

SUPPLEMENTARY MATERIAL

Synthetic surfactants in Swiss sewage sludges: Analytical challenges, concentrations and per capita loads

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S1. Analytical methods and internal standard mixture

Analysis of SAS, LAS, and AEO

The target analytes SAS, LAS, and AEO were determined using an ultra-performance liquid chromatography (UPLC) system coupled to a Waters Xevo G2-XS time-of-flight mass spectrometer (TOF-MS). The LC-MS parameters were modified based on Lara-Martín et al. (2011) and Lara-Martín et al. (2006). A 10 μL volume of each sample was injected onto a RP-C18 column (Acquity UPLC BEH C18 (Waters); 2.1 mm \times 50 mm and 1.7 μm particle size) and eluted with mobile phases of MeOH (A) and 10 mM formic acid (B) at a flow rate of 0.4 mL min^{-1} . The initial solvent mixture of 70% A and 30% B was held for 2 min before ramping up to 100% A within 1 min and holding for 5 min. The initial solvent mixture was then restored and held until column re-equilibration. The MS was operated with electrospray ionization (ESI) in full scan negative and positive modes for detecting anionic and nonionic surfactants, respectively. Desolvation gas flow and temperature were 750 L h^{-1} and 350°C. The source temperature was 150°C. Capillary voltage was 2500 V and 4000 V in negative and positive ESI mode, respectively. The cone gas flow was 10 L h^{-1} and cone voltages were 45 V in negative and 10 V in positive mode. Lock mass was a solution of 200 ng/mL of leu-enkephalin. Ions were extracted using MassLynx software.

Analysis of NP, NPEC, and *tert*-OP

The target analytes *tert*-OP, NP, and NPEC were determined using a UPLC system coupled to an Agilent 6495 triplequad MS with a jet stream ESI ion source and ion funnel technology. A 10 μL volume of each sample was injected onto a RP-C18 column (Acquity UPLC HSS T3 3 \times 100 mm; 1.8 μm) and eluted with MeOH (A) and 10 mM ammonium formate (B) as mobile phases at a flow rate of 0.4 mL min^{-1} . The initial solvent mixture of 100% A and 0% B was held for 15 minutes before ramping up to 80% B within 1 min and holding for 4 min. The initial solvent mixture was restored and held until re-equilibration. A solution of 0.5 mM ammonium fluoride was introduced to the sample post-column at a flow rate of 0.05 mL min^{-1} . The MS was operated with electrospray ionization (ESI) in MRM negative mode for monitoring fragmentation reactions of deprotonated molecule ions of *tert*-OP (205.2 \rightarrow 133), NP (219.2 \rightarrow 133) and NP1EC (277.2 \rightarrow 219). Internal standards (D_8 -NP and D_2 -NP1EC) were used to correct for signal fluctuations. Ions were extracted and quantified using the Agilent Masshunter software. Other MS parameters were: gas temperature 200°C, gas flow 17 L min^{-1} , nebulizer 25 psi, sheath gas heater 400°C, sheath gas flow 12 L min^{-1} , and capillary voltage

3000V. Ion funnel parameters for negative high- and low-pressure radio frequency (RF) were 90 and 60, respectively.

Analysis of NPEO (EO1–3)

The target NP1EO, NP2EO, and NP3EO analytes were analyzed with a UPLC system coupled to a Bruker EvoQ Elite tandem mass spectrometer (MS/MS). The LC-MS parameters were optimized from a previously existing methodology (Lara-Martín et al. 2012). A 10 μL volume of each sample was injected onto a RP-C18 column (Bruker Intensity Solo, 2.1 mm \times 100 mm and 2 μm particle size) and eluted with MeOH (A) and 10 mM ammonium formate (B) as mobile phases at a flow rate of 0.4 mL min^{-1} . The initial solvent mixture of 70% A and 30% B was ramped up to 100% A within 1 min and held for 4 min. The initial solvent mixture was restored and held until re-equilibration. The MS was operated with electrospray ionization (ESI) in the MRM positive mode to monitor the fragmentation reactions of the ammonium adduct ions of NP1EO (282 \rightarrow 127; 282 \rightarrow 71), NP2EO (326 \rightarrow 183; 326 \rightarrow 121), and NP3EO (370 \rightarrow 227; 370 \rightarrow 121). Internal standards ($^{13}\text{C}_6$ -NP1EO and $^{13}\text{C}_6$ -NP2EO) were used to correct for signal fluctuations. Ions were extracted and quantified using the MS Workstation 8.1 software. Other MS parameters were: spray voltage 4500 V, cone temperature 350°C, and probe temperature 450°C.

Internal standard mixture

The final sludge samples (1 mL) were spiked with 10 μL of an internal standard mixture (before dilution). Blank and calibration samples were spiked, too. The internal standard mixture contained the following analytes in the respective concentrations: C_{16} -LAS (400 $\mu\text{g } \mu\text{L}^{-1}$), $\text{C}_{10}\text{EO8-AEO}$ (80 $\mu\text{g } \mu\text{L}^{-1}$), $\text{C}_{12}\text{-SAS}$ (160 $\mu\text{g } \mu\text{L}^{-1}$), $\text{D}_8\text{-NP}$ (40 $\mu\text{g } \mu\text{L}^{-1}$), $\text{D}_2\text{-NP1EC}$ (0.4 $\mu\text{g } \mu\text{L}^{-1}$) and $^{13}\text{C}_6\text{-NPEO}$ (160 $\mu\text{g } \mu\text{L}^{-1}$).

83 **Table S 1.** Monitored quantifier/qualifier ions and instrumental LOQs of the analyzed surfactants.

| Compound | Quantifier m/z | Qualifier m/z | LOQ* (instrumental) ng mL ⁻¹ |
|---|---------------------|--------------------------------------|--|
| C ₁₀ -LAS | 297.1530 | 298.1562 | 1.34 |
| C ₁₁ -LAS | 311.1686 | 312.1719 | 3.21 |
| C ₁₂ -LAS | 325.1843 | 326.1875 | 3.06 |
| C ₁₃ -LAS | 339.1999 | 340.2032 | 2.24 |
| C ₁₄ -SAS | 277.1843 | 278.1875 | 3.1 |
| C ₁₅ -SAS | 291.1999 | 292.2032 | 3.2 |
| C ₁₆ -SAS | 305.2156 | 306.2188 | 2.3 |
| C ₁₇ -SAS | 319.2312 | 320.2345 | 1.4 |
| AEO homologues M → quasi molecular ion | [M+Na] ⁺ | [¹³ C-M+Na] ⁺ | 10 |
| NP1EO | 282 → 127 | 282 → 71 | 0.3 |
| NP2EO | 326 → 183 | 326 → 121 | 0.3 |
| NP3EO | 370 → 227 | 370 → 121 | 0.3 |
| NP | 219.2 → 133 | n/a | 10 |
| NP1EC | 277.2 → 219 | n/a | 10 |
| tert-OP | 205.2 → 133 | n/a | 1 |

* LOQ: The limit of quantification refers to the instrumental LOQ and is also the lowest calibration standard. It is at least five times higher than the detected noise.

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S2. Recoveries

Recoveries were determined in a preliminary experiment. A sludge sample was spiked (in triplicate) with a known amount of analyte to determine the loss during the sample preparation. The resulting recovery values were then applied to correct the quantified levels in all sludge samples. Recoveries for all target analytes are listed in **Table S 2** to **Table S 6**. Recovery values for AEO are discussed in more detail in the main manuscript. Recovery values for LAS are discussed in the following paragraphs.

For the spiking experiments, the spiked concentration needs to be higher (e.g., 2-fold higher) than the level in the unspiked sample. Sewage sludge is highly contaminated with LAS, so it requires very high spike levels, which can be achieved either by using a highly concentrated spike solution or by using a larger spike volume. The spike solution concentration was limited by the solubility of LAS in solvent. Adding water to the solvent increased the solubility but also resulted in foaming and subsequent irreproducible spiking. The spike volume was further limited by the sample amount (0.2g), since the sample should not be dispersed in the solvent. Ultimately, in this study, the spiked LAS level accounted only for one fifth of the actual environmental level, resulting in uncertain recovery values with high standard deviation. Recoveries were $166 \pm 41\%$, $124 \pm 15\%$, $126 \pm 5\%$, and $284 \pm 14\%$ for C_{10} , C_{11} , C_{12} , and C_{13} , respectively. A previous study applied the same preparation method to sediment samples, which are less contaminated; hence, spiking experiments were possible without the mentioned restraints. That study achieved recoveries between 70 and 107% (Lara-Martín et al. 2008). We assume that the method would give similar results when applied to sewage sludge. Applying a conservative approach, a recovery of 100% for LAS homologues is assumed in this study. Due to the high concentrations of LAS in sewage sludge, the final sample extract must be diluted to avoid saturation of the mass spectrometer detector.

C_{10} -SAS was used to verify that no major sample losses occurred during sample preparation. C_{10} -SAS has a good recovery (about $85 \pm 13\%$) with the sample preparation method used here. The sample extract was spiked with C_{10} -SAS before further processing. If the detected C_{10} -SAS deviated by 20% from the average recovery, the sample was excluded from further analysis. Only one sample was ultimately excluded from the data set (deviation of more than 70%). All other samples had deviations of less than 10%.

118 **Table S 2.** Recoveries for LAS of different chain lengths (C₁₀-C₁₃).

| C10-LAS | C11-LAS | C12-LAS | C13-LAS |
|-----------|------------|-----------|------------|
| 166 ± 41% | 124% ± 15% | 126% ± 5% | 284% ± 14% |

120 **Table S 3.** Recoveries for SAS of different chain lengths (C₁₄-C₁₇).

| C14-SAS | C15-SAS | C16-SAS | C17-SAS |
|-----------|------------|------------|-----------|
| 67 ± 0.6% | 96% ± 2.6% | 76% ± 2.4% | 74 ± 3.5% |

122 **Table S 4.** Recoveries for SAS of different chain lengths (C₁₂, C₁₄, C₁₆, C₁₈) and different ethoxymer units (EO02, EO03, EO06, and EO08).

| | EO02 | EO03 | EO06 | EO08 |
|-----|-----------|-----------|-----------|-----------|
| C12 | 27 ± 2.2% | 38 ± 1.7% | 54 ± 0.9% | 55 ± 3.2% |
| C14 | 69 ± 2.3% | 74 ± 5.3% | 81 ± 5.5% | 90 ± 3.3% |
| C16 | 56 ± 3.6% | 63 ± 1.9% | 73 ± 2.0% | 75 ± 2.6% |
| C18 | 50 ± 1.7% | 63 ± 2.1% | 68 ± 2.0% | 98 ± 3.9% |

125 **Table S 5.** Recoveries for NP and NPEO (with 1–3 ethoxymer units).

| NP | NP1EO | NP2EO | NP3EO |
|----------|-----------|----------|-----------|
| 87 ± 13% | 87 ± 18%* | 87 ± 24% | 79 ± 5.9% |

* Estimated by averaging the recoveries of NP and NP2EO.

128 **Table S 6.** Recoveries for n-OP and NP1EC. The recovery of linear n-OP is used for *tert*-OP.

| n-OP | NP1EC |
|----------|-----------|
| 58 ± 12% | 74 ± 4.2% |

S3. Mass spectrometric responses of AEO homologues

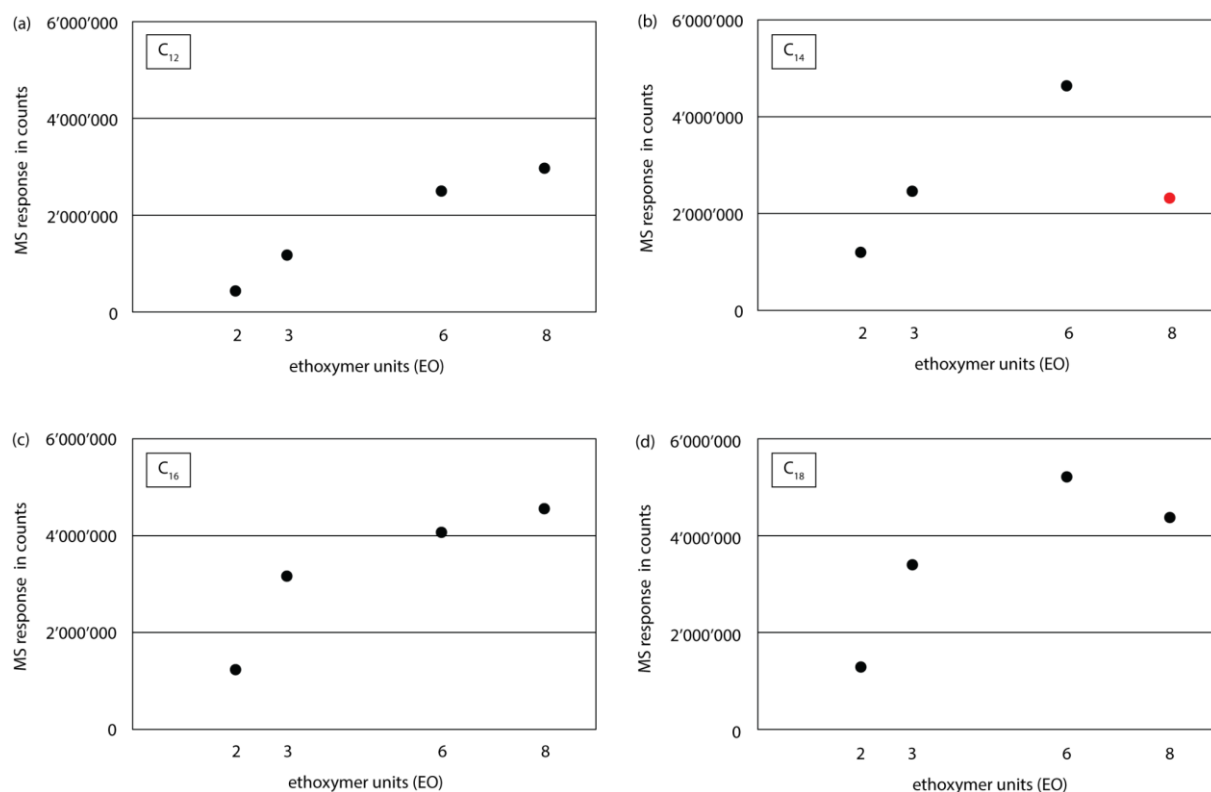


Figure S 1. Mass spectrometric responses for the available AEO standard material at a concentration of 0.1 mg L⁻¹. The responses are plotted for homologues with chain lengths (a) C₁₂, (b) C₁₄, (c) C₁₆, and (d) C₁₈ and with ethoxymers units EO2, EO3, EO6, and EO8. The ionization efficiency increases with increasing ethoxylation and, to some degree, with increasing alkyl chain length. The response for C₁₄EO8-AEO is considered an outlier (red). The calibration curve for C₁₄-EO6 was applied to C₁₄-homologues with higher ethoxylation degrees.

The sludge matrix causes a strong signal suppression for AEO in the mass spectrometer. At a strong dilution of 1:100 (i.e., an extract of 0.002 g sludge per mL), the matrix suppression is negligible but the concentration of AEO in sludge is then too low to be detected at this dilution. At a lower dilution of 1:5 (i.e., 0.04 g sludge per mL), AEO can be detected, but with a signal suppression of around 80%. The internal standard C₁₀EO08-AEO is used to account for the varying matrix effects of the sludge extracts.

S4. WWTP and sample specifications

Information about the sampled WWTPs and the samples are provided in the two following tables.

Sample containers were sent to the listed WWTPs with the instruction to sample anaerobic sludge from the digester. The requested sludge has preferably spend 25 days (on average) in the digester.

The (anaerobic) sludge samples were taken from the digester's outlet by the staff of the WWTPs and then shipped to our lab. Sample sizes were about 50 g or 500 mL, depending on whether the sludge was already dewatered or not.

The WWTPS ZOF, HER,VISP and ZAL do not apply sludge digestion and have provided non-digested sludge.

Table S 7a. Information about the sampled WWTPs and the samples.

| Sample ID | Name (location) of WWTP | Stabilized sludge production (kg d ⁻¹) | Connected population | Sampling year | State of sample | Average residence time in digester (d ⁻¹) |
|-----------|----------------------------|---|----------------------|---------------|--------------------|--|
| BUEH | Bühler | 280 | 5000 | 2019 | liquid | 45 |
| NEU | Neuchatel | 3800 | 40600 | 2017 | n/a | 39 |
| CHDF | La-Chaux-De-Fonds | 2300 | 37500 | 2017 | n/a | 50 |
| VEV | Vevey/Aviron | 5500 | 51900 | 2017 | n/a | 35 |
| UNT | Unterseen (Interlaken) | 1400 | 25400 | 2017 | n/a | 30 |
| WIN | Winterthur | 12600 | 130000 | 2017 | n/a | 30 |
| ZOF | Zofingen | 8000 | 34100 | 2019 | dewatered | 0 (no anaerobic digestion) |
| VERN | Vernier/Aire | 24500 | 440800 | 2017 | n/a | n/a |
| FEH | Fehraltorf-Russikon | 860 | 9600 | 2019 | liquid & dewatered | 27 |
| ZUE | Zuerich (Werdhoelzli) | 25700 | 429400 | 2017 | n/a | n/a |
| CHUR | Chur | 7600 | 54600 | 2017 | n/a | 18 |
| BRUE | Bruegg (Muera Biel) | 10000 | 85000 | 2017 | n/a | 30 |
| BRI | Brig-Glis (Briglina) | 3300 | 27700 | 2017 | n/a | n/a |
| POSC | Poschiavo | 480 | 3300 | 2017 | n/a | 20 |
| AAR | Aarau | 6600 | 74000 | 2017 | n/a | 25 |
| RAM | Ramsen (Bibertal-Hegau) | 5100 | 97600 | 2017 | n/a | 35 |
| GLA | Glarnerland | 3500 | 44700 | 2019 | solid | 23 |
| OBE | Oberriet | 720 | 8500 | 2019 | liquid | 23 |
| UETE | Uetendorf (Thun) | 5200 | 123000 | 2017 | n/a | n/a |
| BUCH | Buchs | 2200 | 23900 | 2017 | n/a | 28 |
| REIN | Reinach (Oberwynental) | 1800 | 17400 | 2017 | n/a | 25 |
| OPF | Opfikon-Kloten | 2900 | 37200 | 2017 | n/a | 6 |
| DUE | Duebendorf | 2600 | 36200 | 2017 | n/a | 14 |
| HER | Herisau Bachwis | 150 | 16200 | 2019 | liquid & dewatered | 0 (no anaerobic digestion) |
| SAM | Samedan | 650 | 3600 | 2019 (+2020) | liquid & dewatered | 23 |
| LAU | Lausanne | 23100 | 235000 | 2017 | n/a | n/a |
| BIO | Bioggio (Lugano) | 13400 | 125000 | 2017 | n/a | n/a |
| WEIN | Weinfelden (Mittelthurgau) | 2200 | 30100 | 2017 | n/a | 11 |
| MUEN | Münchwilen | 1000 | 20500 | 2019 | liquid & dewatered | 21 |
| BAS | Basel | 36300 | 268000 | 2017 | n/a | n/a |
| RAN | Rancate/Mendrisio | 4300 | 29000 | 2017 | n/a | 25 |
| BIRS | Birsfelden (Birs II) | 5400 | 82500 | 2017 | n/a | 15 |
| DAVG | Davos (Gadenstatt) | 1000 | 9900 | 2019 (+2020) | liquid | 42 |
| VISP | Visp/Lonza | 18600 | 13000 | 2019 | liquid | 0 (no anaerobic digestion) |
| POR | Porrentruy (Sepe) | 1500 | 16200 | 2017 | n/a | n/a |
| ZAL | Zala AG | 3200 | 39300 | 2019 | solid | 0 (no anaerobic digestion) |

162 **Table S 7b.** Information about the sampled WWTPs and their treatment processes.

| Sample ID | |
|-----------|--|
| BUEH | mechanical and biological treatment with phosphorous elimination |
| NEU | mechanical and biological treatment with advanced phosphorous elimination |
| CHDF | mechanical and biological treatment with advanced phosphorous elimination |
| VEV | mechanical and biological treatment with advanced phosphorous elimination |
| UNT | mechanical and biological treatment with advanced phosphorous elimination |
| WIN | mechanical and biological treatment with advanced phosphorous elimination, nitrification and denitrification |
| ZOF | mechanical and biological treatment with phosphorous elimination and nitrification |
| VERN | mechanical and biological treatment with phosphorous elimination and nitrification |
| FEH | mechanical and biological treatment with phosphorous elimination, nitrification and denitrification |
| ZUE | mechanical and biological treatment with advanced phosphorous elimination, nitrification and denitrification |
| CHUR | mechanical and biological treatment with advanced phosphorous elimination |
| BRUE | mechanical and biological treatment with advanced phosphorous elimination |
| BRI | mechanical and biological treatment with advanced phosphorous elimination |
| POSC | mechanical and biological treatment with advanced phosphorous elimination |
| AAR | mechanical and biological treatment with advanced phosphorous elimination |
| RAM | mechanical and biological treatment with phosphorous elimination, nitrification and denitrification |
| GLA | mechanical and biological treatment with phosphorous elimination, nitrification and denitrification |
| OBE | mechanical and biological treatment with phosphorous elimination and nitrification |
| UETE | mechanical and biological treatment with phosphorous elimination, nitrification and denitrification |
| BUCH | mechanical and biological treatment with phosphorous elimination and nitrification |
| REIN | mechanical and biological treatment with phosphorous elimination, nitrification and denitrification |
| OPF | mechanical and biological treatment with advanced phosphorous elimination, nitrification and denitrification |
| DUE | mechanical and biological treatment with advanced phosphorous elimination, nitrification and denitrification |
| HER | mechanical and biological treatment with phosphorous elimination |
| SAM | mechanical and biological treatment with advanced phosphorous elimination |
| LAU | mechanical and biological treatment with advanced phosphorous elimination |
| BIO | mechanical and biological treatment with advanced phosphorous elimination, nitrification and denitrification |
| WEIN | mechanical and biological treatment with phosphorous elimination and nitrification |
| MUEN | mechanical and biological treatment with phosphorous elimination, nitrification and denitrification |
| BAS | mechanical and biological treatment with advanced phosphorous elimination |
| RAN | mechanical and biological treatment with advanced phosphorous elimination and nitrification |
| BIRS | mechanical and biological treatment with advanced phosphorous elimination |
| DAVG | mechanical and biological treatment with advanced phosphorous elimination |
| VISP | mechanical and biological treatment with advanced phosphorous elimination |
| POR | mechanical and biological treatment with advanced phosphorous elimination |
| ZAL | mechanical and biological treatment with phosphorous elimination, nitrification and denitrification |

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S5. Concentrations of surfactants in sludge

This section summarizes the measured concentrations of the target analytes in Swiss sewage sludge. **Table S 8** to **Table S 12** give detailed data for LAS, SAS, AEO, NPEO, NP, NPEC, and *tert*-OP in the different WWTPs. The tables also list maximum, minimum, and average concentrations, as well as median values. **Figure S 2** shows summarizing boxplots for sludge concentrations and per capita loads for individual surfactant classes. **Figure S 3** plots sludge concentrations of (a) *tert*-OP and (b) NPEC for the different WWTPs. **Figure S 4** plots homologue distributions for (a) LAS and (b) SAS. The proportions of NP and *tert*-OP are plotted in **Figure S 5**. More detailed information about AEO homologues is given in the following **section S6**.

Table S 8. Levels of LAS in sewage sludge in $\mu\text{g g}^{-1}$ d.w. (dry weight) and the respective per capita loads.

| ID | C ₁₀ -LAS $\mu\text{g g}^{-1}$ d.w. | C ₁₁ -LAS $\mu\text{g g}^{-1}$ d.w. | C ₁₂ -LAS $\mu\text{g g}^{-1}$ d.w. | C ₁₃ -LAS $\mu\text{g g}^{-1}$ d.w. | Sum LAS $\mu\text{g g}^{-1}$ d.w. | LAS per capita load per capita mg d ⁻¹ |
|----------------|---|---|---|---|--------------------------------------|--|
| BUEH | 450 | 2500 | 3600 | 2900 | 9500 | 540 |
| NEU | 300 | 2100 | 3400 | 2600 | 8400 | 780 |
| CHDF | 340 | 2100 | 3100 | 2100 | 7800 | 480 |
| VEV | 270 | 1700 | 2600 | 2000 | 6600 | 700 |
| UNT | 300 | 1800 | 2500 | 1800 | 6400 | 360 |
| WIN | 300 | 1700 | 2500 | 1800 | 6300 | 610 |
| ZOF | 360 | 1700 | 2400 | 1800 | 6200 | 1500 |
| VERN | 250 | 1600 | 2400 | 1800 | 6000 | 340 |
| FEH | 280 | 1600 | 2200 | 1800 | 5900 | 530 |
| ZUE | 240 | 1500 | 2200 | 1600 | 5600 | 330 |
| CHUR | 200 | 1300 | 2000 | 1600 | 5100 | 700 |
| BRUE | 200 | 1300 | 2000 | 1600 | 5100 | 590 |
| BRI | 200 | 1400 | 2100 | 1400 | 5100 | 600 |
| POSC | 200 | 1300 | 2000 | 1400 | 4800 | 710 |
| AAR | 190 | 1300 | 1900 | 1400 | 4800 | 430 |
| RAM | 170 | 1100 | 1700 | 1300 | 4200 | 220 |
| GLA | 160 | 940 | 1500 | 1400 | 3900 | 310 |
| OBE | 160 | 870 | 1300 | 1000 | 3300 | 280 |
| UETE | 140 | 850 | 1300 | 970 | 3300 | 140 |
| BUCH | 74 | 630 | 1300 | 1200 | 3200 | 300 |
| REIN | 160 | 880 | 1200 | 800 | 3100 | 310 |
| OPF | 160 | 830 | 1200 | 730 | 2900 | 220 |
| DUE | 140 | 810 | 1100 | 760 | 2900 | 200 |
| HER | 120 | 580 | 930 | 990 | 2600 | 24 |
| SAM | 100 | 590 | 920 | 860 | 2500 | 450 |
| LAU | 89 | 530 | 860 | 680 | 2100 | 210 |
| BIO | 50 | 430 | 870 | 790 | 2100 | 230 |
| WEIN | 93 | 550 | 820 | 630 | 2100 | 160 |
| MUEN | 65 | 390 | 760 | 760 | 2000 | 98 |
| BAS | 89 | 470 | 720 | 530 | 1800 | 240 |
| RAN | 73 | 430 | 690 | 560 | 1800 | 260 |
| BIRS | 58 | 320 | 460 | 340 | 1200 | 76 |
| DAVG | 34 | 170 | 280 | 280 | 770 | 78 |
| VISP | 67 | 200 | 180 | 110 | 550 | 780 |
| POR | 8 | 62 | 130 | 140 | 340 | 31 |
| ZAL | 2 | 17 | 26 | 28 | 74 | 6 |
| Maximum | 450 | 2500 | 3600 | 2900 | 9500 | 1500 |
| Minimum | 2 | 17 | 26 | 28 | 74 | 6 |
| Average | 170 | 1000 | 1500 | 1200 | 3900 | 380 |
| Median | 160 | 880 | 1300 | 1100 | 3300 | 310 |

Table S 9. Levels of SAS in sewage sludge in $\mu\text{g g}^{-1}$ d.w. (dry weight) and the respective per capita loads.

| ID | C ₁₄ -SAS $\mu\text{g g}^{-1}$ d.w. | C ₁₄ -SAS $\mu\text{g g}^{-1}$ d.w. | C ₁₄ -SAS $\mu\text{g g}^{-1}$ d.w. | C ₁₄ -SAS $\mu\text{g g}^{-1}$ d.w. | Sum SAS $\mu\text{g g}^{-1}$ d.w. | SAS per capita load per capita mg d ⁻¹ |
|----------------|---|---|---|---|--------------------------------------|--|
| BUEH | 46 | 57 | 95 | 64 | 260 | 15 |
| NEU | 80 | 120 | 250 | 210 | 660 | 61 |
| CHDF | 76 | 110 | 220 | 180 | 600 | 36 |
| VEV | 62 | 87 | 170 | 140 | 460 | 49 |
| UNT | 76 | 97 | 170 | 130 | 470 | 26 |
| WIN | 53 | 71 | 130 | 97 | 350 | 34 |
| ZOF | 13 | 15 | 29 | 16 | 73 | 17 |
| VERN | 50 | 71 | 140 | 120 | 380 | 21 |
| FEH | 19 | 27 | 60 | 43 | 150 | 13 |
| ZUE | 46 | 61 | 120 | 90 | 320 | 19 |
| CHUR | 46 | 68 | 140 | 110 | 360 | 50 |
| BRUE | 44 | 63 | 130 | 98 | 330 | 39 |
| BRI | 39 | 58 | 120 | 81 | 290 | 35 |
| POSC | 25 | 34 | 63 | 44 | 170 | 25 |
| AAR | 37 | 56 | 110 | 76 | 270 | 25 |
| RAM | 17 | 25 | 52 | 37 | 130 | 6.9 |
| GLA | 22 | 29 | 67 | 49 | 170 | 13 |
| OBE | 24 | 33 | 69 | 44 | 170 | 14 |
| UETE | 27 | 37 | 71 | 51 | 190 | 7.9 |
| BUCH | 11 | 29 | 79 | 74 | 190 | 18 |
| REIN | 29 | 36 | 59 | 36 | 160 | 16 |
| OPF | 27 | 31 | 59 | 35 | 150 | 12 |
| DUE | 22 | 28 | 51 | 31 | 130 | 9.3 |
| HER | 2.2 | 0.94 | 1.8 | 1.2 | 6.2 | 0.06 |
| SAM | 26 | 38 | 81 | 63 | 210 | 38 |
| LAU | 9.7 | 14 | 30 | 20 | 73 | 7.2 |
| BIO | 2.1 | 5.1 | 19 | 18 | 45 | 4.8 |
| WEIN | 10 | 13 | 26 | 17 | 66 | 4.9 |
| MUEN | 7.9 | 11 | 23 | 17 | 59 | 3.0 |
| BAS | 10 | 12 | 22 | 13 | 56 | 7.6 |
| RAN | 5.9 | 7.2 | 15 | 10 | 39 | 5.7 |
| BIRS | 8.4 | 8.5 | 16 | 9.8 | 43 | 2.8 |
| DAVG | 2.8 | 2.8 | 7 | 4.4 | 17 | 1.7 |
| VISP | 0.09 | 0.03 | 0.03 | 0.01 | 0.15 | 0.21 |
| POR | 0.31 | 0.30 | 0.92 | 0.97 | 2.5 | 0.23 |
| ZAL | 0.29 | 0.04 | 0.10 | 0.01 | 0.44 | 0.04 |
| Maximum | 78 | 120 | 250 | 214 | 660 | 61 |
| Minimum | 0.09 | 0.03 | 0.03 | 0.01 | 0.15 | 0.04 |
| Average | 27 | 38 | 75 | 57 | 200 | 18 |
| Median | 23 | 30 | 62 | 44 | 160 | 14 |

Table S 10. Levels of AEO in sewage sludge in $\mu\text{g g}^{-1}$ d.w. (dry weight) and the respective per capita loads. The average chain lengths and proportion of branched AEO are also listed.

| ID | C ₁₂ -AEO | C ₁₃ -AEO | C ₁₄ -AEO | C ₁₅ -AEO | C ₁₆ -AEO | C ₁₇ -AEO | C ₁₈ -AEO | Sum AEO | AEO per capita load | Average chain length | Proportion of branched AEO |
|---------|---------------------------|---------------------------|---------------------------|---------------------------|---------------------------|---------------------------|---------------------------|---------------------------|-------------------------------|----------------------|----------------------------|
| | $\mu\text{g g}^{-1}$ d.w. | $\mu\text{g g}^{-1}$ d.w. | $\mu\text{g g}^{-1}$ d.w. | $\mu\text{g g}^{-1}$ d.w. | $\mu\text{g g}^{-1}$ d.w. | $\mu\text{g g}^{-1}$ d.w. | $\mu\text{g g}^{-1}$ d.w. | $\mu\text{g g}^{-1}$ d.w. | per capita mg d ⁻¹ | | |
| BUEH | 1.46 | 0.53 | 0 | 0.75 | 6.25 | 11.2 | 10.6 | 30.8 | 1.73 | 16.8 | 99.9 |
| NEU | 1.03 | 0.45 | 0.25 | 2.12 | 7.48 | 6.22 | 6.62 | 24.2 | 2.24 | 16.5 | 92.8 |
| CHDF | 0 | 0.07 | 0.12 | 1.46 | 3.82 | 4.86 | 3.76 | 14.1 | 0.85 | 16.7 | 99.4 |
| VEV | 0 | 0.05 | 0.05 | 0.96 | 3.6 | 3.37 | 1.8 | 9.82 | 1.05 | 16.6 | 99.5 |
| UNT | 0 | 8.74 | 0 | 0.71 | 5.51 | 5.53 | 3.8 | 24.3 | 1.34 | 15.4 | 99.8 |
| WIN | 0 | 0.17 | 0 | 0.9 | 2.14 | 3.74 | 2.11 | 9.06 | 0.88 | 16.7 | 96.3 |
| ZOF | 0.09 | 0.07 | 0.7 | 2.44 | 2.97 | 2.52 | 3.46 | 12.2 | 2.88 | 16.4 | 62.7 |
| VERN | 0 | 0 | 0 | 0.62 | 1.98 | 2.43 | 1.92 | 6.96 | 0.39 | 16.8 | 100.0 |
| FEH | 0 | 0 | 0 | 0.46 | 1.63 | 2.81 | 1.88 | 6.77 | 0.61 | 16.9 | 99.9 |
| ZUE | 0 | 0.01 | 0 | 0.39 | 2.44 | 2.4 | 1.62 | 6.85 | 0.41 | 16.8 | 99.9 |
| CHUR | 0 | 0.02 | 0 | 1.02 | 2.17 | 3.64 | 2.29 | 9.15 | 1.27 | 16.8 | 99.7 |
| BRUE | 0 | 0.02 | 0 | 0.28 | 0.79 | 1.52 | 0.98 | 3.59 | 0.42 | 16.9 | 98.3 |
| BRI | 0 | 0.06 | 0 | 0.52 | 0.81 | 2.31 | 1 | 4.7 | 0.56 | 16.8 | 98.6 |
| POSC | 1.71 | 3.46 | 0.02 | 0.03 | 0.08 | 1.14 | 0.67 | 7.11 | 1.05 | 13.9 | 100.0 |
| AAR | 1.13 | 4.75 | 0.03 | 0.47 | 6.51 | 1.94 | 5.55 | 20.4 | 1.82 | 15.7 | 98.9 |
| RAM | 0.64 | 1.72 | 0.03 | 0.33 | 1.53 | 2.09 | 2.27 | 8.6 | 0.45 | 15.8 | 99.7 |
| GLA | 0.01 | 0 | 0.37 | 3.39 | 6.3 | 7.24 | 7.22 | 24.5 | 1.92 | 16.7 | 100.0 |
| OBE | 0 | 0 | 0 | 0.17 | 0.71 | 1.27 | 0.97 | 3.12 | 0.26 | 17.0 | 100.0 |
| UETE | 0 | 0 | 0 | 0.49 | 1.72 | 1.92 | 1.31 | 5.43 | 0.23 | 16.7 | 100.0 |
| BUCH | 0 | 0 | 0 | 0.2 | 0.67 | 1.62 | 0.6 | 3.08 | 0.28 | 16.8 | 100.0 |
| REIN | 0 | 0 | 0 | 0.04 | 0.22 | 0.5 | 0.32 | 1.09 | 0.11 | 17.0 | 95.2 |
| OPF | 0 | 0 | 0 | 0.16 | 1.09 | 1.31 | 1.26 | 3.82 | 0.3 | 17.0 | 100.0 |
| DUE | 1.4 | 3.73 | 0 | 0 | 0.6 | 1.12 | 0.69 | 7.54 | 0.53 | 14.1 | 100.0 |
| HER | 0.11 | 2.13 | 0 | 0.21 | 1.49 | 3.09 | 2.99 | 10.0 | 0.09 | 16.2 | 97.9 |
| SAM | 0 | 0 | 0 | 0.94 | 1.69 | 2.44 | 1.86 | 6.93 | 1.26 | 16.8 | 98.5 |
| LAU | 0.02 | 0.54 | 0.08 | 1.56 | 0.83 | 1.1 | 1.55 | 5.69 | 0.56 | 16.1 | 53.0 |
| BIO | 0.07 | 0 | 0 | 0 | 0.03 | 0.66 | 0.44 | 1.19 | 0.13 | 17.1 | 94.3 |
| WEIN | 1.65 | 5.48 | 0.02 | 0 | 0.51 | 1.03 | 0.76 | 9.46 | 0.7 | 13.8 | 99.8 |
| MUEN | 0 | 0.3 | 0 | 0.14 | 0.59 | 1.18 | 0.53 | 2.74 | 0.14 | 16.4 | 100.0 |
| BAS | 1.06 | 0.7 | 0.64 | 1.72 | 1.36 | 0.71 | 1.39 | 7.58 | 1.02 | 15.2 | 34.9 |
| RAN | 0.02 | 0.61 | 0 | 0 | 0.81 | 1.15 | 1.23 | 3.82 | 0.56 | 16.5 | 99.6 |
| BIRS | 0.94 | 6.28 | 0 | 0.02 | 1.28 | 1.25 | 1.12 | 10.9 | 0.71 | 14.2 | 99.0 |
| DAVG | 0 | 0 | 0.06 | 0.7 | 0.83 | 1.62 | 0.69 | 3.9 | 0.4 | 16.6 | 100.0 |
| VISP | 2.41 | 8.5 | 1.21 | 1.11 | 0.5 | 0.56 | 0.55 | 14.8 | 21.3 | 13.5 | 55.7 |
| POR | 0 | 0.21 | 0 | 0 | 0 | 0.02 | 0.03 | 0.26 | 0.02 | 13.8 | 78.3 |
| ZAL | 0 | 0 | 0 | 0 | 0 | 0 | 0.05 | 0.05 | 0.004 | 18.0 | 0.0 |
| Maximum | 2.41 | 8.74 | 1.21 | 3.39 | 7.48 | 11.2 | 10.6 | 30.8 | 21.3 | 18.0 | 100.0 |
| Minimum | 0 | 0 | 0 | 0 | 0 | 0 | 0.03 | 0.05 | 0.004 | 13.5 | 0.0 |
| Average | 0.38 | 1.35 | 0.1 | 0.67 | 1.97 | 2.43 | 2.11 | 9.02 | 1.35 | 16.1 ± 1.1 | 90.3 ± 22.1 |
| Median | 0 | 0.07 | 0 | 0.46 | 1.32 | 1.77 | 1.35 | 7.03 | 0.56 | 16.7 | 99.6 |

Table S 11. Levels of NP, NPEO (EO1-3), NP1EC, and *tert*-OP in sewage sludge in $\mu\text{g g}^{-1}$ d.w. (dry weight). Per capita loads are listed in Table S 12.

| ID | NP $\mu\text{g g}^{-1}$ d.w. | NP1EO $\mu\text{g g}^{-1}$ d.w. | NP2EO $\mu\text{g g}^{-1}$ d.w. | NP3EO $\mu\text{g g}^{-1}$ d.w. | Sum NPEO $\mu\text{g g}^{-1}$ d.w. | NP1EC $\mu\text{g g}^{-1}$ d.w. | <i>tert</i> -OP $\mu\text{g g}^{-1}$ d.w. |
|----------------|---------------------------------|------------------------------------|------------------------------------|------------------------------------|---------------------------------------|------------------------------------|--|
| BUEH | 18.5 | 23.6 | 0.18 | 0.27 | 24.1 | 1.65 | 2.15 |
| NEU | 6.66 | 13.1 | 0.32 | 0.30 | 13.7 | 0.20 | 103 |
| CHDF | 13.6 | 5.4 | 0.14 | 0.25 | 5.8 | 0.43 | 4.31 |
| VEV | 2.03 | 32.6 | 0.33 | 0.52 | 33.5 | 0.41 | 2.42 |
| UNT | 7.94 | 10.4 | 0.17 | 0.31 | 10.9 | 0.58 | 0.98 |
| WIN | 4.16 | 12.8 | 0.24 | 0.34 | 13.4 | 0.27 | 0.81 |
| ZOF | 0.42 | 362 | 0.34 | 0.68 | 363 | 0.02 | 0.72 |
| VERN | 2.81 | 13.4 | 0.32 | 0.34 | 14.0 | 0.26 | 1.33 |
| FEH | 2.23 | 12.6 | 0.08 | 0.17 | 12.9 | 0.10 | 0.80 |
| ZUE | 4.47 | 18.1 | 0.22 | 0.35 | 18.7 | 0.46 | 2.62 |
| CHUR | 2.53 | 23.2 | 0.22 | 0.30 | 23.7 | 0.24 | 1.08 |
| BRUE | 9.91 | 9.4 | 0.22 | 0.27 | 9.9 | 0.29 | 2.59 |
| BRI | 8.03 | 15.6 | 0.31 | 0.46 | 16.4 | 0.40 | 1.54 |
| POSC | 6.24 | 8.6 | 0.33 | 0.43 | 9.4 | 0.18 | 1.15 |
| AAR | 2.16 | 29.8 | 0.27 | 0.76 | 30.9 | 0.24 | 1.37 |
| RAM | 4.38 | 12.2 | 0.16 | 0.23 | 12.6 | 0.08 | 1.78 |
| GLA | 6.93 | 16.1 | 0.32 | 0.35 | 16.8 | 0.18 | 38.1 |
| OBE | 6.91 | 5.2 | 0.06 | 0.14 | 5.4 | 0.25 | 0.98 |
| UETE | 3.87 | 20.4 | 0.26 | 0.33 | 21.0 | 0.36 | 1.56 |
| BUCH | 8.28 | 5.2 | 0.07 | 0.18 | 5.4 | 0.69 | 2.07 |
| REIN | 2.88 | 9.3 | 0.16 | 0.27 | 9.8 | 0.15 | 0.58 |
| OPF | 4.07 | 22.8 | 0.30 | 0.47 | 23.6 | 0.36 | 3.74 |
| DUE | 2.00 | 14.8 | 0.19 | 0.20 | 15.2 | 0.23 | 6.98 |
| HER | 1.18 | 177 | 0.24 | 0.21 | 177 | 0.57 | 1.63 |
| SAM | 2.71 | 21.6 | 0.16 | 0.29 | 22.1 | 0.08 | 1.07 |
| LAU | 0.59 | 164 | 0.28 | 0.47 | 164 | 0.04 | 1.30 |
| BIO | 7.91 | 10.9 | 0.18 | 0.25 | 11.3 | 1.97 | 7.09 |
| WEIN | 4.82 | 7.6 | 0.21 | 0.21 | 8.0 | 0.28 | 0.85 |
| MUEN | 3.50 | 11.1 | 0.15 | 0.22 | 11.5 | 0.25 | 0.78 |
| BAS | 0.17 | 153 | 0.32 | 0.52 | 154 | 0.36 | 0.92 |
| RAN | 1.65 | 14.1 | 0.15 | 0.20 | 14.4 | 0.32 | 1.41 |
| BIRS | 2.05 | 29.5 | 0.19 | 0.22 | 29.9 | 0.10 | 1.36 |
| DAVG | 1.05 | 36.2 | 0.19 | 0.35 | 36.8 | 0.04 | 0.61 |
| VISP | 0.20 | 8.6 | 0.07 | 0.19 | 8.9 | 0 | 0.19 |
| POR | 6.76 | 4.3 | 0.10 | 0.21 | 4.6 | 1.30 | 1.16 |
| ZAL | 0.28 | 74.4 | 0.11 | 0.20 | 74.7 | 0.10 | 0.53 |
| Maximum | 18.5 | 362 | 0.34 | 0.76 | 363 (36.8*) | 1.97 (1.97**) | 103 (7.09***) |
| Minimum | 0.17 | 4.3 | 0.06 | 0.14 | 4.6 (4.6*) | 0 (0.02**) | 0.19 (0.19***) |
| Average | 4.55 | 39.1 | 0.21 | 0.32 | 39.7 (16.0*) | 0.37 (0.38**) | 5.61 (1.78***) |
| Median | 3.69 | 14.4 | 0.20 | 0.28 | 14.8 (13.7*) | 0.26 (0.26**) | 1.34 (1.31***) |

* calculated after excluding WWTPs that exceeded the suggested limit value of $50 \mu\text{g g}^{-1}$ d.w. (n = 5); ** calculated after excluding the WWTP VISP, where NP1EC was not detected;

*** calculated after excluding the two extreme outliers (the WWTPs NEU and GLA)

Table S 12. Per capita loads of NP, NPEO (EO1-3), NP1EC, and *tert*-OP.

| ID | NP per capita mg d ⁻¹ | NP1EO per capita mg d ⁻¹ | NP2EO per capita mg d ⁻¹ | NP3EO per capita mg d ⁻¹ | Sum NPEO per capita mg d ⁻¹ | NP1EC per capita mg d ⁻¹ | <i>tert</i> -OP per capita mg d ⁻¹ |
|----------------|-------------------------------------|--|--|--|---|--|--|
| BUEH | 1.04 | 1.33 | 0.010 | 0.015 | 1.35 | 0.09 | 0.12 |
| NEU | 0.62 | 1.22 | 0.030 | 0.028 | 1.27 | 0.02 | 9.58 |
| CHDF | 0.82 | 0.33 | 0.008 | 0.015 | 0.35 | 0.03 | 0.26 |
| VEV | 0.22 | 3.48 | 0.036 | 0.056 | 3.57 | 0.04 | 0.26 |
| UNT | 0.44 | 0.58 | 0.010 | 0.017 | 0.60 | 0.03 | 0.05 |
| WIN | 0.40 | 1.24 | 0.023 | 0.033 | 1.30 | 0.03 | 0.08 |
| ZOF | 0.10 | 85.0 | 0.081 | 0.159 | 85.3 | 0.01 | 0.17 |
| VERN | 0.16 | 0.74 | 0.018 | 0.019 | 0.78 | 0.01 | 0.07 |
| FEH | 0.20 | 1.13 | 0.007 | 0.015 | 1.15 | 0.01 | 0.07 |
| ZUE | 0.27 | 1.09 | 0.013 | 0.021 | 1.12 | 0.03 | 0.16 |
| CHUR | 0.35 | 3.21 | 0.031 | 0.042 | 3.28 | 0.03 | 0.15 |
| BRUE | 1.16 | 1.10 | 0.026 | 0.031 | 1.16 | 0.03 | 0.30 |
| BRI | 0.96 | 1.86 | 0.037 | 0.055 | 1.95 | 0.05 | 0.18 |
| POSC | 0.92 | 1.28 | 0.048 | 0.063 | 1.39 | 0.03 | 0.17 |
| AAR | 0.19 | 2.66 | 0.024 | 0.068 | 2.75 | 0.02 | 0.12 |
| RAM | 0.23 | 0.64 | 0.008 | 0.012 | 0.66 | 0.00 | 0.09 |
| GLA | 0.54 | 1.26 | 0.025 | 0.028 | 1.31 | 0.01 | 2.99 |
| OBE | 0.58 | 0.43 | 0.005 | 0.012 | 0.45 | 0.02 | 0.08 |
| UETE | 0.16 | 0.87 | 0.011 | 0.014 | 0.89 | 0.02 | 0.07 |
| BUCH | 0.77 | 0.48 | 0.006 | 0.016 | 0.50 | 0.06 | 0.19 |
| REIN | 0.29 | 0.94 | 0.017 | 0.028 | 0.98 | 0.02 | 0.06 |
| OPF | 0.32 | 1.76 | 0.024 | 0.037 | 1.82 | 0.03 | 0.29 |
| DUE | 0.14 | 1.04 | 0.013 | 0.014 | 1.07 | 0.02 | 0.49 |
| HER | 0.01 | 1.64 | 0.002 | 0.002 | 1.64 | 0.01 | 0.02 |
| SAM | 0.49 | 3.94 | 0.028 | 0.053 | 4.02 | 0.01 | 0.19 |
| LAU | 0.06 | 16.1 | 0.027 | 0.046 | 16.1 | 0.004 | 0.13 |
| BIO | 0.84 | 1.16 | 0.019 | 0.027 | 1.21 | 0.21 | 0.76 |
| WEIN | 0.36 | 0.56 | 0.016 | 0.015 | 0.60 | 0.02 | 0.06 |
| MUEN | 0.17 | 0.55 | 0.008 | 0.011 | 0.57 | 0.01 | 0.04 |
| BAS | 0.02 | 20.7 | 0.043 | 0.071 | 20.8 | 0.05 | 0.12 |
| RAN | 0.24 | 2.07 | 0.022 | 0.030 | 2.12 | 0.05 | 0.21 |
| BIRS | 0.13 | 1.93 | 0.013 | 0.015 | 1.96 | 0.01 | 0.09 |
| DAVG | 0.11 | 3.17 | 0.020 | 0.036 | 3.77 | 0.004 | 0.06 |
| VISP | 0.29 | 12.4 | 0.094 | 0.269 | 12.7 | 0 | 0.27 |
| POR | 0.63 | 0.40 | 0.009 | 0.019 | 0.43 | 0.12 | 0.11 |
| ZAL | 0.02 | 6.07 | 0.009 | 0.016 | 6.10 | 0.01 | 0.04 |
| Maximum | 1.16 | 85.0 | 0.094 | 0.269 | 85.3 (12.7*) | 0.21 (0.21**) | 9.58 (0.76***) |
| Minimum | 0.01 | 0.33 | 0.002 | 0.002 | 0.35 (0.35*) | 0 (0.003**) | 0.02 (0.02***) |
| Average | 0.40 | 5.14 | 0.023 | 0.039 | 5.20 (1.84*) | 0.03 (0.03**) | 0.50 (0.16***) |
| Median | 0.29 | 1.25 | 0.018 | 0.027 | 1.31 (1.21*) | 0.02 (0.02**) | 0.13 (0.12***) |

* calculated after excluding WWTPs that exceeded the suggested limit value of 50 µg g⁻¹ d.w. (n = 5); ** calculated after excluding the WWTP VISP, where NP1EC was not detected;

*** calculated after excluding the two extreme outliers (the WWTPs NEU and GLA)

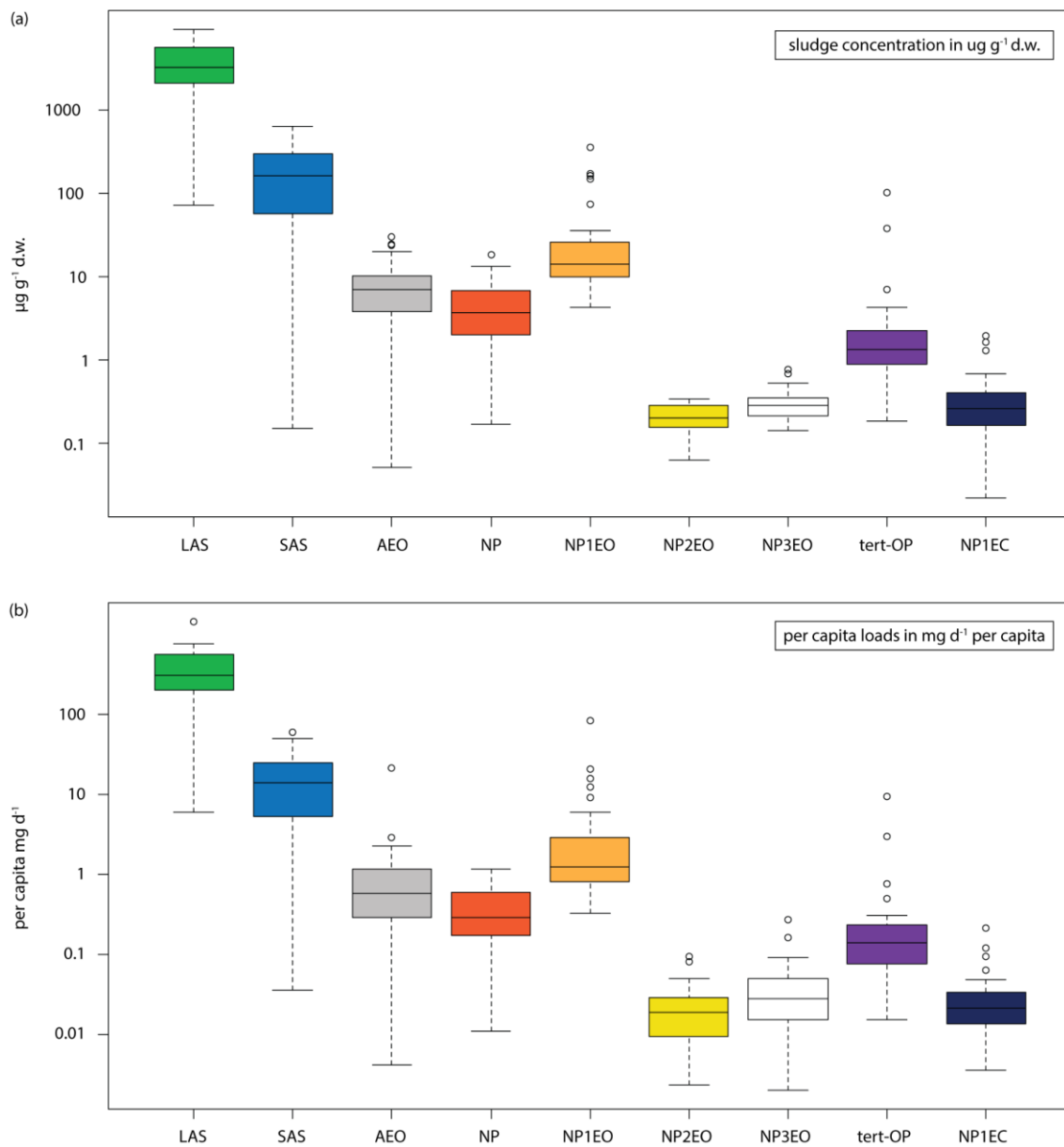


Figure S 2. Summarizing boxplots for levels of target analytes in all sampled WWTPs for (a) sludge concentrations ($\mu\text{g g}^{-1} \text{d.w.}$) and (b) per capita loads ($\text{mg d}^{-1} \text{per capita}$).

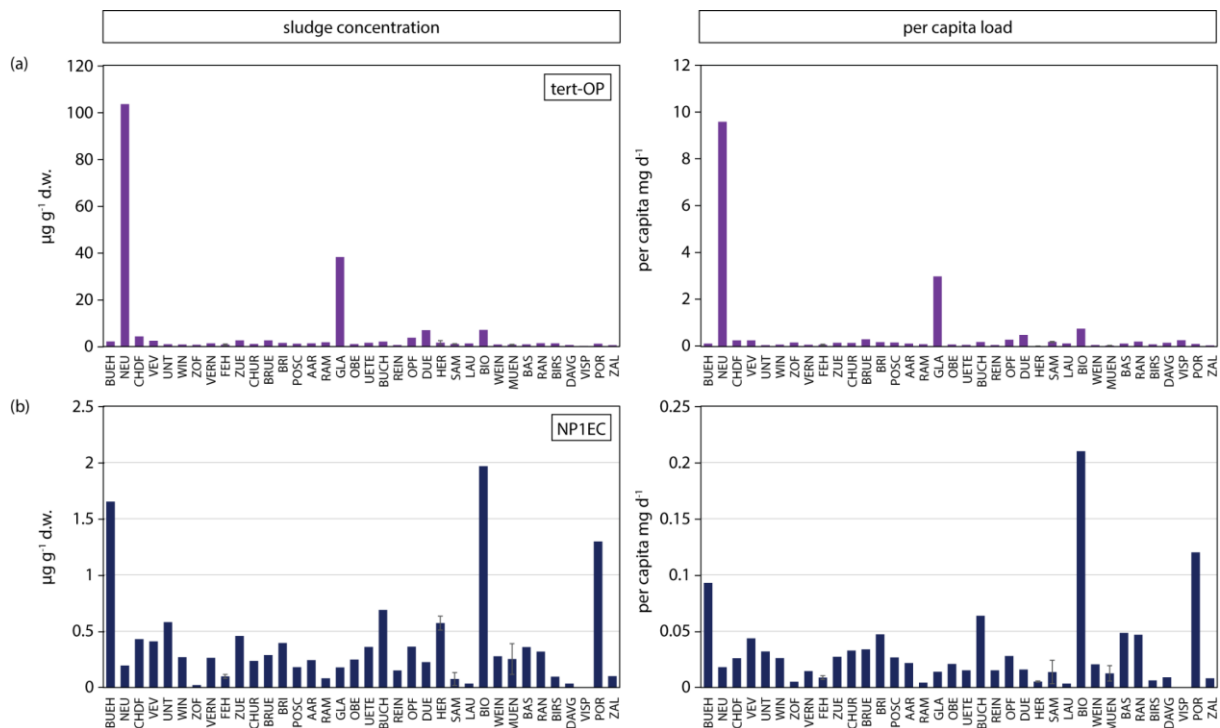


Figure S 3. Concentration levels of (a) *tert*-OP and (b) NP1EC in sampled WWTPs. Right panels show actual sludge concentrations (sludge dry weight, d.w.) whereas left panels show daily per capita loads in sludge. Range bars are depicted for the WWTPs that provided more than one sludge sample.

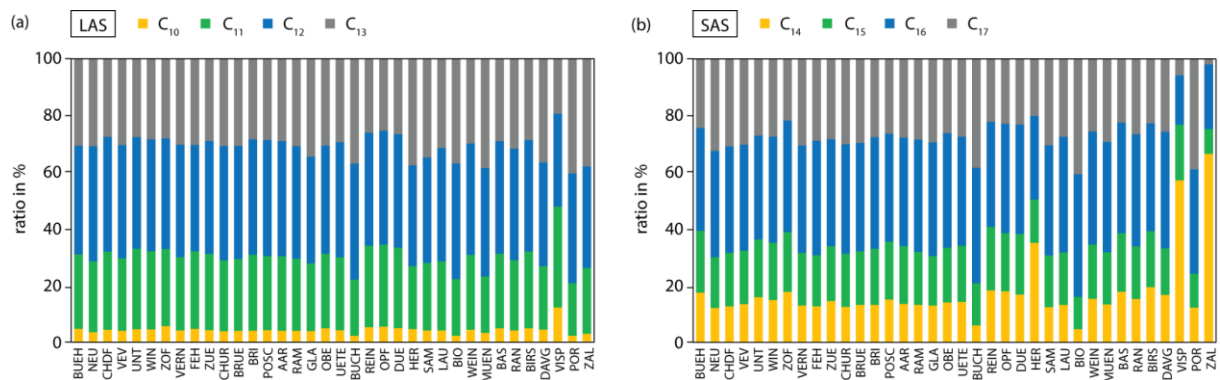


Figure S 4. Carbon chain length distribution of the homologues for (a) LAS and (b) SAS.

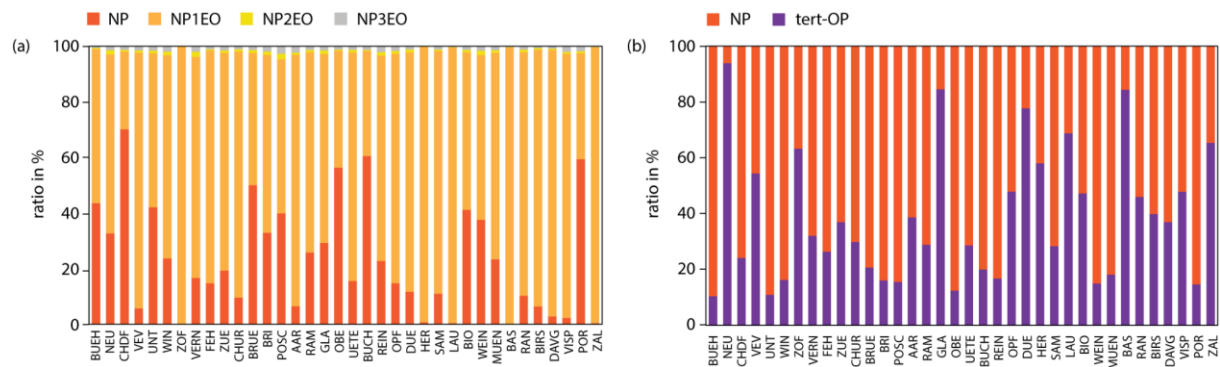


Figure S 5. Distribution of (a) NP and NPEO levels, and (b) NP and *tert*-OP levels in sludge.

S6. AEO homologue data

Table S 13. Average chain lengths of AEO isomers for linear and branched homologues, as well as for the sum of them. The average AEO chain length was derived for each individual sample. These average values were then used to derive the values of the table.

| | Linear | Branched | Sum |
|---------|------------|------------|------------|
| Average | 14.9 ± 2.0 | 16.2 ± 1.1 | 16.1 ± 1.1 |
| Median | 14.1 | 16.7 | 16.7 |
| Maximum | 18.0 | 17.4 | 18.0 |
| Minimum | 12.0 | 13.5 | 13.5 |

Table S 14. Average ethoxymer units of AEO isomers for linear homologues. The average AEO ethoxymer unit was derived for each individual sample. These average values were then used to derive the values of the table.

| Linear | Average | Median | Maximum | Minimum |
|-----------------|--------------|--------|---------|---------|
| C ₁₂ | 10.3 ± 2.4 | 9.5 | 13.5 | 6.9 |
| C ₁₃ | 8.0 ± 2.2 | 7.8 | 12.0 | 3.0 |
| C ₁₄ | 7.6 ± 1.4 | 7.8 | 9.0 | 5.2 |
| C ₁₅ | 7.0 ± 0.9 | 7.0 | 8.4 | 6.0 |
| C ₁₆ | 6.3 ± 0.9 | 6.0 | 7.5 | 5.3 |
| C ₁₇ | Not detected | - | - | - |
| C ₁₈ | 7.5 ± 1.2 | 7.6 | 9.8 | 5.5 |
| All | 8.2 ± 1.9 | 8.0 | 13.0 | 3.0 |

Table S 15. Average ethoxymer units of AEO isomers for branched homologues. The average AEO ethoxymer unit was derived for each individual sample. These average values were then used to derive the values of the table.

| Branched | Average | Median | Maximum | Minimum |
|-----------------|-----------|--------|---------|---------|
| C ₁₂ | 8.9 ± 1.2 | 8.6 | 11.0 | 7.5 |
| C ₁₃ | 9.0 ± 0.7 | 8.7 | 10.3 | 7.9 |
| C ₁₄ | 6.2 ± 1.6 | 6.9 | 8.0 | 4.0 |
| C ₁₅ | 6.3 ± 0.8 | 6.1 | 8.0 | 4.2 |
| C ₁₆ | 7.4 ± 0.9 | 7.5 | 9.1 | 5.0 |
| C ₁₇ | 7.5 ± 0.8 | 7.5 | 9.0 | 4.0 |
| C ₁₈ | 7.4 ± 1.0 | 7.6 | 8.7 | 5.0 |
| All | 7.7 ± 0.6 | 7.7 | 9.0 | 6.6 |

Table S 16. Average ethoxymer units of AEO isomers for the sum of linear and branched homologues. The average AEO ethoxymer unit was derived for each individual sample. These average values were then used to derive the values of the table.

| Sum | Average | Median | Maximum | Minimum |
|-----------------|-----------|--------|---------|---------|
| C ₁₂ | 9.5 ± 1.9 | 8.8 | 13.5 | 6.9 |
| C ₁₃ | 8.3 ± 1.2 | 8.4 | 10.5 | 4.7 |
| C ₁₄ | 6.8 ± 1.7 | 7.5 | 9.0 | 4.0 |
| C ₁₅ | 6.3 ± 0.8 | 6.1 | 8.1 | 4.2 |
| C ₁₆ | 7.3 ± 0.9 | 7.4 | 9.1 | 5.0 |
| C ₁₇ | 7.5 ± 0.8 | 7.5 | 9.0 | 4.0 |
| C ₁₈ | 7.5 ± 1.0 | 7.6 | 9.3 | 5.0 |
| All | 7.7 ± 0.7 | 7.7 | 9.4 | 6.6 |

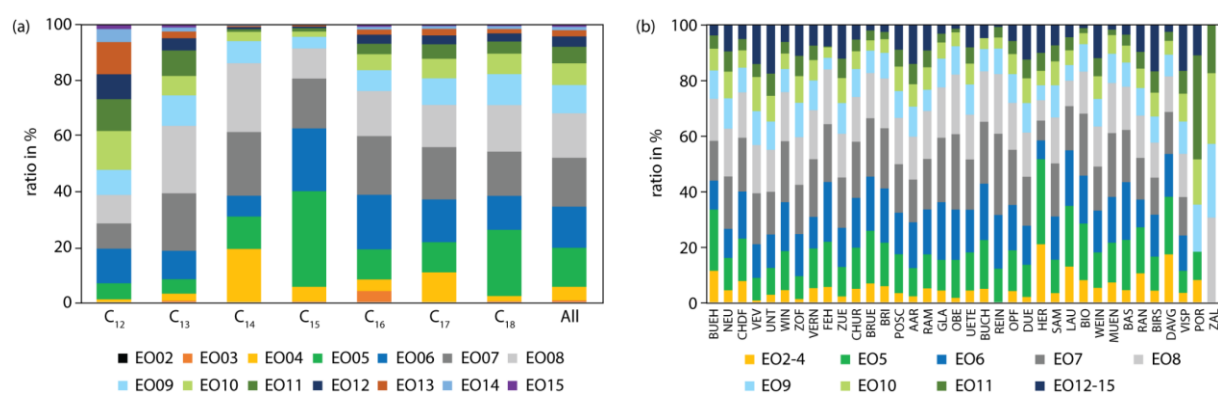


Figure S 6. Relative abundances of different ethoxymer units. (a) Ethoxymer distributions for homologues of different chain lengths. The last bar displays the weighted average of ethoxymer units of all chain length homologues. (b) The average ethoxymer distribution is comparable between most WWTPs.

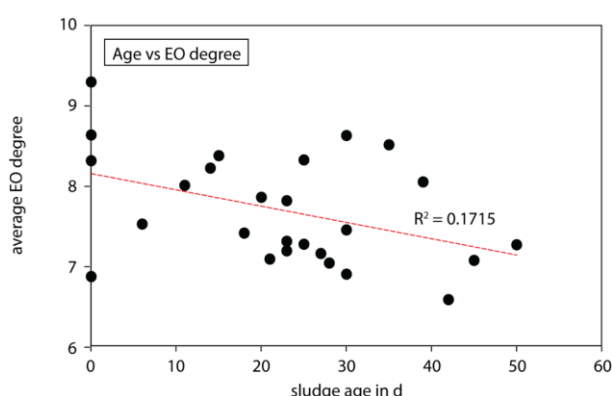


Figure S 7. No apparent relation is observed between the sludge age and the average ethoxylation degree of AEO homologues present in the sludge.

S7. Data correlations

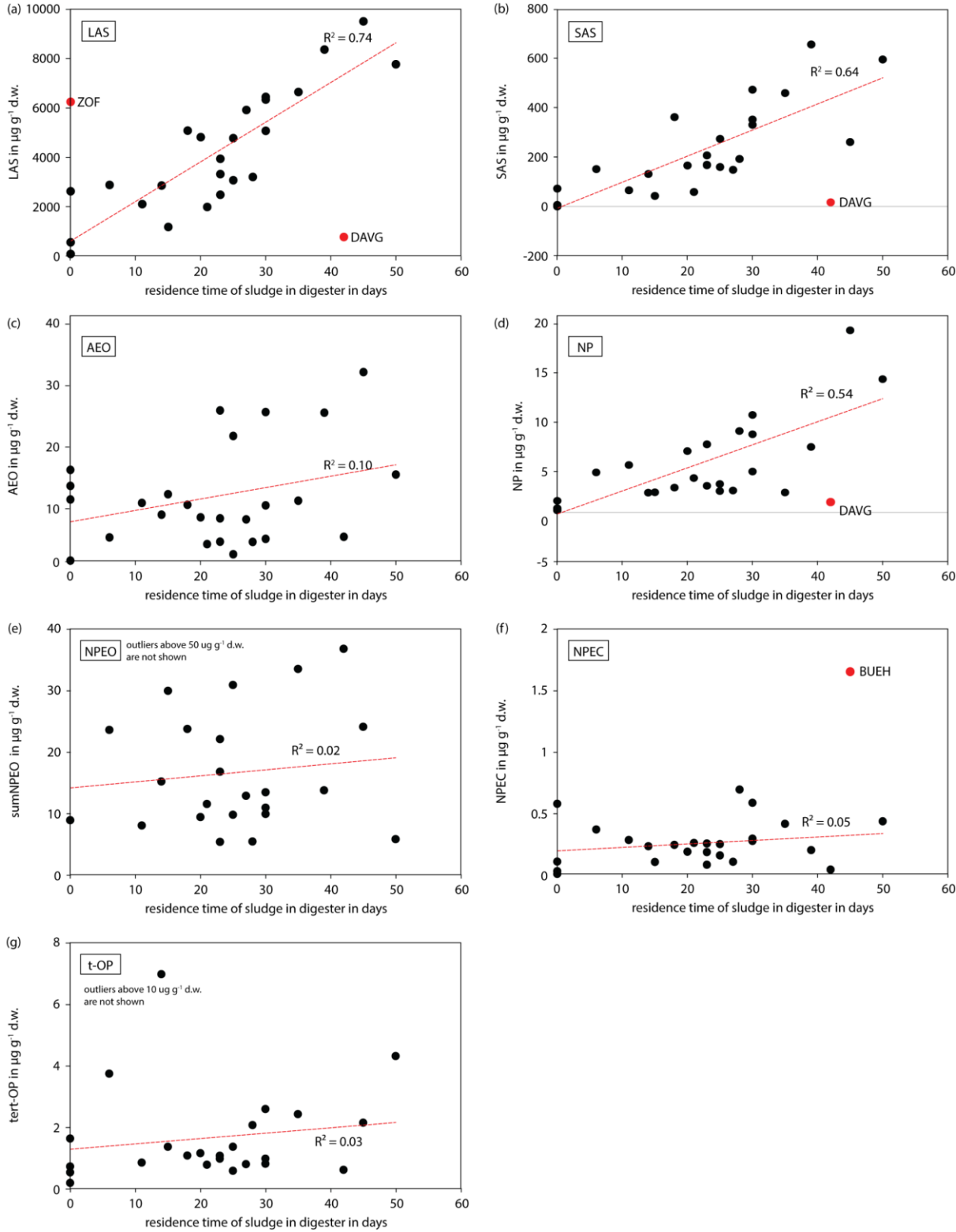


Figure S 8. Correlations between surfactant levels and residence time in the digester (anaerobic). A correlation is evident between sludge age and (a) LAS levels, (b) SAS levels, and (d) NP levels. No correlation is apparent between sludge age and (c) AEO levels, (e) sum NPEO levels, (f) NPEC levels, and (g) *tert*-OP levels. Visually outlying data points that were not considered for the linear regression are highlighted in red.

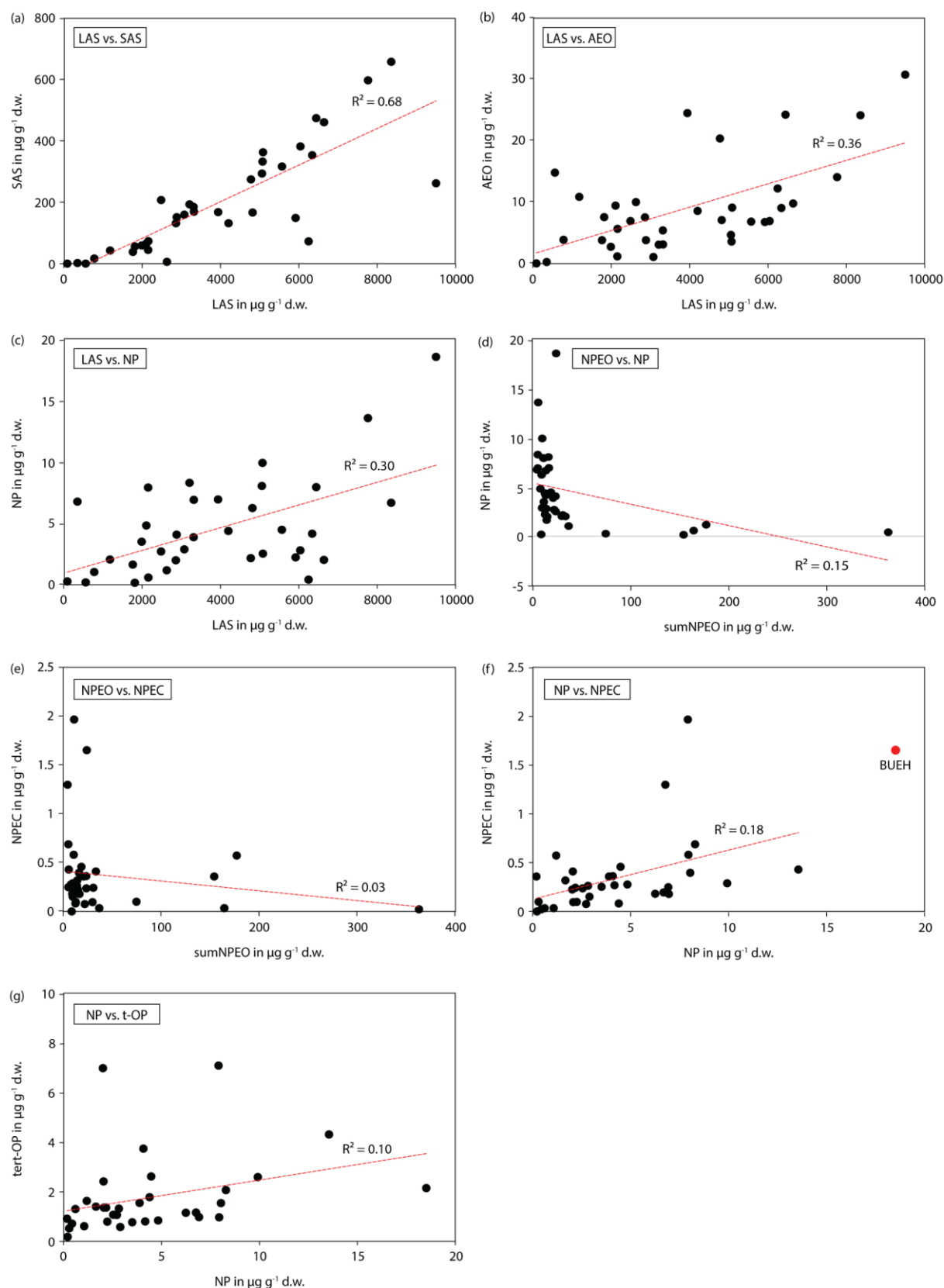


Figure S 9. Correlations between surfactant classes. Levels of surfactants are plotted against each other. A correlation is evident between LAS and (a) SAS levels. A trend toward a linear relationship is evident between (b) LAS levels and AEO levels and (c) (b) LAS levels and NP levels, but with higher dispersion than for LAS vs. SAS. Outlying data points that were not considered for the linear regression are highlighted in red.

S8. References

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