pugh, T. A.; Muller, C.; Elliott, J.; Deryng, D.; Folberth, C.; Olin, S.; Schmid, E.; Arneth,
- A., Climate analogues suggest limited potential for intensification of production on current
- 620 croplands under climate change. *Nat. Commun.* **2016**, 7, 12608.
- 79. Ray, D. K.; Gerber, J. S.; MacDonald, G. K.; West, P. C., Climate variation explains a
- third of global crop yield variability. *Nat. Commun.* **2015**, *6*, 5989.

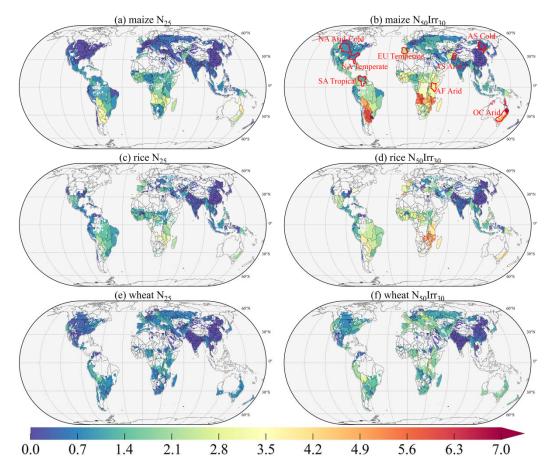


Figure 1. Differences in crops yields [t per ha] between N₂₅ and baseline (a, c, e), as well as between N₅₀Irr₃₀ and baseline (b, d, f). Note: eight continent-climate regions in panel b are used to detail the responses of yields and nitrogen emissions to nitrogen inputs under irrigated and rainfed cultivations. Maps of different food production units were obtained from Kummu et al.⁴⁴ AF: Africa, AS: Asia, EU: Europe, NA: North America, OC: Oceania, SA: South America.

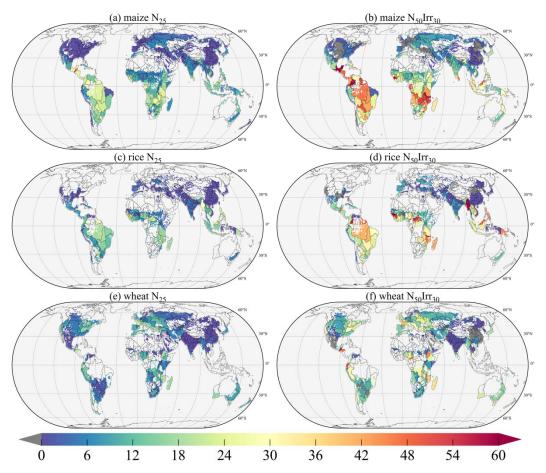


Figure 2. Differences in nitrogen (N) emissions [kg N per ha] between N_{25} and baseline (a, c, e), as well as between $N_{50}Irr_{30}$ and baseline (b, d, f). Maps of different food production units were obtained from Kummu et al.⁴⁴

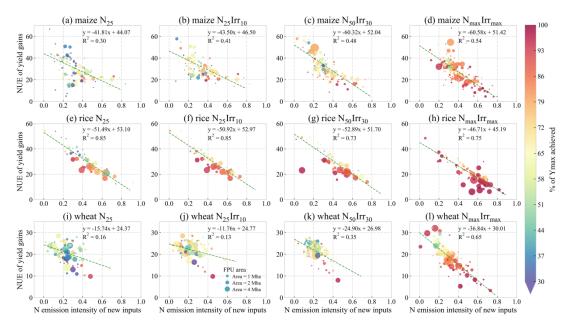


Figure 3. Relationships between nitrogen (N) use efficiency (NUE) of yield gains [kg grain per kg N] and N emission intensity of new inputs [-] under different intensification scenarios at the food production units (FPUs) level. Ymax is estimated yields by PEPIC under the N_{max}Irr_{max} scenario. The FPUs with the smallest areas (for a total of 2% of global total cropland areas of each crop) are not shown. Size represents cropland area for each FPU. Colors represent the fractions to which Ymax have been achieved. For a given intensification scenario, the colors show the situation of its previous scenario with an intensification order of baseline, N₂₅, N₂₅Irr₁₀, N₅₀Irr₃₀ and N_{max}Irr_{max} scenarios. That is to say, for instance, colors in N₂₅ give the achieved Ymax fractions of the baseline scenario. Equations represent the linear relationship between NUE of yield gains and N emission intensity of new inputs (points with negative N emission intensity of new inputs are not included for regression analysis). R² is the coefficient of determination of equation.

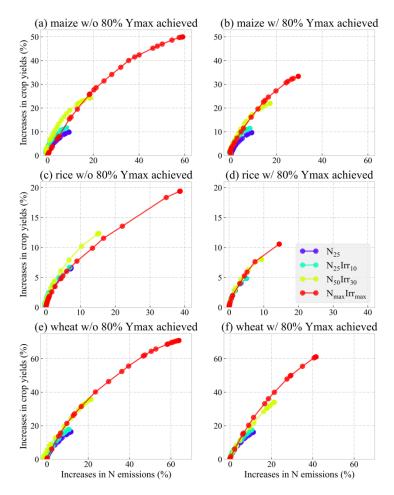


Figure 4. Percentage increases in global average crop yields and nitrogen (N) emissions under different intensification scenarios without (w/o) considering 80% of maximum yields (Ymax) achieved (a, c, e) and with (w/) considering 80% of Ymax achieved (b, d, f). Considering 80% of Ymax achieved means that the food production units (FPUs) with 80% of Ymax achieved under a given intensification scenario adopt the first scenario of baseline, N_{25} , $N_{25}Irr_{10}$ and $N_{50}Irr_{30}$, which has already achieved 80% of Ymax. Points in each curve are derived by intensifying FPUs one-by-one with an ascending order of N emission intensity of new inputs and then a descending order of N use efficiency of yield gains on the basis of Fig. 3. For better visualization, every 10 points between the first and last points are plotted for each curve. Note: lines for $N_{max}Irr_{max}$ are not always highest because different sorting sequences are used for different scenarios.

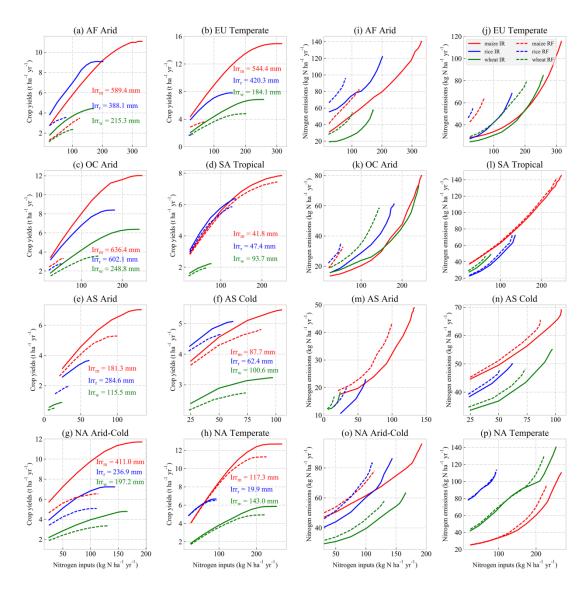


Figure 5. Response curves of crop yields (a-h) and nitrogen (N) emissions (i-p) to N inputs (N_{in}) under irrigated (IR, solid lines) and rainfed (RF, dashed lines) cultivations in different continental-climate regions. Irr_m: irrigation water requirements of maize; Irr_r: irrigation water requirements of rice; Irr_w: irrigation water requirements of wheat. A map of the continental-climate regions can be found in Fig. 1b. AF: Africa, AS: Asia, EU: Europe, NA: North America, OC: Oceania, SA: South America. The response curves are derived from simulations with automatic N fertilization by setting different N_{in} caps: 25, 50, ... 400, 450, 500, and 1000 kg N ha⁻¹ yr⁻¹. Actual N_{in} is lower than N_{in} cap when a high-level cap is used. Rainfed cultivation has lower maximum N_{in} than irrigated cultivation.

Table 1. Description of intensification scenarios in terms of increasing nitrogen (N) inputs and irrigation (Irr) areas.

Scenario	N inputs	Irrigation				
		areas				
baseline	N _{in-base}	A_{ir}				
N_{25}	$N_{in\text{-}base} + 0.25 \times \Delta N_{in\text{-}max}$	A_{ir}				
$N_{25}Irr_{10} \\$	$N_{in\text{-}base} + 0.25 \times \Delta N_{in\text{-}max}$	$A_{ir} + 0.10 \\ \times A_{rf}$				
$N_{50}Irr_{30} \\$	$N_{\text{in-base}} + 0.50 \times \Delta N_{\text{in-max}}$	$A_{ir} + 0.30 {\times} A_{rf}$				
$N_{max} Irr_{max} \\$	$N_{\text{in-max}}$	Full irrigation				

 $N_{\text{in-base}} = \min(N_{\text{in-a}}, N_{\text{in-auto}}); N_{\text{in-a}}: \text{ actual } N \text{ inputs based on the EarthStat dataset; } N_{\text{in-auto}}: \text{ optimal } N \text{ inputs derived from PEPIC simulations considering baseline irrigation land condition based on the MIRCA2000 dataset; <math>N_{\text{in-max}}: \text{ optimal } N \text{ inputs derived from PEPIC simulations considering croplands with full irrigation; } \Delta N_{\text{in-max}} = N_{\text{in-max}} - N_{\text{in-base}}; A_{\text{ir}} \text{ and } A_{\text{rf}}: \text{ baseline areas for irrigation and rainfed cultivations.}}$

Table 2. Global average irrigation water (Irr) [mm], nitrogen inputs (N_{in}) [kg N per ha], crop yields (Y) [t per ha], and nitrogen emissions (N_e) [kg N per ha] under baseline and different intensification scenarios by intensifying all croplands.

	maize				rice				wheat			
	Irr	Nin	Y	Ne	Irr	Nin	Y	Ne	Irr	Nin	Y	Ne
baseline	50.2	93.2	5.5	68.3	41.3	86.8	4.5	74.2	44.7	70.7	2.7	43.4
N_{25}	50.2	111.0	6.0	74.7	41.3	97.5	4.8	79.5	44.7	92.1	3.1	48.5
$N_{25}Irr_{10} \\$	62.0	111.5	6.1	73.9	42.1	97.5	4.8	79.3	55.4	92.3	3.2	48.0
$N_{50}Irr_{30} \\$	86.8	132.8	6.8	81.0	43.9	108.9	5.1	85.4	78.0	114.2	3.6	52.7
$N_{max}Irr_{max} \\$	173.0	193.0	8.2	108.6	50.1	136.0	5.4	102.9	156.7	168.2	4.6	71.1

Table 3. Global average irrigation water (Irr) [mm], nitrogen inputs (N_{in}) [kg N per ha], crop yields (Y) [t per ha], and nitrogen emissions (N_e) [kg N per ha] under different intensification scenarios by avoiding further intensifying croplands with 80% of maximum Y achieved.

	maize				rice				wheat			
	Irr	$N_{\rm in}$	Y	Ne	Irr	$N_{\rm in}$	Y	Ne	Irr	N_{in}	Y	Ne
N_{25}	50.2	110.6	6.0	74.5	41.3	94.4	4.7	77.9	44.7	91.7	3.1	48.3
$N_{25}Irr_{10} \\$	61.5	111.1	6.1	73.9	41.8	94.4	4.7	77.9	54.9	91.8	3.2	48.0
$N_{50}Irr_{30} \\$	79.7	129.0	6.7	80.0	42.7	100.2	4.9	81.1	75.2	112.0	3.6	52.7
$N_{\text{max}} Irr_{\text{max}}$	117.6	150.7	7.3	88.4	44.8	106.8	5.0	84.9	131.0	146.5	4.3	61.5