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Fe(II)-catalyzed ligand-controlled dissolution of iron(hydr)oxides

Jagannath Biswakarma[1,2], Kyounglim Kang[3], Susan C. Borowski[1#], Walter D.C. Schenkeveld[3†],

Stephan M. Kraemer[3], Janet G. Hering[1,2,4] and Stephan J. Hug[1]*

[1] Eawag, Swiss Federal Institute of Aquatic Science and Technology, W&T, CH-8600 Dübendorf, Switzerland

[2] ETH, Swiss Federal Institute of Technology Zurich, IBP, CH-8092 Zürich, Switzerland

[3] University of Vienna, Dept. of Environmental Geosciences, 1090 Vienna, Austria

[4] EPFL, Swiss Federal Institute of Technology Lausanne, ENAC, CH-1015 Lausanne, Switzerland

* Corresponding author. E-mail: stephan.hug@eawag.ch

current address: Randolph-Macon College, Department of Chemistry, 114 College Avenue, Ashland VA. 23005-5505, USA

† Current address: Copernicus Institute, Faculty of Geosciences, Utrecht University Princetonlaan 8A, 3584 CB Utrecht, the Netherlands

Abstract

Dissolution of iron(III)phases is a key process in soils, surface waters and the ocean. Previous studies found that traces of Fe(II) can greatly increase ligand controlled dissolution rates at acidic pH, but the extent that this also occurs at circumneutral pH and what mechanisms are involved are not known. We addressed these questions with infrared spectroscopy and ^{57}Fe isotope exchange experiments with lepidocrocite (Lp) and 50 µM ethylenediaminetetraacetate (EDTA) at pH 6 and 7. Addition of 0.2-10 µM Fe(II) led to an acceleration of the dissolution rates by factors of 7-31. Similar effects were observed after irradiation with 365 nm UV light. The catalytic effect persisted under anoxic conditions, but ceased as soon as air or phenanthroline was introduced. Isotope exchange experiments showed that added ^{57}Fe remained in solution, or quickly re-appeared in solution when EDTA was added after ^{57}Fe(II), suggesting that catalyzed dissolution occurred at or near the site of ^{57}Fe incorporation at the mineral surface. Infrared spectra indicated no change in the bulk, but changes in the spectra of adsorbed EDTA after addition of Fe(II) were observed. A kinetic model shows that the catalytic effect can be explained by electron transfer to surface Fe(III) sites and rapid detachment of Fe(III)EDTA due to the weaker bonds to reduced sites. We conclude that the catalytic effect of Fe(II) on dissolution of Fe(III)(hydr)oxides is likely important under circumneutral anoxic conditions and in sunlit environments.
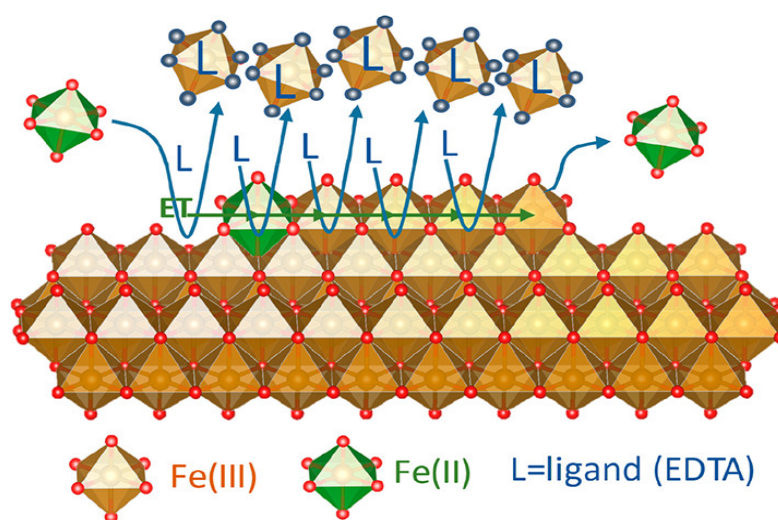
Introduction

The bioavailability of iron in oxic environments with circumneutral pH is limited due to the low solubility and slow dissolution kinetics of iron (oxyhydr)oxide minerals[1] (here collectively termed Fe(III)(hydr)oxides). A common strategy of plants and micro-organisms for Fe acquisition involves the release of ligands (L) that promote the dissolution of Fe(III)(hydr)oxides. Laboratory and field studies have demonstrated that iron availability and rates of Fe(III)(hydr)oxide dissolution are increased by natural ligands (e.g. siderophores such as desferroxamine-B (DFOB) and aerobactin) [2] and by synthetic ligands (e.g. ethylenediaminetetraacetate (EDTA).[3] It has been further found that dissolution rates can be accelerated by synergistic effects between different ligands, for example between oxalate and DFOB [4, 5] and other low molecular weight organic acids and siderophores. [6, 7] Similar synergistic effects were observed in the dissolution of Fe(III)(hydr)oxides associated with release of trace elements. [8] [9]

While the synergistic effects between ligands were reported to accelerate dissolution rates by factors of 2-10, [4] a group of earlier studies found that a trace of Fe(II) in presence of EDTA accelerated the rate of ligand-controlled dissolution by factors of 10-100 at low pH under anoxic conditions.[10-16] There were no follow-up studies on the catalytic effect of Fe(II) to our knowledge until recently, and it is not known if and under what conditions the strong catalytic effect is also efficient at circumneutral pH. In a recent study, Wang et al. observed a synergistic effect between DFOB or N,N′-di(2-hydroxybenzyl)-ethylenediamine-N,N′-diacetic acid (HBED) and the reducing ligand ascorbate in the dissolution of goethite at pH 6. [17] Similar synergistic effects were observed between ascorbate and DFOB, 2′-deoxymugineic acid, citrate and esculetin. [18] The Fe(II) concentrations resulting from ascorbate addition were not determined, but a careful analysis of the results showed that Fe(II) acted as a catalyst for accelerated dissolution. In a subsequent study using density function theory, Kubicki et al. [19] found that adsorption of Fe(II) near an Fe(III) corner site with adsorbed oxalate lead to a weakening of bonds between the corner site and the lattice, explaining a faster rate of detachment of this site from the lattice. A synergistic effect between oxalate and ascorbate was also recently reported in the extraction of arsenic from contaminated soils. [20]

The catalytic effect of Fe(II) in the earlier dissolution studies conducted at low pH has been explained as follows: Fe(II), either added or generated through reduction of surface Fe(III) by a reductant, can form ternary surface complexes, e.g. ≡Fe(III)-L-Fe(II) (bonds to the lattice are symbolized by ≡). Electron transfer (ET) then leads to formation of ≡Fe(II)-L-Fe(III) and rapid detachment of Fe(III)L into solution due to the relatively weaker Fe(II)-L bonds compared to Fe(III)-L bonds. Detachment of Fe(II) then leads to formation of new ternary surface complexes and to accelerated dissolution in a catalytic cycle.

Several more recent studies found that Fe(II) adsorption on Fe(III)(hydr)oxides, in the absence of ligands, can lead to isotopic exchange and recrystallization of the mineral [21-28] and suggested a mechanism for the recrystallization that might also provide alternative explanations for Fe(II) catalyzed dissolution. The recrystallization was explained by a model in which electrons transferred from adsorbed Fe(II) to the surface are mobile within the Fe(III)(hydr)oxide bulk solid. This leads to formation of Fe(III) and growth at the site of initial Fe(II) adsorption and to formation and detachment of Fe(II) and thus dissolution at a remote surface site. [21] Based on the studies on Fe(II)-catalyzed recrystallization and on the observations and suggestions by Wang et al. and Schenkeveld et al. [17, 18], delocalization of charge in the mineral structure could play an important role in the accelerated Fe(II)-catalyzed dissolution of Fe(III)(hydr)oxides.

In a new approach to study dissolution of Fe(III)(hydr)oxides, we applied Attenuated Total Reflectance Fourier-transformed Infrared (ATR-FTIR) spectroscopy to examine the dissolution of lepidocrocite (Lp) and to observe structural changes in the bulk and on the surface before and during the dissolution process continuously with high time resolution (no filtration is required). We used ATR-FTIR previously to study the structure and the photoreactions of surface complexes of citrate,[29] DFOB and aerobactin [30] and of dicarboxylates, [31] but not to follow changes in Lp structure and its dissolution. In studies by Borer et al.,[32-35] we investigated light induced photo-reductive dissolution of Lp in the absence and presence of DFOB and aerobactin, considering Fe(II) as a reaction product but not its potential autocatalytic role. [32-35]. In the current study, we address the catalytic effect of Fe(II) in the non-reductive ligand-controlled dissolution in the dark, with added and photo-generated Fe(II). Isotope ex-

104 periments with ^{57}Fe(II) were used to probe the fate of added Fe(II) and the importance of charge migra-

105 tion and isotopic exchange. Further, we applied kinetic modeling to test if proposed mechanisms are

106 able to describe experimental results. Lp was chosen due its relevance in the environment and EDTA

107 as a model ligand for non-reducing siderophores with predominantly carboxylate groups and due to its

108 applications in fertilizers to increase the solubility of iron in soils. Fe(II) catalyzed dissolution might be

109 an important pathway in the dissolution of iron(III) phases in sub- and anoxic environments, e.g. in soils

110 and sediments, and in the oxic-anoxic interface in shallow lakes, wetlands and irrigated fields. Photo-

111 produced Fe(II) might accelerate dissolution of iron(III) phases in sunlit oxic environments, e.g. in the

112 ocean and in the atmosphere. Studies addressing these questions at circumneutral pH (which have not

113 been conducted so far) are of particular interest because the solubility of Fe(III) phases is minimal in

114 this pH-range. In a parallel study, the effect of added Fe(II) is investigated with Lp, goethite, hematite

115 and 2-line ferrihydrite and the ligands HBED and DFOB over a larger pH range. [36]

116

Materials and Methods

Chemicals and Solutions

119 All chemicals used were of analytical grade and are listed in Table S1. Aqueous solutions were prepared

120 using high-purity doubly-deinonized (DDI) water (Barnstead Nanopure). To perform studies on isotope

121 exchange and Lp dissolution, a 20 mM ^{57}Fe(II) stock solution was prepared by dissolving 5.7 mg ^{57}Fe

122 (Sigma-Aldrich, 95% pure) in 100 µl 2.5 M HCl and dilution to 5.00 ml with DDI H$_2$O. The ^{57}Fe powder

123 was 99% pure in Fe, with an isotopic composition of 95.06 % ^{57}Fe, 3.04% ^{56}Fe and 1.86% ^{58}Fe (Certif-

124 icate of Analysis of the supplier).

Lp (γ-FeOOH) synthesis and characterization

126 The synthesis of Lp was modified from Schwertmann [37] and is decribed in more detail in the Supporting

127 Information (SI). Briefly, 60 mM FeCl$_2$ was purged with a stream of air and titrated with 1M NaOH to

128 within a pH range of 6.65 - 6.76 in a strongly stirred solution, until NaOH consumption ceased. After 2

129 h, the bright orange suspended particles were collected by centrifugation and repeatedly washed by re-

130 suspension and centrifugation with DDI water. Finally, the solid was collected and dried with stream of

131   N_2. The purity and the properties of the synthesized Lp were determined by X-ray diffraction (XRD),

132   ATR-FTIR, and scanning electron microcopy (SEM), see SI, Figures S1-S3. The specific surface area

133   of Lp was measured as 63 m^2/g by a multipoint N_2 Brunauer–Emmett–Teller (BET) adsorption method

134   (Quantachrome Nova 3200).

ATR-FTIR measurements

136   ATR-FTIR spectroscopy was employed to investigate the adsorption of EDTA onto Lp and to follow

137   the dissolution of Lp at pH 6 and 7. Measurements were performed on a Biorad FTS 575C instrument

138   equipped with a liquid N_2-cooled mercury cadmium telluride (MCT) detector and a nine reflection dia-

139   mond ATR unit (SensIR Technologies, Danbury, CT), as described in detail in the SI. Briefly, a thin

140   layer of 40-60 µg Lp was deposited on the ATR crystal (diamond, ϕ 4 mm). IR absorbance spectra of

141   the layer were recorded after drying with a gentle stream of N_2. Subsequently, the layer was covered

142   with 40 ml 9.5 mM NaCl (38 ml 10 mM NaCl) and 5 mM MES/MOPS (2 ml 100 mM MES/MOPS

143   stock solution). Continuous purging of the aqueous solution with high purity N_2 gas led to removal of

144   dissolved O_2 and desorption of adsorbed CO_2 and served to stir the solution during measurements. Re-

145   sidual O_2 concentrations after 3 h were < 10 nM (measured with a photo-luminescent probe directly in

146   the reaction cell, see SI). At this point, a single-beam background spectrum with the Lp-layer in contact

147   with the anoxic background electrolyte was recorded. Subsequently, difference absorbance (ΔA) spectra

148   were measured continuously, every 43 s or 71 s.

149   **EDTA and Fe(II) addition.** 50 µM EDTA (400 µl 5 mM EDTA deoxygenated stock solution) was

150   added 180-200 s after the background spectrum was measured. Fe(II) (0.2-10 µM) was added 1800 s

151   after EDTA addition. Toward the end of the measurement after 6300 s, synthetic air or 1mM phenan-

152   throline was added to examine the effect of O_2 or phenanthroline on dissolution. The same experiments

153   were performed with addition of Fe(II) 1800 s before addition of EDTA. Experiments at each of the

154   seven different Fe(II) concentrations were performed in duplicate. All procedures except the illumina-

155   tion with UV-light were performed under weak yellow light (>550 nm), to avoid photochemical reac-

156   tions with solid and dissolved Fe(III) species.

157   **Photochemically produced Fe(II).** A UV (365 nm, bandwidth 10 nm) light-emitting diode (Dr. Groe-

158   bel UVElektronik GmbH) was used to produce Fe(II) at the Lp surface under both anoxic (N_2) and oxic

159 conditions (synthetic, CO_2-free air). The Lp-layer immersed in the aqueous solution, was illuminated at

160 a distance of 7.25 cm from the lamp through a UV-transparent glass plate covering the solution. Illumi-

161 nation was done twice for 900 s: (a) 2700-3600 s and (b) 6200-7100 s. The light intensity of the irradi-

162 ation at the Lp-layer was 150 W/m^2 (approximately double the irradiance of natural sunlight in the range

163 of 300-450 nm) as measured by ferrioxalate actinometry (see SI).

164 **^{57}Fe(II) isotope exchange and Lp dissolution with ICP-MS analysis.**

165 ^{57}Fe(II) was used to investigate the fate of the added Fe(II) during the dissolution process. Batch disso-

166 lution experiments were conducted under anoxic conditions at pH 6 at room temperature. Lp suspensions

167 (1125 μM ≈ 0.100 g L^{-1}) were prepared by dispersing 10.0 mg Lp in 100 ml aqueous solution (9.5 mM

168 NaCl and 5mM MES). Suspensions were purged with N_2 for at least 4 h before addition of EDTA or

169 ^{57}Fe(II). Fe(II) was either applied after or before EDTA: (a) 1.2 μM ^{57}Fe(II) was injected 30 min after

170 50 μM EDTA addition, and (b) 1.0 or1.2 μM ^{57}Fe(II) was added 30 min before EDTA addition. The

171 time between Fe^{2+} and EDTA addition made it possible to a) examine Lp dissolution in the presence of

172 EDTA and b) to investigate isotope exchange in the absence of EDTA. Aliquots of 1.5 ml were with-

173 drawn periodically, filtered through a 0.1 μm nylon filters (Whatman® Puradisc 13 syringe filters) and

174 diluted in 1% HNO_3 (Merck, suprapure) for inductively coupled plasma-mass spectrometry (ICP-MS,

175 Agilent 7500ce) analysis. The dissolution of Lp was followed at mass 56, with counts calibrated to

176 represent the total concentration of all iron isotopes from Lp. Dissolution rates were determined as the

177 slope of the linear fit of the dissolved Fe concentration as a function of time over the time-interval 600-

178 2400s after Fe(II) addition. We use the following definitions:

179

180 $[^{56}Fe]^*_{diss}$ Concentration of dissolved ^{56}Fe multiplied with 1.092 to represent the total dissolved

181   iron (sum of all iron isotopes) released from Lp. (5.85% ^{54}Fe, 91.57% ^{56}Fe, 2.12% ^{57}Fe,

182   0.28% ^{58}Fe). This is the concentration of Fe as usually measured with ICP-MS at mass

183   56, with counts calibrated with certified standard solutions. ^{56}Fe comes from dissolution

184   of Lp or from exchange of added ^{57}Fe with ^{56}Fe from Lp. We corrected for the 3% ^{56}Fe

185   impurity in 1-1.2 μM added ^{57}Fe. This correction is small (<0.04 μM) compared to the

186   concentrations of ^{56}Fe released from Lp by isotope exchange and dissolution.

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187	$[^{57}Fe]_{diss}$	Concentration of dissolved ^{57}Fe. Measured with ICP-MS at mass 57 and calibrated with
188		normal certified Fe standards, considering that they contain 2.12% ^{57}Fe.
189	$[^{57}Fe]_{tracer, diss}$	To follow the concentration of ^{57}Fe added as a tracer, we subtracted the ^{57}Fe coming
190		from dissolution or isotope exchange with Lp: $[^{57}Fe]_{tracer, diss} = [^{57}Fe]_{diss} -$
191		$0.0212[^{56}Fe]^*_{diss}$.
192		(The subtracted concentration of ^{57}Fe from dissolution of Lp is 1.06 µM, when 50 µM
193		Fe(III) is dissolved in the presence of 50 µM EDTA).

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Results and Discussion

Lp dissolution monitored with ATR-FTIR

Fig 1A shows the ATR-FTIR absorbance spectrum of a Lp layer and a SEM image of the Lp in the inset. The Lp consists of plate-like (50-100 nm width and 200-500 nm length) aggregates of 200-500 nm long rods of 10 nm width and less than 10 nm thickness. The IR spectrum shows the characteristic vibrations of Lp at 1160 cm^{-1}, 1021 cm^{-1}. The full spectrum shown in Figure S1 also shows the characteristic band of rod-shaped Lp at 752 cm^{-1} (in contrast to plate-shaped Lp where this band is located at 742cm^{-1}).[38] Figure 1B shows ATR-FTIR difference absorbance (ΔA) spectra recorded during EDTA adsorption over 1800 seconds at pH 6 under anoxic condition. Addition of 50 µM EDTA at t=180-220 s lead to characteristic IR absorbance bands in the range of 1700-1200 cm^{-1} that increased in the first few minutes and then reached a stable level. There are two distinct peaks of adsorbed EDTA at 1570 cm^{-1} and 1408 cm^{-1}, which were assigned to the asymmetric and symmetric vibrations of carboxylate groups, respectively. [39] There was an increase, but no significant change in the shape of the spectra during EDTA adsorption. This shows that at pH 6.0 one single surface complex, or several complexes in constant proportions, were formed on the Lp layer. Complexes with different spectral characteristics are formed at lower pH (SI, Figure S4). The adsorption of EDTA led to a small decrease to the left of and a small increase to the right of the absorbance peak of Lp. These small changes in absorbance must be due to local changes in the coordination environment of the Lp lattice at the surface related to adsorption of EDTA, but are not interpreted further. Importantly, there was only a minimal decrease at 1021 cm^{-1} for

the entire duration of the first 1800 s, indicating that there was only minor dissolution of Lp in the presence of EDTA alone.

The group of spectra in (C) show a decrease in absorbance at 1021 cm^{-1}, indicating a strong acceleration in dissolution of Lp after addition of 1 µM Fe(II). Minor spectral changes (which will be discussed later) were recorded in the spectra of adsorbed EDTA upon addition of 1 µM Fe(II). Introduction of synthetic air or 1 mM phenanthroline at 6300 s stopped the accelerated dissolution of Lp. The effect of air indicates a rapid oxidation of Fe(II) at the surface, which ends the catalytic effect of Fe(II) on dissolution. Phenanthroline addition caused rapid formation of Fe(II)-phenanthroline complexes in the solution and apparently also fast desorption of Fe(II) from the surface or complexation of Fe(II) at the surface, which ended the catalytic action of Fe(II). These results show the importance of the presence of Fe(II) or Fe(II)EDTA for the catalytic effect on the dissolution processes. Similar observations were made when 10 µM Fe(II) was added 3000 s after 50 µM EDTA at pH 7 (see SI, Fig S6).

Fig. 2 shows the kinetics of EDTA adsorption and Lp dissolution at pH 6 under anoxic condition (data points every 43 s). EDTA adsorption was examined at 1408 cm^{-1} and Lp dissolution at 1021 cm^{-1}. EDTA adsorbed rapidly and reached equilibrium within 600 s after EDTA addition. There is virtually no decrease of the absorbance at 1021cm^{-1}, indicating a low rate of dissolution in the presence of only EDTA. Increasing concentrations of Fe(II) (0.2-10 µM) were added 1800 s after 50 µM EDTA addition. A clear catalytic effect was observed already for sub-micro molar concentrations of added Fe(II) with a strongly accelerated decrease of absorbance at 1021 cm^{-1}. The same effect at sub-micromolar concentrations were observed in a parallel study with two different ligands (HBED and DFOB) and several Fe(III)(hydr)oxides [36]. The plots in Fig. 2 show that the dissolution rates started to increase quickly after addition of Fe(II) and that the decrease in absorbance after Fe(II) addition was linear over time from 3000-6000 s. At the same time, the absorbance of EDTA at 1408 cm^{-1} showed no visible change for 0.1-1.0 µM added Fe(II). With 6-10 µM added Fe(II), the absorbance decreased quickly and then remained at a stable level. This indicates that a steady-state dissolution rate was reached within 30-120 s, which remained constant throughout our ATR-FTIR experiments. The fast decrease in absorbance at 1408 cm^{-1} and the changes in the spectra of adsorbed EDTA for 6 µM and 10 µM Fe(II) will be discussed in the last section. The order in which EDTA and Fe(II) were added did not alter the main observations. When

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242   1 µM Fe(II) was added 1800 s before 50 µM EDTA, no changes were recorded for Lp at 1021 cm^{-1} (see

243   SI, Fig. S4). After EDTA was added, approximately the same rate of decline in absorbance (-0.033 units

244   after 6000 s) was reached as when EDTA was added first (-0.032 units after 6000s) . 1 mM phenanthro-

245   line addition after ~ 8800 s again stopped the accelerated Lp dissolution.

246   To determine dissolution rates, slopes were calculated from the linear ranges of the decreasing

247   absorbances at 1021 cm^{-1} in Figure 2. (The ranges were 900-1200 s for EDTA only and 3000-5000 s in

248   the presence of Fe(II)). The rates of Lp dissolution expressed in % of Lp dissolved per time were cal-

249   culated from the slopes divided by the initial absorbance of Lp at 1021 cm^{-1} for each experiment. The

250   initial absorbances of the Lp layers in the different experiments varied in the range of 0.7-0.9 and are

251   listed in SI, Table S2. Figure 3 presents the rate of Lp dissolution as a function of added Fe(II) concen-

252   trations. In the range of 0.2-2 µM,Fe(II), the rate of Lp dissolution increased close to linearly with in-

253   creasing concentrations of added Fe(II), after which the increase leveled off. The solid line corresponds

254   to a fit with a kinetic model which is described in a later section.

255

256   **Photochemically produced Fe(II).** The kinetics of EDTA adsorption and Lp dissolution, during ex-

257   periments in which Fe(II) was photo-chemically generated at pH 6.0, are shown in Fig. 4.. EDTA (50

258   µM) was added after 180-220 s and EDTA adsorption reached an equilibrium within 600 s under both

259   anoxic and oxic conditions. Again, there was only very minimal Lp dissolution in the presence of EDTA

260   alone. Under UV-illumination (for 900 s) absorbances at 1408 and 1021 cm^{-1} immediately started to

261   change. The absorbance at 1408 cm^{-1} (adsorbed EDTA) decreased and stabilized at a lower level during

262   illumination, while the absorbance at 1021 cm^{-1} (Lp) decreased linearly under both anoxic and oxic

263   conditions. We attribute this mainly to photolysis of Fe(III)-EDTA surface complexes and formation of

264   Fe(II) directly on the Lp surface, due to the high photoreactivity of the adsorbed EDTA, [16] but intrinsic

265   photo-processes of Lp with formation of Fe(II) and OH-radicals could also have contributed to EDTA

266   degradation.[34] From the decrease in the absorbance of Lp, we estimate that 0.14-0.25 µM Fe(II) were

267   formed during each illumination (see SI). After the illumination stopped, photolyzed EDTA was re-

268   placed by EDTA from solution which can be seen by the recovery of the absorbance at 1408 cm^{-1}. Also,

269   Lp dissolution under oxic condition almost instantly came to a halt after illumination. In contrast, under

anoxic conditions, Lp dissolution continued in the dark, yet at a slower rate. During a second UV-illumination (from 6200-7100 s) under anoxic conditions, more Lp was dissolved than during the first illumination and the subsequent dissolution rate in the dark continued at a faster rate. Apparently, Fe(II) produced during the second illumination was added to Fe(II) formed in the first illumination and thus increased the concentration of photogenerated Fe(II) and the catalytic effect of Fe(II) during and after the illumination. The catalytic effect was not observed under oxic conditions, where the second illumination lead to the same rate and amount of dissolution as the first illumination and to essentially no dissolution in the absence of light. We ascribe this to fast oxidation of the formed adsorbed Fe(II) by dissolved O_2. Similar observations were made in experiments at pH 7 (see SI, Fig S7). These results demonstrate a continuing catalytic effect of photochemically produced Fe(II) on non-reductive ligand-controlled dissolution at pH 6 and 7 after illumination stops under anoxic conditions. Qualitatively, the dissolution rates after illumination are similar to those with 0.2-0.5 µM added Fe(II) immediately after the illumination stops, and about half of these rates a few minutes later. Since Fe(II) is formed directly at the surface and equilibration with the solution might take a few minutes, the observed Fe(II) catalyzed dissolution rates agree well with the effect of the estimated 0.14-0.25 µM Fe(II) in the solution volume of 40 ml.

^{57}Fe(II) isotope exchange and Lp (lepidocrocite) dissolution. To determine the fate of Fe added as Fe(II) before and during Lp dissolution, we performed dissolution experiments in which we added Fe(II) as ^{57}Fe(II).

Figure 5A shows Lp dissolution as a function of time before and after addition of 1.2 µM ^{57}Fe(II) to a Lp suspension (1125 µM) containing 50 µM EDTA at pH 6.0 under anoxic conditions. The dissolved concentrations of iron coming from dissolution of Lp, $[^{56}Fe]^*_{diss}$, are plotted as purple triangles on the right y-axis. In agreement with the FTIR results, the dissolution of Lp was very slow in the absence of Fe(II). After addition of ^{57}Fe(II), the dissolution was strongly accelerated and $[^{56}Fe]^*_{diss}$ reached 42 µM after 1600s, approaching the expected limit of Fe(III) solubility with 50 µM EDTA. The dissolved concentrations of added ^{57}Fe are plotted as orange squares on the left y-axis ($[^{57}Fe]_{tracer, diss}$). Importantly, added ^{57}Fe remained in the solution in the presence of EDTA. This shows that the added ^{57}Fe (added as ^{57}Fe(II)) was not incorporated into the surface or bulk during dissolution. A decrease in

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298     $[^{57}Fe]_{tracer, diss}$ would have been observed if adsorption of $^{57}Fe(II)$ and incorporation as $^{57}Fe(III)$ by elec-

299   tron transfer would have led to preferential dissolution of surface sits or surfaces which are different

300   from the sites of adsorption and incorporation.

301   In the presence of EDTA, rapid electron transfer at the Lp-surface could have resulted in oxidation

302   of added $^{57}Fe(II)$ to $^{57}Fe(III)$. We cannot comment on the oxidation state of $^{57}Fe_{tracer,diss}$ in the presence

303   of EDTA; addition of phenanthroline (phen) to filtered solutions resulted in only very low and incon-

304   sistent UV absorbance, which did not allow quantification of $Fe(II)(phen)_3$. Even so, it is clear that ^{57}Fe

305   added in the presence of EDTA was not incorporated into the solid.

306   Figure 5B shows ^{57}Fe isotope exchange and Lp dissolution with reversed addition of $^{57}Fe(II)$

307   and EDTA. In the following, we first discuss the observations before addition of EDTA, when the oxi-

308   dation state of measurable dissolved Fe can only be Fe(II), followed by the observations after EDTA

309   addition, where the oxidation state of dissolved Fe is mixed, but predominantly Fe(III).

310   In contrast to addition of $^{57}Fe(II)$ in the presence of EDTA, we observed an immediate decrease

311   in the concentration of the added 1.0-1.2 µM $^{57}Fe(II)$ (orange squares, left y-axis) in the absence of

312   EDTA, presumably by adsorption. This was followed by a slower decrease in $[^{57}Fe(II)]$, coupled with

313   an increase of $[^{56}Fe(II)]^*_{diss}$ (red triangles, left y-axis), indicating isotope exchange between Fe(II) in

314   solution and Fe from Lp over the 1800 s before EDTA was added. The sum of these concentrations

315   (green circles on the left y-axis) was 1.0-1.1 µM, meaning that 0.1-0.2 µM of the added Fe(II) remain

316   adsorbed. Since $[^{56}Fe]^*_{diss}$ increased from 0 to 0.70 µM, up to 0.7 µM $^{57}Fe(II)$ must have been incorpo-

317   rated into the surface as $^{57}Fe(III)$. Thus, before EDTA addition, both adsorption and isotope exchange,

318   but no dissolution was observed. This is in agreement with studies showing isotope exchange between

319   dissolved Fe(II) and FeOOH that leads to recrystallization on longer time scales [21-28].

320   After addition of EDTA, a rapid increase of $[^{56}Fe]^*_{diss}$ (purple triangles, right axis) due to cata-

321   lyzed Lp dissolution was observed. At the same time, the added (tracer) ^{57}Fe that was previously ad-

322   sorbed and/or exchanged was released back into solution. Over 80% of the adsorbed/exchanged ^{57}Fe

323   was in solution after dissolution of 10 µM Lp within the first 1000 s after addition of EDTA.

324   This observation shows that added ^{57}Fe was not preferentially incorporated into surface sites or surfaces

325   that were less prone to dissolution upon addition of EDTA. Although this observation does not exclude

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326 ET to other sites and other surfaces, such ET (if it occurs) does not result in the preferential dissolution

327 of those other sites or surfaces.

328   We can, in the main, exclude preferential dissolution distant from the site at which added ^{57}Fe was

329 adsorbed, since addition of EDTA resulted in release of >80% of the added tracer ^{57}Fe back into solu-

330 tion. The initially adsorbed ^{57}Fe (i.e., 1 µM of the added 1.2 µM ^{57}Fe) corresponds to only 5.7 % of the

331 total active surface site concentration of 17.5 µM (calculated for 100 mg Lp/L with a surface area of

332 $6.3 \cdot 10^{18}$ nm^2/L assuming 1.67 active site per nm^2.[40] If ET had resulted in preferential dissolution at

333 different surface sites and surfaces (with only the natural abundance of ^{57}Fe) then no increase in

334 $[^{57}Fe]_{tracer, diss}$ should have been observed (recall that the reported $[^{57}Fe]_{tracer, diss}$ was corrected for the

335 natural abundance of ^{57}Fe in Lp) . Even if dissolution had occurred randomly at all surface sites (enriched

336 to 5.7 % ^{57}Fe), dissolution of 10 µM Lp would have resulted in a maximum tracer concentration

337 $[^{57}Fe]_{tracer, diss}$ of 0.57 µM. Incorporation of a minor faction of ^{57}Fe (added before EDTA) into more

338 stable sites and surfaces is, however, consistent with the association of 10-20% of the added ^{57}Fe with

339 the solid even after dissolution of 50 µM Lp (i.e., measured as $[^{56}Fe]^*_{diss}$).

340   Our observations do not exclude charge injection and charge migration in the Lp surface and bulk

341 lattice that lead to recrystallization on a longer time scale. In the presence of a ligand that promotes

342 dissolution, mobile charge can be transferred back to previously adsorbed or incorporated ^{57}Fe and ac-

343 celerate its dissolution. An alternative explanation for the fast dissolution of the added ^{57}Fe is the for-

344 mation of a less crystalline or amorphous ^{57}Fe(III) phases that are dissolved more quickly than Lp after

345 addition of EDTA. However, a surface precipitate with addition of only 1-1.2 µM Fe(II) to a Lp suspen-

346 sion with estimated 17.5 µM surface sites seems less likely than adsorption and isotope exchange by

347 incorporation of ^{57}Fe into the surface.

348 **Possible phase changes.** We observed no evidence for transformation of Lp to goethite or magnetite

349 with 0.2-10 µM added Fe(II) (to which FTIR is very sensitive, see Figure S3) in our experiments. Pre-

350 vious studies found exchange of ^{55}Fe between Lp and dissolved Fe(II) (0.2-1.0 mM) Fe(II) at pH 6.5, [41,]

351 [42] but only minor formation of magnetite even with 1mM dissolved Fe(II) after 2 days [41] and no for-

352 mation of goethite over 7 days. [42]

Kinetic model for the catalytic effect of Fe(II). In our experiments, ligand controlled dissolution was slow compared to Fe(II)-catalyzed dissolution. We define the measured catalytic effect (CE) as $R_{diss,L,Fe(II)}$ / $R_{diss,L}$, where $R_{diss,L,Fe(II)}$ and $R_{diss,L}$ are the dissolution rates in the presence and absence of Fe(II), respectively. As defined, CE was 1 without added Fe(II) and ranged from 7-31 with 0.2-10 µM added Fe(II), as listed in Table S3.

To explore reaction mechanisms that can explain the catalytic effect of Fe(II), a kinetic model was developed that allowed us to fit the model outputs to measured data by optimizing rate coefficients and concentrations of active surface sites in the model. Tentative reaction mechanisms for Fe(II) cata-lyzed dissolution have been presented before for dissolution at low pH, [10, 43, 44] but it has not been shown that they can quantitatively explain experimental results. The model with reaction equations is presented in Table 1 and is described in more detail in the SI.

We assumed that all equilibrium reactions are fast and not rate determining for dissolution on the inves-tigated time scale. As consequence of the fast equilibration in reactions R1-R5, the order of the addition of EDTA and Fe(II) should not change the observation of the catalytic effect, as was experimentally observed (Figure S4). The competitive adsorption of formed Fe(III)EDTA (R6) can explain the non-linear dissolution in the batch-experiments, with dissolution rates that decrease more strongly from 1800-5000 s (Figure 5) than expected from the decreasing concentration of uncomplexed EDTA. In the dissolution measured with FTIR (Figure 2) this effect was not observed and dissolution proceeded line-arly, which can be explained by the much lower Lp concentrations in the AFR-FTIR experiments. With maximally 8% dissolution of 60 µg Lp in 40 ml, dissolved Fe(III)EDTA concentration remained below 1.4 µM. When 20 µMFe(III) was added (Figure S8) the dissolution rate also decreased. Similar non-linear dissolution kinetics were observed previously in the photo-reductive dissolution of Lp with aero-bactin, [33] where decreasing rates in the presence of added and formed Fe(III) were also ascribed to competing adsorption of formed Fe(III)-ligand complexes with the free ligand. By variation of rate co-efficients (see SI), we found that either the electron transfer (ET) in R7 or the detachment of Fe(III)L can be defined as the rate determining step in this model. Since Fe(II)EDTA is strongly reducing, we assumed that ET is the faster step and adjusted the rate coefficient for detachment of Fe(III)L (k_8) in

380 reaction R8 as the rate determining step. The listed fitted rate coefficients lead to good fits of the cur-

381 rently available data, but they are tentative because the rate coefficients for reactions R4 and R5 corre-

382 spond to reactions (i.e., adsorption of EDTA on adsorbed Fe(II) or adsorption of Fe(II)EDTA) that yield

383 the same surface complex in our model. Since more than 99.8% of Fe(II) is complexed by EDTA, the

384 rate coefficients for these reactions are not independent, as explained in more detail in the SI.

385   Important parameters in the model are the concentrations of active sites for the dissolution re-

386 actions (p_1 to p_4). The site concentrations for the adsorption of Fe(II), EDTA and Fe(III)EDTA were

387 obtained from adsorption data as shown in Figure S11. They agree with ranges for iron(hydr)oxides

388 reported in the literature. [45, 46] A lower concentration of sites that are active for dissolution has been used

389 in models describing the dissolution of hematite [47] and agrees with the understanding that dissolution

390 proceeds on kink and step sites [48] whose concentrations are smaller than the sum of adsorption sites

391 measured in adsorption experiments.

392 **Structure of surface complexes.** In the kinetic model, we formulate the surface complexes that involve

393 both Fe(II) and EDTA as $\equiv Fe^{III}\text{-}Fe^{II}\text{-}L$, but they could alternatively be formulated as $\equiv Fe^{III}\text{-}L\text{-}Fe^{II}$ or as

394 complexes with Fe(II) adsorbed next to a site were EDTA is adsorbed. Ideally, the infrared spectra could

395 provide information about the structure of the complex involving both Fe(II) and EDTA. Addition of

396 Fe(II) caused shifts in the spectra of adsorbed EDTA to higher energy and induced spectral changes

397 around several peak positions (see SI, Fig. S9, SI Table S4). This clearly indicates that the interaction

398 of Fe(II) with adsorbed EDTA led to formation of additional complexes at the surface, or that additional

399 complexes were formed by adsorption of Fe(II)EDTA from solution. However, the shifts could not be

400 interpreted in terms of structures. In a recently published study on the formation of oxalate complexes

401 on Lp, we found that several surface complexes with different structures were formed. For oxalate,

402 detailed DFT calculations allowed us to draw conclusions about the structures of these different surface

403 complexes.[49] DFT calculations with the larger ligand EDTA in the presence of Fe(II) are computation-

404 ally more demanding and were not attempted in this study.

405 **Reaction mechanisms and possible implications.** Fe(II) accelerates the dissolution of Lp by EDTA

406 when added before or after addition of EDTA. At pH 6, most of the added Fe(II) (>80%) remains in

407 solution even in the absence of EDTA. In the presence of 50 μM EDTA, over >99.8% of dissolved

408   Fe(II) and >99.9% of Fe(III) are complexed with EDTA (Table S5). Most likely, the Fe(II) catalyzed

409   dissolution is driven by adsorption of Fe(II)EDTA and formation of mixed Fe(II)-Fe(III) ternary surface

410   complexes (reaction R5 in Table 1), followed by electron transfer and detachment of Fe(III)EDTA (R7-

411   R8). We can currently not determine the fraction of adsorbed Fe(II)EDTA and whether the newly formed

412   Fe(II) (R9) is complexed with EDTA directly at the surface (reaction R4) or leaves the surface (back

413   reaction of R3) and reabsorbs as Fe(II)EDTA (R5) to continue the catalytic cycle. The model constitutes

414   a useful list of possible reactions that can explain our observation, but more work is required to quanti-

415   tatively determine the contribution of each of these reactions and to determine the corresponding rate

416   coefficients.

417

418   **Environmental significance**

419   We observed that added and photo-produced Fe(II) can accelerate the dissolution of Lp with EDTA by

420   a factor of 7-30 at pH 6 and at pH 7 under anoxic conditions. Fe(II) catalyzed dissolution could be an

421   efficient strategy for accelerated mobilization and uptake of iron in environments with changing redox

422   conditions and in sunlit environments where Fe(II) is produced photochemically. Examples are soils

423   where reducing conditions are not sufficient for reductive dissolution of Fe-bearing phases, but where

424   traces of Fe(II) could lead to strongly accelerated Fe(II)-catalyzed dissolution in the presence of syn-

425   thetic or natural ligands (e.g. siderophores), [18] as also shown in our parallel study for a range of minerals

426   with HBED and DFOB. [36] Other environments where Fe(II)-catalyzed dissolution could be important

427   are oxic-anoxic transition zones in lakes, wetlands, and irrigated agricultural fields. Added Fe(II) and

428   photoproduced Fe(II) did not appear lead to formation of mobile charge or Fe(II) that is protected from

429   oxidation by oxygen, e.g. by charge residing in the bulk of the solid, at least not with Lp. Nevertheless,

430   steady state concentrations of photo-produced Fe(II) at the surface of iron oxide particles under sunlit

431   conditions could make Fe(II)-catalyzed dissolution an important pathway in their photoreductive disso-

432   lution. It is also possible that negative charge from added or produced Fe(II) can persist longer in the

433   bulk of more conductive iron(hydr)oxide minerals and that this results in a longer lasting catalytic effect

16

434 of Fe(II) under oxic conditions. The time resolved ATR-FTIR measurements and ^{57}Fe isotope experi-

435 ments presented here have proven useful and will be used to investigate the effect of Fe(II) on the dis-

436 solution of other Fe(III)(hydr)oxide phases in different environments.

437

438 **Associated Content**

439 Supporting Information. List of chemicals; lepdicrocite synthesis and characterization (XRD

440 and FTIR spectra); details on ATR-FTIR measurements, list of duplicate experiments; FTIR

441 spectra of adsorbed EDTA with pH and concentration dependence, spectral changes with added

442 Fe(II)); input file for kinetic model, model fits for dissolution kinetics and adsorption isotherms.

443 (5 Tables, 12 Figures).

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455 **References**

456

457 1. Kraemer, S. M., Iron oxide dissolution and solubility in the presence of siderophores. *Aquat. Sci.*
458   **2004,** *66* (1), 3-18.
459 2. Sandy, M.; Butler, A., Microbial iron acquisition: Marine and terrestrial siderophores. *Chem. Rev.*
460   **2009,** *109* (10), 4580-4595.
461 3. Shenker, M.; Chen, Y., Increasing Iron Availability to Crops: Fertilizers, Organo-Fertilizers, and
462   Biological Approaches. *J. Soil Sci. Plant Nutr.* **2005,** *51*, (1), 1-17.
463 4. Cheah, S. F.; Kraemer, S. M.; Cervini-Silva, J.; Sposito, G., Steady-state dissolution kinetics of
464   goethite in the presence of desferrioxamine B and oxalate ligands: Implications for the microbial
465   acquisition of iron. *Chem. Geol.* **2003,** *198* (1-2), 63-75.
466 5. Cervini-Silva, J.; Sposito, G., Steady-state dissolution kinetics of aluminum-goethite in the
467   presence of desferrioxamine-B and oxalate ligands. *Environ. Sci. Technol.* **2002,** *36*, (3), 337-342.

6. Reichard, P. U.; Kraemer, S. M.; Frazier, S. W.; Kretzschmar, R., Goethite dissolution in the presence of phytosiderophores: Rates, mechanisms, and the synergistic effect of oxalate. *Plant Soil* **2005,** *276* (1-2), 115-132.

7. Reichard, P. U.; Kretzschmar, R.; Kraemer, S. M., Dissolution mechanisms of goethite in the presence of siderophores and organic acids. *Geochim. Cosmochim. Acta* **2007,** *71* (23), 5635-5650.

8. Wolff-Boenisch, D.; Traina, S. J., The effect of desferrioxamine B, enterobactin, oxalic acid, and Na-alginate on the dissolution of uranyl-treatedgoethite at pH 6 and 25°C. *Chem. Geol.* **2007,** *243*, (3-4) 357-368.

9. Liu, X.; Fu, J. W.; Da Silva, E.; Shi, X. X.; Cao, Y.; Rathinasabapathi, B.; Chen, Y.; Ma, L. Q., Microbial siderophores and root exudates enhanced goethite dissolution and Fe/As uptake by As-hyperaccumulator Pteris vittata. *Environ. Pollut.* **2017,** *223*, 230-237.

10. Blesa, M. A.; Borghi, E. B.; Maroto, A. J. G.; Regazzoni, A. E., Adsorption of EDTA and iron-EDTA complexes on magnetite and the mechanism of dissolution of magnetite by EDTA. *J. Colloid Interface Sci.* **1984,** *98* (2), 295-305.

11. Suter, D.; Siffert, C.; Sulzberger, B.; Stumm, W., Catalytic dissoluion of iron(III)(hydr)oxides by oxalic acid in the presence of Fe(II). *Naturwissenchaften* **1988,** *75* (11), 571-573.

12. Borghi, E. B.; Regazzoni, A. E.; Maroto, A. J. G.; Blesa, M. A., Reductive dissolution of magnetite by solutions containing EDTA and FeII. *J. Colloid Interface Sci.* **1989,** *130* (2), 299-310.

13. Wehrli, B.; Sulzberger, B.; Stumm, W., Redox processes catalyzed by hydrous oxide surfaces. *Chem. Geol.* **1989,** *78* (3-4), 167-179.

14. Torres, R.; Blesa, M. A.; Matijević, E., Interactions of metal hydrous oxides with chelating agents. IX. Reductive dissolution of hermatite and magnetite by aminocarboxylic acids. *J. Colloid Interface Sci.* **1990,** *134* (2), 475-485.

15. Suter, D.; Banwart, S.; Stumm, W., Dissolution of Hydrous Iron(III) Oxides by Reductive Mechanisms. *Langmuir* **1991,** *7* (4), 809-813.

16. Karametaxas, G.; Hug, S. J.; Sulzberger, B., Photodegradation of EDTA in the Presence of Lepidocrocite. *Environ. Sci. Technol.* **1995,** *29* (12), 2992-3000.

17. Wang, Z.; Schenkeveld, W. D. C.; Kraemer, S. M.; Giammar, D. E., Synergistic Effect of Reductive and Ligand-Promoted Dissolution of Goethite. *Environ. Sci. Technol.* **2015,** *49*, (12), 7236-7244.

18. Schenkeveld, W. D. C.; Wang, Z.; Giammar, D. E.; Kraemer, S. M., Synergistic Effects between Biogenic Ligands and a Reductant in Fe Acquisition from Calcareous Soil. *Environ. Sci. Technol.* **2016,** *50* (12), 6381-6388.

19. Kubicki, J. D.; Tunega, D.; Kraemer, S., A density functional theory investigation of oxalate and Fe(II) adsorption onto the (010) goethite surface with implications for ligand- and reduction-promoted dissolution. *Chem. Geol.* **2017,** *464*, 14-22.

20. Lee, J. C.; Kim, E. J.; Baek, K., Synergistic effects of the combination of oxalate and ascorbate on arsenic extraction from contaminated soils. *Chemosphere* **2017,** *168*, 1439-1446.

21. Handler, R. M.; Beard, B. L.; Johnson, C. M.; Scherer, M. M., Atom exchange between aqueous Fe(II) and goethite: An Fe isotope tracer study. *Environ. Sci. Technol.* **2009,** *43* (4), 1102-1107.

22. Gorski, C. A.; Scherer, M. M., Fe2+ sorption at the Fe oxide-water interface: A revised conceptual framework. In *ACS Symposium Series*, American Chemical Society: 2011; Vol. 1071, pp 315-343.

23. Gorski, C. A.; Handler, R. M.; Beard, B. L.; Pasakarnis, T.; Johnson, C. M.; Scherer, M. M., Fe atom exchange between aqueous Fe2+ and magnetite. *Environ. Sci. Technol.* **2012,** *46* (22), 12399-12407.

24. Latta, D. E.; Gorski, C. A.; Scherer, M. M., Influence of Fe2+-catalysed iron oxide recrystallization on metal cycling. Biochem. Soc. Trans. **2012,** *40*, (6), 1191-1197.

25. Handler, R. M.; Frierdich, A. J.; Johnson, C. M.; Rosso, K. M.; Beard, B. L.; Wang, C.; Latta, D. E.; Neumann, A.; Pasakarnis, T.; Premaratne, W. A. P. J.; Scherer, M. M., Fe(II)-catalyzed recrystallization of goethite revisited. *Environ. Sci. Technol.* **2014,** *48* (19), 11302-11311.

26. Joshi, P.; Fantle, M. S.; Larese-Casanova, P.; Gorski, C. A., Susceptibility of Goethite to Fe2+-Catalyzed Recrystallization over Time. *Environ. Sci. Technol.* **2017,** *51* (20), 11681-11691.

27. Joshi, P.; Gorski, C. A., Anisotropic Morphological Changes in Goethite during Fe2+-Catalyzed Recrystallization. *Environ. Sci. Technol.* **2016,** *50* (14), 7315-7324.

28. Zarzycki, P.; Rosso, K. M., Stochastic Simulation of Isotopic Exchange Mechanisms for Fe(II)-Catalyzed Recrystallization of Goethite. *Environ. Sci. Technol.* **2017,** *51* (13), 7552-7559.

29. Borer, P.; Hug, S. J.; Sulzberger, B.; Kraemer, S. M.; Kretzschmar, R., Photolysis of citrate on the surface of lepidocrocite: An in situ attenuated total reflection infrared spectroscopy study. *J. Phys. Chem. C* **2007,** *111* (28), 10560-10569.

30. Borer, P.; Hug, S. J.; Sulzberger, B.; Kraemer, S. M.; Kretzschmar, R., ATR-FTIR spectroscopic study of the adsorption of desferrioxamine B and aerobactin to the surface of lepidocrocite (γ-FeOOH). *Geochim. Cosmochim. Acta* **2009,** *73* (16), 4661-4672.

31. Borer, P.; Hug, S. J., Photo-redox reactions of dicarboxylates and α-hydroxydicarboxylates at the surface of Fe(III)(hydr)oxides followed with in situ ATR-FTIR spectroscopy. *J. Colloid Interface Sci.* **2014,** *416*, 44-53.

32. Borer, P. M.; Sulzberger, B.; Reichard, P.; Kraemer, S. M., Effect of siderophores on the light-induced dissolution of colloidal iron(III) (hydr)oxides. *Mar. Chem.* **2005,** *93*, 179– 193.

33. Borer, P.; Kraemer, S. M.; Sulzberger, B.; Hug, S. J.; Kretzschmar, R., Photodissolution of lepidocrocite (γ-FeOOH) in the presence of desferrioxamine B and aerobactin. *Geochim. Cosmochim. Acta* **2009,** *73* (16), 4673-4687.

34. Borer, P.; Sulzberger, B.; Hug, S. J.; Kraemer, S. M.; Kretzschmar, R., Wavelength-dependence of photoreductive dissolution of lepidocrocite (γ-FeOOH) in the absence and presence of the siderophore DFOB. *Environ. Sci. Technol.* **2009,** *43* (6), 1871-1876.

35. Borer, P.; Sulzberger, B.; Hug, S. J.; Kraemer, S. M.; Kretzschmar, R., Photoreductive dissolution of iron(III) (Hydr)oxides in the absence and presence of organic ligands: Experimental studies and kinetic modeling. *Environ. Sci. Technol.* **2009,** *43* (6), 1864-1870.

36. Kang, K.; Schenkeveld, W. D. C.; Biswakarma, J.; Borowski, S. C.; Hug, S. J.; Hering, J. G.; Kraemer, S. M., Low Fe(II) concentrations catalyze the dissolution of various Fe(III) (hydr)oxide minerals in the presence of diverse ligands and over a broad pH range. *Environ. Sci. Technol.* **2018**.

37. Schwertmann, U.; Cornell, R. M., *Iron Oxides in the Laboratory*. 2nd, Completely Revised and Extended Edition ed.; Wiley-VCH Weinheim, Germany, 2000.

38. Cornell, R. M.; Schwertmann, U., *The Iron Oxides: Structure, Properties, Reactions, Occurrences, and Uses*. 2nd ed.; ed.; Wiley-VCH: New York, NY, USA, 2003.

39. Norén, K.; Loring, J. S.; Bargar, J. R.; Persson, P., Adsorption mechanisms of EDTA at the water-iron oxide interface: Implications for dissolution. *J. Phys. Chem. C* **2009,** *113*, (18), 7762-7771.

40. Zhang, Y.; Charlet, L.; Schindler, P. W., Adsorption of protons, Fe(II) and Al(III) on lepidocrocite (γ-FeOOH). *Colloids Surf.* **1992,** *63*, (3-4), 259-268.

41. Pedersen, H. D.; Postma, D.; Jakobsen, R.; Larsen, O., Fast transformation of iron oxyhydroxides by the catalytic action of aqueous Fe(II). *Geochim. Cosmochim. Acta* **2005,** *69*, (16), 3967-3977.

42. Jones, A. M.; Collins, R. N.; Rose, J.; Waite, T. D., The effect of silica and natural organic matter on the Fe(II)-catalysed transformation and reactivity of Fe(III) minerals. *Geochim. Cosmochim. Acta* **2009,** *73* (15), 4409-4422.

43. Ballesteros, M. C.; Rueda, E. H.; Blesa, M. A., The influence of iron (II) and (III) on the kinetics of goethite dissolution by EDTA. *J. Colloid Interface Sci.* **1998,** *201*, (1), 13-19.

44. Banwart, S.; Davies, S.; Stumm, W., The role of oxalate in accelerating the reductive dissolution of hematite (α-Fe2O3) by ascorbate. *Colloids Surf.* **1989,** *39* (2), 303-309.

45. Nowack, B.; Lutzenkirchen, J.; Behra, P.; Sigg, L., Modeling the adsorption of metal-EDTA complexes onto oxides. *Environ. Sci. Technol.* **1996,** *30* (7), 2397-2405.

46. Nowack, B.; Sigg, L., Adsorption of EDTA and metal-EDTA complexes onto goethite. *J. Colloid Interface Sci.* **1996,** *177* (1), 106-121.

47. Samson, S. D.; Eggleston, C. M., The depletion and regeneration of dissolution-active sites at the mineral-water interface: II. Regeneration of active sites on alpha-Fe2O3 at pH 3 and pH 6. *Geochim. Cosmochim. Acta* **2000,** *64* (21), 3675-3683.
48. Wehrli, B., Monte Carlo simulations of surface morphologies during mineral dissolution. *J. Colloid Interface Sci.* **1989,** *132* (1), 230-242.
49. Borowski, S. C.; Biswakarma, J.; Kang, K.; Schenkeveld, W. D. C.; Hering, J. G.; Kubicki, J. D.; Kraemer, S. M.; Hug, S. J., Structure and reactivity of oxalate surface complexes on lepidocrocite derived from infrared spectroscopy, DFT-calculations, adsorption, dissolution and photochemical experiments. *Geochim. Cosmochim. Acta* **2018,** *226*, 244-262.

Table. 1. Kinetic model with list of reactions and fitted equilibrium constants, rate coefficients and surface site concentrations for pH 6.0 and 9.5 mM NaCl [a]

Nr.	Reaction			Nr. of K or k	Description	*Fixed* or fitted log(K) or *log(k)*
R1	Fe^{II} + L	⇌	Fe^{II}L	K_1	Complexation of Fe^{II} in solution	10.3 [b]
R2	≡Fe^{III} + L	⇌	≡Fe^{III}-L	K_2	Adsorption of ligand (L)	5.0 [c]
R3	≡Fe^{III} + Fe^{II}	⇌	≡Fe^{III}-Fe^{II}	K_3	Adsorption of Fe^{II}	5.5 [d]
R4	≡Fe^{III}-Fe^{II} + L	⇌	≡Fe^{III}-Fe^{II}-L	K_4	Adsorption of L on adsorbed Fe^{II}	4.7 [e]
R5	≡Fe^{III} + Fe^{II}L	⇌	≡Fe^{III}-Fe^{II}-L	K_5	Adsorption of Fe^{II}L	5.4
R6	≡Fe^{III} + Fe^{III}L	⇌	≡Fe^{III}-Fe^{III}-L	K_6	Competing adsorption of Fe^{III}L	5.7 [d]
R7	≡Fe^{III}-Fe^{II}-L	→	≡Fe^{II}-Fe^{III}-L	$k_{7(ET)}$	Electron transfer	*1*
R8	≡Fe^{II}-Fe^{III}-L	→	≡Fe^{II} + Fe^{III}L	k_8	Fe(II)-catalyzed detachment of Fe^{III}-L	*-0.64*
R9	≡Fe^{II} + Lp_{bulk}	→	≡Fe^{III} -Fe^{II}	k_9	Formation of new surface site	*6* [f]
R10	≡Fe^{III}-L	→	≡ + Fe^{III}L	k_{10}	Detachment of Fe^{III}-L	*-2.5*
R11	≡ + Lp_{bulk}	→	≡Fe^{III}	k_{11}	Formation of new surface site	*6* [f]
Fitted active surface site conc. (last column, [≡Fe^{III}]/ [Lp] (M/M) and ≡Fe^{III}/nm^2) for dissolution, and surface site conc. for adsorption of EDTA, Fe(II) and Fe(III)EDTA in batch experiments						
	≡Fe^{III}			p_1	Dissolution in batch and FTIR experiments	6.3e-5 [g] 6.7e-3 ≡Fe^{III} /nm^2
	≡Fe^{III}			p_2	Adsorption of EDTA with FTIR (10 µM Lp)	7.4e-3 [c] 0.79 ≡Fe^{III} /nm^2
	≡Fe^{III}			p_3	Ads. of Fe(II) in batch experiments	1.0e-3 [d] 0.11 ≡Fe^{III} /nm^2
	≡Fe^{III}			p_4	Ads. of Fe(III)EDTA	1.3e-3 [d] 0.14 ≡Fe^{III} /nm^2

[a] Surface sites ≡Fe^{III}-OH are abbreviated as ≡Fe^{III} and surface complexes ≡Fe^{III}-O-Fe^{II} as ≡Fe^{III}-Fe^{II}. Surface complexes that involve both ligand and Fe(II) in close proximity are formulated as ≡Fe^{III}-Fe^{II}-L. This could mean a ternary surface complexes or adsorbed EDTA next to a reduced site.

[b] Conditional complex formation constant for Fe(II)EDTA at pH 6 (see SI). (Over 99.8% of Fe(II) is present as $Fe^{II}EDTA^{2-}$ with 50 µM EDTA).

[c] Best fits to adsorption isotherm for EDTA determined with FTIR, see SI Figures S10 and S11.

[d] Best fits to adsorption isotherms of Fe(II) and Fe(III)EDTA, see SI Table S5 and Figure S11.

[e] The rate coefficients of R4 are interdependent with the rate of Fe^{II} detachment (back reaction R3), formation of Fe^{II}L (R1) and adsorption of Fe^{II}L (R5), see SI.

[f] Arbitrary, fast and non-rate determining rate coefficients.

[g] Best fits to Lp-dissolution in batch (1.13 mM Lp) and FTIR (10 µM Lp = 35.6 µg/40ml) experiments, see SI.
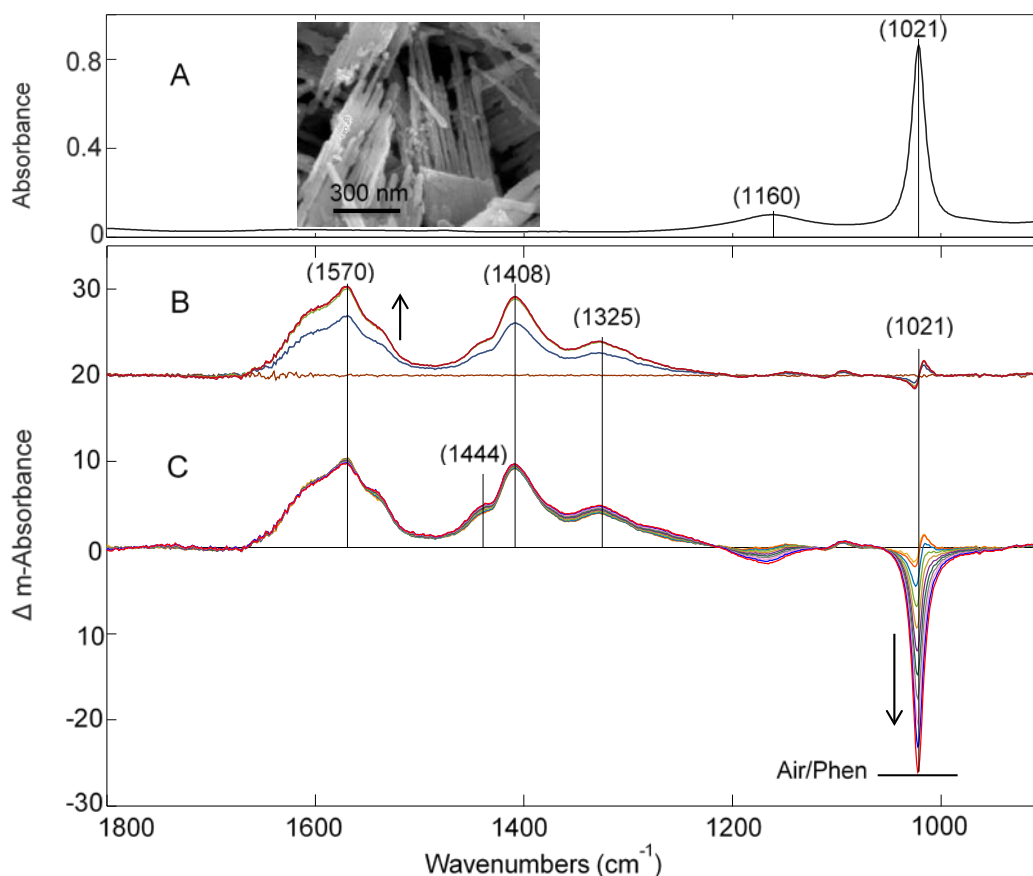
Fig.1. (A) ATR-FTIR absorbance spectrum of lepidocrocite (Lp) and SEM image of Lp (inset).

(B) Difference absorbance spectra recorded during EDTA adsorption (offset by 20 units), upon addition

of 50 µM EDTA to a layer of 40-60 µg Lp in contact with 40 ml aqueous solution (10 mM NaCl and

5 mM MES, pH 6.0). (C) Dissolution of Lp after addition of 1 µM Fe(II) in presence of 50 µM EDTA.

Spectra were recorded continuously every 43 s. Averages of every 10 continuous spectra are shown for

clarity. All measurements were conducted at pH 6 under anoxic condition by purging the solution with

high purity N_2. The group of spectra in (B) shows adsorption of EDTA by the increase of peaks at

1570 cm^{-1} and 1408 cm^{-1}. The spectra in (C) show the strongly accelerated dissolution of Lp by the

decrease of the absorbance of Lp at 1021 cm^{-1} after addition of 1 µM Fe(II) ($t_{Fe(II)\ addition}$=1843 s).

Introduction of synthetic air or 1 mM phenanthroline (Phen) (at t = 6300 s) stopped the dissolution.
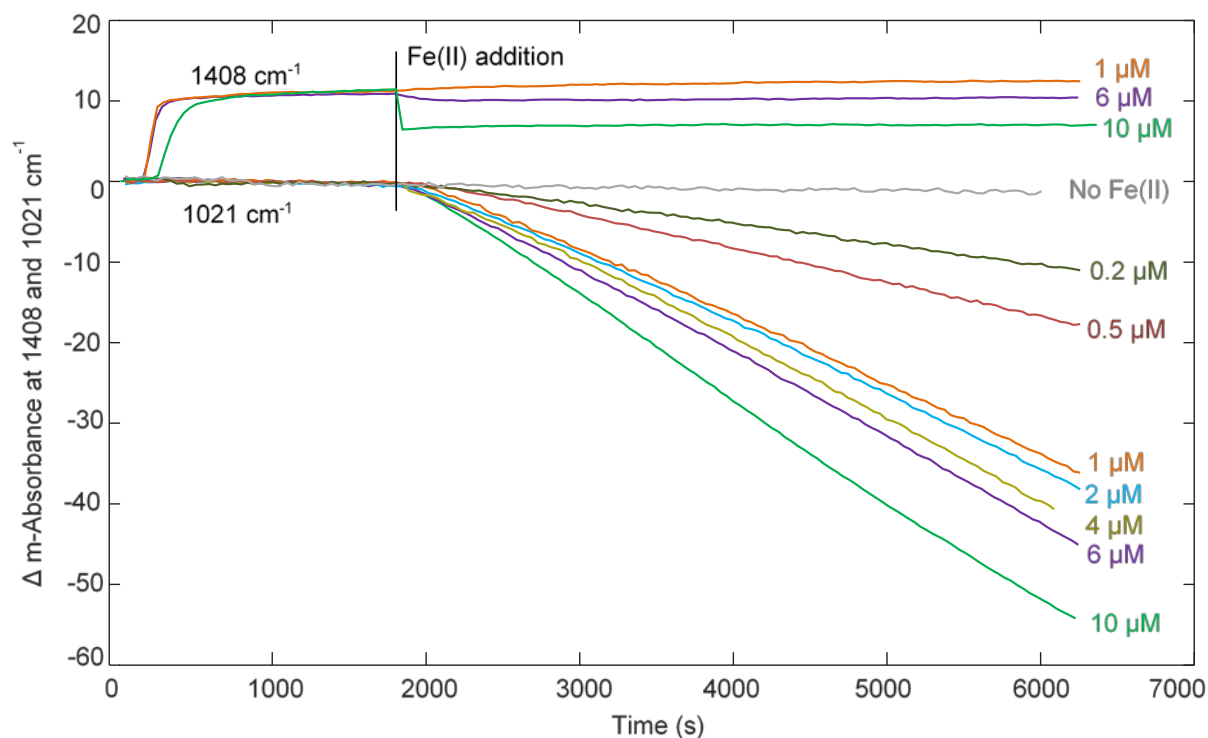
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Fig.2. Kinetics of EDTA adsorption and Lp dissolution (pH 6.0, anoxic conditions) monitored with ATR-FTIR. Adsorption of EDTA was monitored at 1408 cm^{-1} and dissolution of Lp at 1021 cm^{-1}. 50 µM EDTA was added (at t=180-200 s) after purging the aqueous solution covering the Lp layer with N$_2$ for at least 3-4 h, followed by addition of Fe(II) (0.2-10 µM) after 1800 s. EDTA adsorbed rapidly and reached equilibrium in < 600 s. Fe(II) additions lead to accelerated dissolution of Lp. Slopes (at 1021 cm^{-1}) were determined from 3000-5000 s.
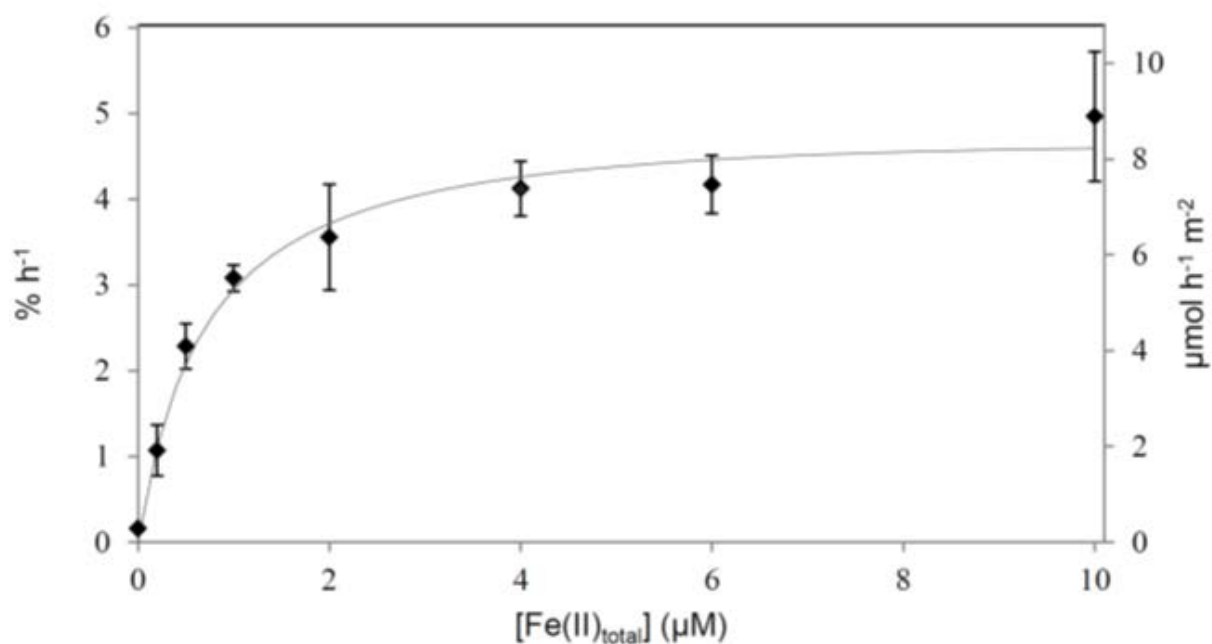
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Fig.3. Rates of Lp dissolution as a function of added (total) Fe(II) concentrations. Rates expressed as % h^{-1} (left Y-axis) were calculated from the slopes of the lines in Fig. 2 divided by the initial absorbance of Lp (at 1021 cm^{-1}) for each experiment. Rates expressed as µmol h^{-1} m^{-2} (right Y-axis) were calculated using the molecular weight (88.85 g mol^{-1}) and surface area (63 m^2 g^{-1}) of Lp. The solid line corresponds to a kinetic model (Table 1) as described in the text.

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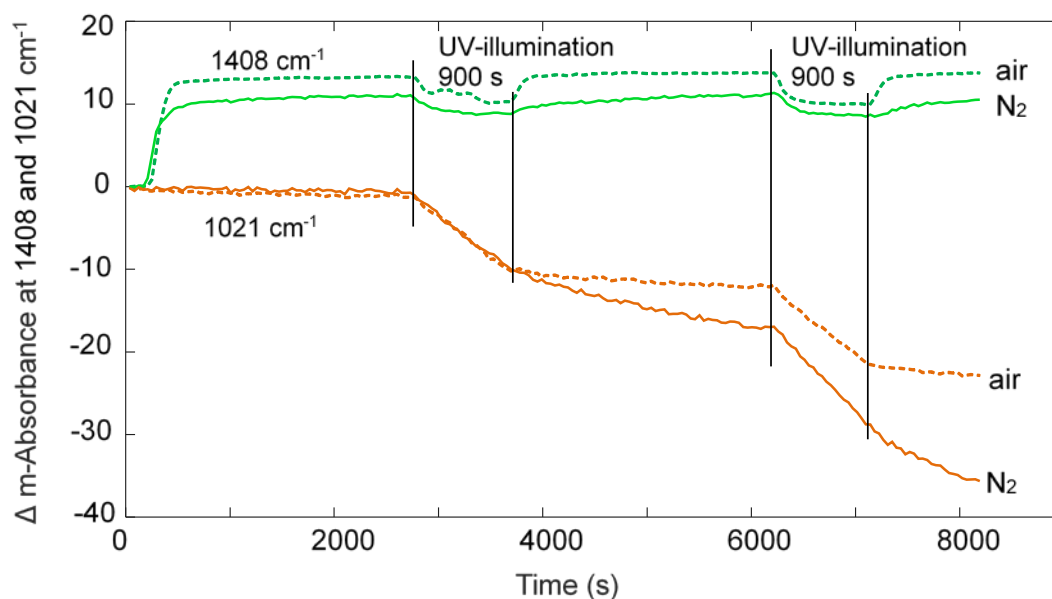
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688 **Fig.4.** Kinetics of EDTA adsorption (monitored at 1408 cm^{-1}) and dissolution of Lp (monitored at

689 1021 cm^{-1}) at pH 6 (I=0.01 M) during photochemical experiments. Dashed lines represent experiments

690 performed under oxic conditions and solid lines under anoxic conditions. EDTA (50 µM) was added at

691 180-220 s. During UV-illumination (for 900 s) with a 365 nm UV-LED lamp, adsorbed EDTA is

692 photolyzed at the surface. After irradiation, photo-transformed EDTA is replaced by EDTA from solu-

693 tion. Lp dissolution was strongly accelerated by UV-illumination under both oxic and anoxic conditions.

694 After irradiation stopped, there was no or very minimal continuing Lp dissolution under oxic condition.

695 In contrast, under anoxic condition, dissolution of Lp continued.
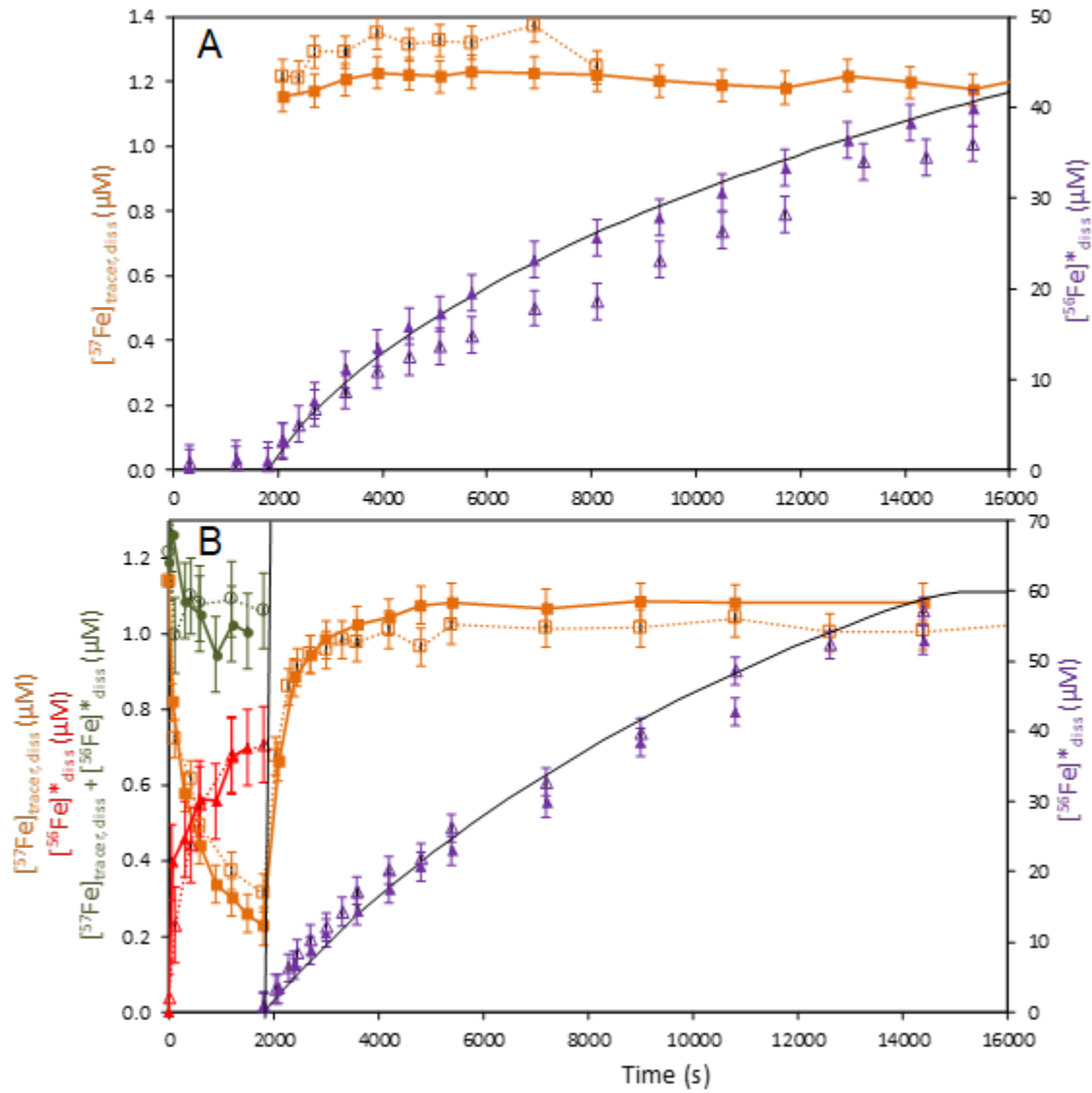
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Figure 5. Dissolution of (1125 µM) Lp at pH 6.0 under anoxic conditions catalyzed by addition of 1.2 µM Fe(II) as ^{57}Fe(II) either (a) 1800 s after the addition of 50 µM EDTA or (b) before EDTA addition. Adsorption and isotopic exchange of added ^{57}Fe(II) were observed when (b) 60 µM EDTA was added 1800 s after Fe(II). Symbols: (purple triangles, right axis) concentration of Fe released into solution by Lp dissolution (reported values of $[^{56}Fe]^*_{diss}$ correspond to the sum of all Fe-isotopes from Lp as described in Materials and Methods); (orange squares, left axis) concentration of the ^{57}Fe tracer remaining in or released back into solution where $[^{57}Fe]_{tracer,diss}$ was corrected for ^{57}Fe released from Lp; (red triangles, left axis) concentration of ^{56}Fe in solution resulting from isotopic exchange of added ^{57}Fe with ^{56}Fe in Lp (t < 1800 s); (green circles, left axis) sum of dissolved Fe measured as $[^{57}Fe]_{tracer, diss}$ and $[^{56}Fe]^*_{diss}$. Results of duplicate experiments are shown in open and closed symbols (note that results shown in Fig. 5b with filled symbols were scaled to compensate for the addition of a lower concentration (1.0 µM) of ^{57}Fe(II) in this experiment). Error bars correspond to the standard deviations of ICP-MS measurements obtained from repeated calibrations. Solid black lines show kinetic model fits (Table 1).