- 1 Effect of drinking water chlorination on the intestinal flora and resistomes of Bangladeshi
- 2 children
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## 31 Abstract

32 Healthy development of the gut microbiome provides long-term health benefits. Children raised 33 in countries with high infectious disease burdens are frequently exposed to antibiotics and 34 diarrheal pathogens, which perturb gut microbiome assembly. A recent cluster-randomized trial 35 in Dhaka, Bangladesh, found that automated water chlorination of shared taps effectively 36 reduced child diarrhea and antibiotic use. Here, we leveraged stool samples collected one year 37 after this intervention to examine differences between treatment and control children's gut 38 microbiota. Water chlorination was associated with increased abundance of several bacterial 39 genera previously linked to improved gut health; however, we observed no effects on the overall 40 richness or diversity of taxa. Several clinically relevant ARGs were relatively more abundant in 41 the gut microbiome of treatment children, possibly due to increases in Enterobacteriaceae. 42 While further studies on the long term health impacts of drinking chlorinated water would be 43 valuable, we conclude that access to chlorinated water did not substantially impact child gut 44 microbiome development in this setting, supporting the use of chlorination to increase global 45 access to safe drinking water.

46

48 **Main** 

Normal assembly of the early-life gut microbiome is critical for human health. The gut microbiome is seeded during birth and stabilizes to an adult-like configuration by the third year of life.<sup>1</sup> The progressive, unperturbed colonization of the intestinal tract during this time window is likely essential to the establishment and maturation of multiple developmental pathways related to metabolism, allergy development, weight gain, disease susceptibility, and mental health.<sup>2–5</sup>

55 Children raised in low- and middle-income countries (LMICs) are at high risk of early-life 56 environmental insults that might disrupt optimal gut microbiome development. Due to poor 57 sanitation and lack of access to clean drinking water, children living in poverty are frequently exposed to enteric pathogens.<sup>6</sup> Pathogen establishment and proliferation in the intestinal tract 58 59 can perturb the normal gut microbiome by triggering local and systemic inflammation.<sup>7,8</sup> In areas 60 with a high infectious disease burden, children also frequently consume antibiotics. In urban 61 Bangladesh, children younger than two years are treated with antibiotics at a rate more than five-times higher than that among similarly aged children in the United States.<sup>9,10</sup> Frequent 62 63 antibiotic use early in life diminishes gut microbiota diversity, enriches for antibiotic resistance genes (ARGs), and reduces microbiome richness while increasing variability.<sup>11,12</sup> In the United 64 States, these perturbations have been linked to increased risk of multiple childhood-onset health 65 66 disorders, including asthma, allergy rhinitis, and attention deficit hyperactivity disorder.<sup>5</sup>

Water chlorination is a promising strategy for reducing exposure to enteric pathogens and associated antibiotic use among young children in LMICs,<sup>13</sup> with potential benefits for the gut microbiome and long-term health. Chlorination inactivates many microorganisms present in water and reduces recontamination during transport and storage.<sup>14</sup> By reducing children's exposures to pathogens, chlorination could prevent the early establishment and proliferation of pathogens in the gut and the subsequent use of antibiotics. However, water chlorination could indirectly affect the developing gut microbiome in other ways. Chlorination does not inactivate all

microorganisms present in drinking water and could systematically impact the types of
waterborne organisms that children are exposed to.<sup>15</sup> For example, the introduction of chlorine
into drinking water systems has been associated with changes in biofilm communities,<sup>16</sup> the
relative abundance of specific bacterial genera in water,<sup>17</sup> and the abundance of ARGs in water
systems,<sup>18</sup> with the potential for ARGs that function as efflux pumps to be enriched.<sup>19</sup> Further,
ingested trace chlorine residuals or chlorine disinfection by-products could perturb the gut
environment in ways that are not fully understood.<sup>20–22</sup>

Members of our team recently conducted a blinded, placebo-controlled cluster-81 82 randomized trial of passive (automated) water chlorination devices installed at shared taps in 83 urban Bangladesh, which disinfected all water used for domestic purposes including drinking, cooking, personal hygiene, and cleaning.<sup>13</sup> The passive water chlorination intervention reduced 84 85 caregiver-reported child diarrhea in the past 7 days by 23% and caregiver-reported antibiotic 86 use in the past two months by 7%, relative to controls. The objectives of the present study were 87 to examine the impacts of water chlorination on children's gut microbiomes, including the 88 resistance genes and bacterial pathogens that they harbored, across different phases of gut 89 microbiome development.

90

## 91 Results

## 92 Child and stool sample characteristics

We examined fecal metagenomes from 130 children from the control (n=64) and treatment
groups (n=66). Samples included in our final analysis were balanced between two study sites
(Dhaka Uddan and Tongi) and three pre-specified age strata (6-14 months, 15-30 months, 31
months and older) corresponding to distinct phases of gut microbiome development.<sup>1</sup>
Characteristics known to impact the early-life gut microbiome were evenly distributed
between the treatment and control groups, including child age,<sup>1</sup> breastfeeding status,<sup>1</sup> recent
diarrhea,<sup>23</sup> and recent antibiotic use (**Table 1**).<sup>4,5</sup> Children were exposed to chlorine (treatment)

and Vitamin C (active control) doses for an average of 10.5 months (range due to open cohort
study design = 1.7 to 14.4 months). Most children (89%) were exposed for at least 6 months.

102 We achieved an average sequencing depth of approximately 6 Gb per sample.

103

## 104 <u>Water chlorination impacted the relative abundance of several human enterobacteria</u>

105 Exposure to chlorinated water significantly impacted the relative abundance of multiple bacterial 106 genera in children's guts, relative to control children (Figure 1A), when adjusting for age and 107 study site. Consistent with previous studies,<sup>1</sup> we observed considerable differences in gut 108 microbiome composition by age group (Figure 2A). We estimated treatment coefficients 109 describing the additive change in the logit-transformed relative abundance of bacterial genera between treatment and control children;<sup>24</sup> positive treatment coefficient values indicate elevated 110 111 levels among children receiving chlorinated water relative to controls. The treatment coefficient 112 generally approximates the log fold change (Supplementary Figure 1). Overall, several 113 bacterial genera that are frequently reported to colonize humans were significantly more 114 abundant among treatment children (fdr-adjusted p-value <0.05), including Akkermansia (treat. 115 coef: 2.4, 95% confidence interval (CI): 1.9, 3.0), *Escherichia* (treat. coef: 1.11, 95% CI: 0.7, 116 1.6), Flavonifractor (treat. coef: 0.89, 95% CI: 0.5, 1.3), and Phascolarctobacterium (treat. coef: 117 2.1, 95% CI: 1.5, 2.7) (Supplementary Table 1). Each of these genera comprised at least 0.1% 118 of bacterial reads in children's fecal metagenomes, on average, with the exception of 119 Escherichia, which comprised 4% of bacterial reads.

The effects of water chlorination on the relative abundance of gut bacteria were modified by age (**Figure 1D-E**). Among the oldest children (age 31-61 months), there were far fewer differentially abundant genera between treatment and controls (*i.e.*, 1 genera compared to 19 differentially abundant genera in children aged 15-30 months and 34 among children aged 6-14 months) (**Supplementary Table 1**). A substantial portion of metagenomic reads could not be classified to any taxonomy, and the proportion of unclassified reads in each sample increased from an average of 13% at age 6-14 months to 42% at age 31-61 months (Supplementary
Figure 2).

128 Because of the open cohort study design, study participants were exposed to the 129 intervention for varying durations. To determine whether longer exposure to the intervention, 130 (*i.e.*,  $\geq$ 6 months) was associated with different impacts on the gut microbiome, we conducted a 131 sub-group analysis among children older than 14 months (since age would be correlated to 132 exposure time for children aged 6-14 months). Among all children older than 14 months 133 (n=103), we identified three genera that were significantly more abundant among treatment and 134 controls (Supplementary Table 2). The direction of the effect was maintained for each of these genera among the subset exposed for at least six months (n=91), although only one effect 135 136 estimate remained significant.

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Water chlorination impacted gut microbial richness, but only among children in the <u>"transitional</u>
 phase" of gut microbiome development

140 Estimated genera richness was fairly constant across each age stratum, and we did not observe 141 any association between richness and treatment status among the youngest (6-14 months) or 142 the oldest (31 months and older) children. However, among children aged 15-30 months, water 143 chlorination was associated with lower estimated richness on average (Figure 1B). Shannon 144 diversity was lowest among the youngest children (age 6-14 months) and progressively higher 145 among each successive age stratum, a trend that is well-described.<sup>2</sup> We observed no 146 association between Shannon diversity and treatment status within any age stratum (Figure 147 1C).

We identified clusters of similar gut metagenomes by calculating pairwise distances between each sample. Among all children, metagenome clusters were primarily defined by age group (*p*<0.001; two-sided chi-square test), rather than exposure to chlorinated water (p=0.09)

- (Figure 2B). Age group stratified analyses revealed treatment status was associated with
   clustering only among children aged 15-30 months (*p*=0.015; Figure 2C-E).
- 153

154 Water chlorination was not associated with the occurrence of specific bacterial pathogens 155 We observed no impacts of automated water chlorination on the occurrence of several bacterial pathogens that can cause enteric infections in children in Bangladesh,<sup>25</sup> after adjusting for 156 157 child's age and study site (Table 2). Using a qualitative multiplex pathogen assay, among the 158 children selected for metagenomic sequencing for whom raw stool aliquots were available for 159 analysis (125/130), we found that 75% of children harbored at least one bacterial pathogen, and 160 children harbored 1.3 bacterial pathogens on average (SD: 1.0 pathogens). Campylobacter 161 spp., Salmonella spp., Enterotoxigenic E. coli [ETEC], and Shigella spp. were detected in at 162 least 10% of children's stool, while Shiga toxin-producing E. coli [STEC]) and Clostridioides 163 *difficile* (<5%) were rarely detected. We repeated this analysis using the full set of children's 164 stool samples from the original trial (n=527) and with a broader set of 14 gastrointestinal 165 pathogens, including viral and protozoan targets. Despite increased power to detect differences, 166 we did not observe significant reductions or increases in any of these pathogens among 167 children receiving chlorinated water (Supplementary Table 3).

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## 169 ARGs were enriched among children receiving chlorinated water

Study children frequently harbored ARGs in their guts, and resistance to the same antibiotic classes was observed in the treatment and control groups (**Supplementary Figure 3A**). We found that resistance to some antibiotic classes was enriched among treatment children (**Supplementary Figure 4**), including sulfonamides and quinolones (all age strata), trimethoprim (children aged 15-30 months), macrolides (15-30 months), and aminoglycosides (31 months and older) (treatment coefficient range for above drug classes: 0.4-1.5). ARGs

conferring resistance to beta lactam antibiotics were less abundant among treatment children
aged 15-30 months (treat. coeff: -0.84, 95% CI: -1.37, -0.32), relative to controls.

178 We also examined chlorination-induced effects on the occurrence and relative 179 abundance of individual ARGs. Several clinically relevant ARGs were detected among both 180 treatment and control children, including *bla*<sub>CTX-M</sub> alleles (conferring third-generation 181 cephalosporin resistance), mph(A) (azithromycin resistance), and qnrS1 (low-level 182 fluoroquinolone resistance) (Supplementary Figure 5). We observed no significant difference 183 (using a *fdr*-adjusted *p*-value threshold of 0.05) in the presence/absence of any ARG between treatment and control children when controlling for age and study site. We did, however, 184 185 observe differences in the relative abundance of several ARGs (Figure 3A). Among all children 186 receiving chlorinated water we observed a higher relative abundance of mdf(A) and tet(A), 187 (multidrug and tetracycline resistance-conferring efflux pumps, respectively), sul2 (sulfonamide 188 resistance) and aadA5 (streptomycin and spectinomycin resistance). Among specific age strata 189 we observed increases in additional ARGs, including tet(X) (tetracycline resistance) among 190 children aged 6-14 months; erm(X) (cross-resistance to macrolides, lincosamides, and 191 streptogramins), mph(A) and nimE (nitroimidazole resistance) among children aged 15-30 192 months; dfrA17 (trimethoprim resistance) among children aged 31-61 months; and blaTEM 193 (penicillin resistance) among children aged 15 months and older (treatment coefficient range for 194 above ARGs: 0.9-7.3). ARGs that encode efflux pumps were slightly enriched in the treatment 195 group (treat. coef.: 0.3, 95% CI: 0, 0.5) but the direction and significance of this effect was not 196 consistent across age groups. The ant(6')-la gene (aminoglycoside resistance) was less 197 abundant among treatment children aged 15-30 months relative to control children (treat. coeff.: 198 -1.3, 95% CI: -2.2, -0.37).

Exposure to chlorinated water was also associated with estimated ARG richness in the gut (p=0.019),<sup>26,27</sup> when controlling for age and study site. Specifically, we observed higher ARG richness among treatment relative to control children in the 15-30 months and 31-61 months

| 202 | age strata ( <i>p</i> =0.015 and p=0.05, respectively) ( <b>Figure 3B</b> ). We also identified a marginal   |
|-----|--|
| 203 | increase in the Shannon diversity of ARGs among treatment children aged 31-61 months   |
| 204 | relative to controls ( <i>p</i> =0.07), but not among other age strata ( <b>Figure 3C</b> ).   |
| 205 | Antibiotic use in the two months prior to stool collection was most frequently reported  |
| 206 | among the youngest children (17/27) and least commonly reported among children aged 31-61  |
| 207 | months (10/52) (Supplementary Figure 3B). Caretakers rarely reported the types of antibiotics  |
| 208 | used. Reported antibiotic use was not associated with ARG richness or ARG Shannon diversity  |
| 209 | in children's stool ( <i>p</i> =0.30 and <i>p</i> =0.14, respectively, by Wilcoxon signed-rank test).  |
| 210 |  |
| 211 | Differentially abundant resistance genes were significantly correlated with specific bacterial taxa  |
| 212 | We observed statistically significant Spearman's correlations (fdr-adjusted $p$ -value <0.05)  |
| 213 | between the relative abundance of several bacterial families and the $ant(6)$ -1a, $bla_{TEM}$ , $erm(X)$ ,  |
| 214 | <i>mdf</i> ( <i>A</i> ), <i>mph</i> ( <i>A</i> ), <i>sul2</i> , and <i>tet</i> ( <i>A</i> ) ( <b>Figure 3D</b> ) genes, indicating potential origins for these |
| 215 | differentially abundant ARGs. In particular, we observed very strong correlations ( <i>rho</i> >0.9)   |
| 216 | between $mdf(A)$ and Enterobacteriaceae <sup>28,29</sup> and $erm(X)$ and Bifidobacteriaceae, <sup>30,31</sup>   |
| 217 | associations which are well-described. We also observed a strong correlation ( <i>rho</i> =0.66)   |
| 218 | between the relative abundance of <i>mph(A)</i> and <i>Enterobacteriaceae</i> . Of note, <i>Escherichia</i> spp.,  |
| 219 | which belong to the family Enterobacteriaceae, were enriched among children receiving  |
| 220 | chlorinated water. Other significant correlations were observed between bacterial families and   |
| 221 | ARGs that were differentially abundant between treatment and control children, but were  |
| 222 | moderate to weak in strength ( <i>i.e.</i> , <i>rho</i> <0.6).   |
| 223 |  |
| 224 | Discussion   |
| 225 | In this analysis of fecal metagenomes from 130 children who participated in a cluster-   |
| 226 | randomized water chlorination trial in urban Bangladesh, automated water chlorination at the   |
|     |  |

227 point of collection impacted children's gut microbiomes, including the resistance genes they

harbored, though shifts in taxa abundance were generally small in magnitude. Children
receiving chlorinated water harbored a higher abundance of bacterial genera that are often
detected in the human gut, including *Akkermansia* spp., *Flavonifractor* spp., *Phascolarctobacterium* spp., and *Escherichia* spp. Despite these changes, water chlorination
did not impact the overall richness or diversity of children's gut microbiomes and was not
associated with differences between children's gut microbiome communities, except among
children aged 15-30 months. These findings suggest that automated water chlorination, an

effective strategy for reducing child diarrhea and associated antibiotic demand,<sup>13</sup> does not
substantially impact children's developing gut microbiomes.

Children in the treatment group harbored a higher abundance of several bacterial general 237 238 previously linked to improved gut health. Akkermansia was more abundant among treatment 239 children overall, relative to controls, and comprised 0.13% of bacterial reads in children's fecal 240 metagenomes. A. muciniphila is a mucin-degrading gut commensal that improves intestinal barrier integrity<sup>32</sup> and reduces gut inflammation.<sup>33</sup> In addition, *Flavonifractor* was approximately 241 242 twice as abundant among treatment compared to control children and comprised 0.19% of 243 bacterial reads in children's fecal metagenomes. F. plautii metabolize flavonoids, which are 244 naturally-occurring compounds common in fruit and vegetables, and the by-products of this metabolic activity modulate gut inflammation and weight gain.<sup>34,35</sup> However, *F. plautii* have been 245 associated with some disease states (e.g., colorectal cancer, bipolar disorder).<sup>36,37</sup> Both A. 246 247 muciniphila and F. plautii are being explored as probiotic targets.<sup>38–40</sup> Phascolarctobacterium 248 spp. are commonly detected in the human gut<sup>41</sup> and were four times as abundant among 249 treatment compared to control children. Phascolarctobacterium spp. solely metabolize luminal 250 succinate, which is a key growth substate for enteric pathogens like *Clostridioides difficile* and 251 Salmonella serovar Typhimurium.<sup>42,43</sup> By reducing bioavailable succinate, Phascolarctobacterium spp. may strengthen gut colonization resistance.<sup>42</sup> Finally, Escherichia 252

spp. were more abundant in treatment compared to control children. Escherichia spp. include

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commensal strains that contribute to colonization resistance against bacterial pathogens,<sup>44</sup> as 254 well as strains that cause diarrhea or other infections.<sup>45,46</sup> Notably, pathogenic *E. coli* (e.g., 255 256 ETEC, STEC) were no more common among treatment relative to control children by the highly 257 sensitive (>90%) and specific (≥99%) qualitative multiplex assay we used,<sup>47</sup> suggesting that the 258 higher relative abundance of *Escherichia* in the treatment group was not driven by increases in 259 pathogenic E. coli. Overall, the higher relative abundance of several human enterobacteria 260 among treatment children compared to controls suggests that chlorination-mediated impacts on 261 the gut microbiome could support engraftment of commensal strains.

262 Water chlorination was associated with microbiota clustering and diminished genera 263 richness among children aged 15-30 months, but this effect were not observed in younger or 264 older children. In a longitudinal study of children born in Europe and the United States, the 15-265 30 month age range was associated with the "transitional phase" of gut microbiome 266 development, characterized by significant shifts in dominant phyla and Shannon diversity.<sup>1</sup> 267 Further studies are needed to examine why children's gut microbiomes may be affected by 268 water chlorination specifically during this age range and what impacts (if any) this may have on 269 future gut composition. Of note, the impacts of water chlorination on the differential abundance 270 of gut taxa were weakest among children in the oldest age stratum (ages 31 months and older). 271 This could suggest that the microbial gut community is less perturbable by water chlorination 272 later on in life. Longitudinal analyses are needed to capture if the subtle impacts of water 273 chlorination on children's developing gut microbiomes that we describe here have any long-term 274 health effects.

Findings from the parent trial that water chlorination reduced child diarrhea could be related to the microbiome shifts we observed here. The original trial (in which this study was nested) leveraged >4,000 child observations to detect a 23% relative reduction in diarrhea prevalence over a 1-year time period.<sup>13</sup> Among the 130 children included in this study, we observed no differences in the presence of eight diarrheal pathogens between treatment and

280 control children. This could be due to our small sample size, seasonal variability in diarrhea 281 prevalence, or because the presence of enteric pathogens does not necessarily imply a disease 282 state. Instead, growing evidence from high child mortality settings suggests that many enteric 283 pathogens are just as commonly detected among children with no clinical manifestations of diarrhea,<sup>48</sup> suggesting that other factors like immunity, exposure dose, inflammation, or 284 intestinal barrier functionality likely play a role in determining disease onset.<sup>49</sup> Here, we 285 286 observed that several gut commensal bacteria were enriched among children receiving 287 chlorinated water; it is possible that the metabolic by-products of these bacteria, their 288 interactions with other gut microbes, or other unmeasured changes to the gut microbiome environment could have improved gut colonization resistance and immunity more broadly.<sup>50</sup> 289 290 Longitudinal stool and environmental sampling in future water chlorination intervention trials 291 could help elucidate whether reductions in child diarrhea from access to chlorinated drinking 292 water are due to disrupted exposures to human, animal, and environmental-origin pathogens, and/or due to improved gut health conferred by shifts in microbiota.<sup>50,51</sup> 293

294 Because children randomized to the water chlorination intervention in the parent trial experienced a 7% reduction in recent antibiotic use, we hypothesized that water chlorination 295 296 could indirectly reduce the relative abundance of ARGs in children's guts.<sup>11</sup> The water 297 chlorination intervention instead increased the relative abundance of several ARGs in children's 298 guts, which was likely the result of a higher relative abundance of *Enterobacteriaceae* spp. 299 among treatment children versus controls. Enterobacteriaceae spp., like E. coli, frequently harbor mobile ARGs,<sup>52</sup> especially in Bangladesh.<sup>28,29,53,54</sup> Our findings suggest that in settings 300 301 like urban Bangladesh, where mobile ARGs are exceptionally common among commensal bacteria,<sup>28,29,53</sup> water chlorination alone may not be sufficient to reduce the burden of ARGs 302 303 circulating in the community. Instead, other interventions might be necessary to curb the selection and spread of antibiotic resistance,<sup>55</sup> not just in human communities but also in the 304 food animal production sector, where many of the same antibiotics are used.<sup>56</sup> Notably, because 305

we did not use functional genomics to characterize all resistance mechanisms present inchildren's fecal microbiomes, we may have missed effects on novel ARGs.

308 Our findings should be interpreted in the context of this study's limitations. First, due to 309 the open cohort study design children were exposed to the intervention for varying durations, 310 though most children were exposed for at least six months. Duration of exposure was strongly 311 correlated with age, so we did not control for this covariate in our analyses. Second, stool 312 samples were not collected at the time of enrollment, which could have helped inform whether 313 there were preexisting differences in gut community composition between the selected subset of 314 treatment and control children, and the extent to which the intervention shifted gut composition. 315 Baseline survey data in the parent trial indicated that the treatment and control groups were well 316 balanced across a range of socioeconomic and child health variables that could plausibly 317 influence gut community composition.<sup>13</sup> Third, a substantial proportion of children's fecal 318 metagenomic reads could not be classified to any known taxonomy, comprising an average of 319 42% of reads among the oldest children. This may have affected our ability to assess the 320 relative effect of chlorination of children's gut microbiomes, given that our analyses were based 321 solely on the classifiable fraction. The unclassified portion that we observed is comparable to other studies that have included fecal metagenome data from LMICs (42%-68%),<sup>57,58</sup> providing 322 323 further evidence of biases in existing taxonomy databases towards high-income, Western countries.<sup>59</sup> Finally, an important caveat of our results is that control children in the parent trial 324 325 received water dosed with trace amounts of Vitamin C, which could have conferred a nutritional benefit. The impacts of Vitamin C on the gut microbiome are not well-known.<sup>60</sup> One pilot trial of 326 327 healthy adults in Europe<sup>61</sup> found that high doses (500 mg/day) administered directly to the colon 328 increased genera richness and the relative abundance of some bacterial taxa, including 329 Akkermansia, but the setting, study population, and administration route are not comparable to 330 the present study. Given that Vitamin C was dosed at very low levels, we anticipate that any

331 subclinical changes to the gut microbiome would not have been distinguishable from natural332 variation in gut microbiome development among children in this setting.

333

## 334 Conclusion

335 To our knowledge, this is the first study to experimentally evaluate the impact of water 336 chlorination, the most common form of drinking water disinfection worldwide, on children's 337 developing gut microbiomes. Because we leveraged stool samples from a double-blinded, 338 cluster-randomized placebo-controlled automated water chlorination trial, water chlorination can 339 be causally attributed to the outcomes measured here. Our findings suggest that water 340 chlorination does not substantially affect the developing gut microbiomes of children in urban 341 Bangladesh. Specifically, while we observed chlorination-induced shifts in the relative 342 abundance of some bacteria taxa, including beneficial gut commensals, we observed no effects 343 on overall gut genera richness or diversity. However, long-term studies may be needed to 344 confirm that these types of subtle changes do not affect health later in life. Overall, the benefits 345 of automated water chlorination with regards to preventing child diarrhea, reducing antibiotic 346 use, and protecting child health appear to outweigh any potential changes to gut microbiome 347 development in this setting.

348

# 349 Materials and Methods

## 350 Stool collection

Stool samples were collected from children participating in a double-blinded, cluster-randomized automated water chlorination trial implemented from July 2015-December 2016 in two lowincome communities in urban Bangladesh: Tongi, a community outside Dhaka city, and Dhaka Uddan, a community within Dhaka city.<sup>13</sup> In brief, shared water taps that served as the primary source of drinking water for children younger than five years old were identified in each community, then randomized to either the treatment (n=50) or control (n=50) arms. Although

357 water taps were selected based on reliance as a primary drinking water source, water from these taps was used for all purposes, including bathing, cleaning, and washing clothes. Over 14 358 359 months, treatment water points were automatically dosed with chlorine using a passive water 360 chlorination device, achieving a mean chlorine residual of 0.37 ppm, while control water points 361 were dosed with trace amounts of Vitamin C (active control).<sup>13</sup> Vitamin supplementation was 362 included as an active control in order to improve acceptability to study participants and the local 363 human subjects protection board. Vitamin C was specifically chosen by study investigators 364 because tablets compatible with the dosing devices could be locally acquired.

365 Survey data were collected approximately every two months from an open cohort of 366 children younger than five years old, of which over 1000 were enrolled at baseline. The trial 367 used an open cohort study design given high migration rates in both communities. Thus, new 368 children living in compounds served by either a treatment or control water tap were continuously 369 enrolled at every follow-up survey round. Field workers confirmed at each study visit that the 370 primary source of drinking water for each enrolled child was either a control or treatment pump. 371 Across all survey rounds in the parent trial, less than 4% of households reported using a 372 secondary water source for drinking.<sup>13</sup> Both study participants and researchers processing the 373 samples and performing data analysis were unaware of which households were served by 374 chlorinated taps (double-blinded). This study and the cluster randomized trial in Bangladesh<sup>13</sup> 375 followed the CONSORT checklist for cluster randomized trials (Supplementary Material). The 376 study protocol for the original trial was approved by the International Centre for Diarrhoeal 377 Diseases Research, Bangladesh (icddr,b) scientific and ethical review committees (protocol 378 number 14022) and the human subjects institutional review board at Stanford University 379 (protocol number 30456). Field staff obtained informed written consent from the owner of each 380 water point enrolled and all study participants, including consent for biospecimens to be used for 381 future unplanned analyses.

382 Children's stool samples were collected approximately one year into the trial.

383 Households were provided stool collection kits that included latex gloves and sterile scoops, and 384 were instructed on safe collection procedures prior to handling stool. Following stool production, 385 the child's mother or other caretaker were instructed to immediately transfer a small amount of 386 feces into pre-prepared vials containing 1 mL of RNAlater, which has been demonstrated to be a suitable preservative for fecal samples,<sup>62</sup> and then invert the tube a few times. Field staff 387 transported children's stool samples to the laboratory within 8 hours. Field staff conducted up to 388 389 three follow-up visits per household to retrieve specimens. Stool aliquots stored in RNAlater 390 were frozen at -80°C upon arrival at the International Centre for Diarrheal Disease Research, 391 Bangladesh (icddr,b) in Dhaka and remained frozen during storage and transport on dry ice to 392 Tufts University in the United States for DNA extraction.

393

## 394 Sample selection for metagenomic sequencing

395 We performed short-read metagenomic sequencing on RNAlater-preserved stool specimens 396 collected from children older than 6 months in the control and treatment groups. Sample 397 selection was done by stratified random sampling of archived child stool samples to balance 398 across treatment and control groups, study community (*i.e.*, Dhaka Uddan versus Tongi), and 399 across three age groups that others have demonstrated to correspond to distinct phases of gut 400 microbiome development (*i.e.*, 6 -14 months, 15-30 months, 31 months and older).<sup>1</sup> Total DNA 401 was extracted from approximately 0.25 g of frozen feces at Tufts University using the QIAamp 402 PowerFecal DNA Kit (Qiagen) according to manufacturer's instructions and quantified using a 403 Qubit 4 fluorometer (Invitrogen). Extraction blanks were included with each batch of extractions; 404 DNA concentrations were below the Qubit level of detection for all. We performed duplicate 405 extractions on two stool samples. DNA extracts were sent to Novogene (UK) Company Limited 406 for short-read, paired-end 150 bp sequencing on an Illumina Novaseq 6000 System using SP4 407 flow cells to achieve 10 Gb per sample.

## 409 <u>Processing metagenomic data</u>

410 Fastg sequences were trimmed and filtered to remove sequencing adaptors and low quality 411 reads using Trimmomatic v0.36. We tabulated the number of raw reads for each sample and 412 excluded from further analysis any sample with fewer reads than two standard deviations below 413 the mean. Human contaminant sequences were removed from each sample by discarding 414 reads that mapped to a non-redundant version of the Genome Reference Consortium Human 415 Build 38 (GRCh38; www.ncbi.nlm.nih.gov) using Bowtie2 v2.2.3. Human sequence-filtered raw 416 reads were deposited in the Sequence Read Archive (SRA; https://www.ncbi.nlm.nih.gov/sra) 417 under the project number PRJNA726052. We randomly chose one extraction duplicate per 418 sample to include in subsequent metagenomic analyses.

419

## 420 Profiling taxonomy and classifying source

Taxonomic assignment of short reads was performed using Kraken2, using its standard built-in
database comprising all complete bacterial, archeal, and viral genomes in NCBI's RefSeq at the
time of build (03 June 2021).<sup>63</sup> We used the Bracken species-level sequence abundance
estimation algorithm to estimate organism abundance at every taxonomic level.<sup>64</sup> For all
analyses, we considered organisms classified as bacteria in NCBI's taxonomy database
(https://ftp.ncbi.nlm.nih.gov/pub/taxonomy/), downloaded 23 November 2021.

427

# 428 Identifying antibiotic resistance genes and calculating abundance

429 We screened for antibiotic resistance genes (ARGs) by mapping short reads to the Resfinder

- 430 database (v. 3.1.1) using the KMA tool.<sup>65</sup> We considered matches with >90% coverage and
- 431 >95% identity to be true hits. To examine the abundance of ARGs across samples, we
- 432 calculated the RPKM (Reads Per Kilobase Million) for each ARG in a sample as the number of
- hits divided by the total number of matched paired-end bacterial reads per million for that

434 sample, then divided by the ARG length in kilobases. To visualize the relative abundance of
435 ARG classes across samples, we summed the RPKM for all ARGs belonging to a class.

436

## 437 <u>Clustering similar metagenomes</u>

438 All-against-all genomic distances were estimated between each metagenome using Mash.<sup>66</sup> We 439 reduced the dimensionality of the resulting matrix using UMAP from the *uwot* package 440 (https://github.com/jlmelville/uwot) and used the MClust R package<sup>67</sup> to classify samples into clusters. The MClust algorithm chooses the number of clusters and the specific clustering model 441 442 (of eight models considered) to maximize the BIC; a minimum of one cluster is considered by default and we did not assign a maximum.<sup>67,68</sup> We repeated this same process within each age 443 444 stratum after subsetting the Mash distance matrix. We used chi-square tests to examine 445 associations between treatment status and cluster classification, both within each age group 446 and overall. All *p*-values are reported in Figure 2.

447

#### 448 Identifying gastrointestinal pathogens

449 Given that rare taxa like gastrointestinal pathogens are difficult to capture using metagenomic 450 sequencing approaches, and that pathogenic E. coli (e.g., ETEC, STEC) may be impossible to 451 distinguish from commensal strains using short-read sequencing data, we used a qualitative 452 multiplex assay with >90% sensitivity and  $\geq$ 99% specificity to determine pathogen occurrence, 453 the Luminex xTAG® Gastrointestinal Pathogen Panel.<sup>47</sup> For children whose stool samples were 454 analyzed using short-read metagenomic sequencing, we examined additional, raw aliguots of 455 their stool specimens for the presence of 8 bacterial pathogens, including 456 Campylobacter; Clostridioides difficile; Enterotoxigenic E. coli (ETEC) LT/ST; Shiga-like toxin-457 producing E. coli (STEC) stx1/stx2; Salmonella spp.; Shigella spp.; Yersinia enterocolitica, and 458 Vibrio cholerae. We also examined all stool specimens collected during the parent trial for a 459 broader set of 14 pathogens, including the aforementioned bacterial pathogens as well as

adenovirus 40/41; norovirus GI/GII; rotavirus A; *Giardia* spp.; *Cryptosporidium* spp.; and *Entamoeba histolytica*. Methods, including DNA and RNA extraction protocols, have been
described in detail elsewhere.<sup>69</sup>

463

464 Comparing extraction duplicates

465 We compared the taxonomic profiles of both sets of extraction duplicates (Supplementary 466 Figure 9) and observed no clear differences in the relative abundance of any bacterial family or 467 genera that comprised at least 1% of bacterial reads, on average, across all of the samples we 468 sequenced (n=132). We observed some discordance in the genera that were identified within 469 each extraction pair (3 discordant genera versus 1552 concordant genera among extraction 470 duplicates for Sample A; 85 discordant genera versus 1063 concordant genera among 471 extraction duplicates for Sample B); however, all discordant genera were of exceptionally low 472 abundance (<0.007%). We randomly chose one extraction duplicate per sample to include in 473 subsequent metagenomic analyses.

474

475 Data Analysis

476 We examined differences in bacterial taxa (primary outcome), ARGs (secondary outcome), and 477 the occurrence of gastrointestinal pathogens (secondary outcome) between treatment and 478 control children using several approaches. Analysis of differences in bacteria taxa and ARGs 479 were not planned as part of the original trial, though differences in the occurrence of 480 gastrointestinal pathogens was pre-specified as a secondary outcome in the original trial's 481 registration and statistical analysis plan (available at: https://osf.io/t98bv/). The original trial was 482 registered with ClinicalTrials.gov, number NCT02606981. Because nearly all children (126/130) 483 lived in different households, each child's fecal metagenome was considered as an independent 484 data point for these analyses.

485 For bacterial taxa, we used beta-binomial regression models that account for variable 486 sequencing depth to identify taxa that were differentially abundant between treatment and control children with the R package *corncob*,<sup>24</sup> while controlling for the child's age and study 487 488 site. This was the primary outcome we examined. The Benjamini-Hochberg method was used to 489 correct for multiple comparisons (default method for *corncob*).<sup>70</sup> Only taxa that were present in 490 at least 20% of samples were included in these analyses. Given frequent discordance in 491 species-level taxonomy assignments among commonly used taxonomic labeling tools for short-492 read sequencing data,<sup>71</sup> we only examined differences at the genera level and above. We 493 filtered differentially abundant taxa that were exclusively driven by influential points from further 494 analyses. Differentially abundant genera and families (as identified by corncob) were visualized using the *metacoder* package.<sup>72</sup> We also calculated two estimates of bacterial alpha diversity for 495 496 each sample: richness, which describes the number of bacterial genera, and the Shannon 497 index. Genera richness was estimated using the *breakaway* package.<sup>27</sup> Shannon diversity was calculated using the *phyloseg* package.<sup>73</sup> Differences between treatment and control children for 498 499 each of these metrics were examined within each age stratum using Wilcoxon signed-rank 500 tests.

501 We used logistic regression models to examine differences in the presence/absence of 502 ARGs between treatment and control children, while controlling for age and study site. Because 503 allelic variants of beta lactamase genes can differ by as few as one single nucleotide 504 polymorphism, we considered these genes as groups (e.g., bla<sub>CTX</sub>, bla<sub>CXA</sub>, bla<sub>TEM</sub>) rather than as 505 individual variants. Resulting p-values were adjusted for multiple comparisons using the 506 Benjamini–Hochberg method. For ARGs that were common among children in this setting 507 (present in at least 20% of samples), we used the R package *corncob* to identify differentially 508 abundant ARGs between treatment and control children, while controlling for age and study site 509 (secondary outcome). We separately used *corncob* to evaluate whether ARGs that function as 510 efflux pumps were differentially abundant between treatment and control children, by creating a

511 composite indicator variable for all efflux pump ARGs in our dataset. We determined the 512 functional annotation of detected ARGs by cross-referencing with the Comprehensive Antibiotic Resistance Database (CARD).<sup>74</sup> For any ARGs not listed in CARD (*i.e.*, *nimE*, *nimJ*, *mefA*, 513 514 tet(O/32/O), tet(O/W), tet(W/32/O), we determined their function by searching the literature. We used *breakaway*'s betta function<sup>26</sup> to examine the effect of chlorination on estimated ARG 515 516 richness while controlling for study site, both overall (analysis also controlled for age) and within 517 each age stratum (i.e., 6 -14 months, 15-30 months, 31 months and older). We examined 518 differences in ARG diversity between treatment and control children using Wilcoxon signed-rank 519 tests.

520 To examine differences in gastrointestinal pathogen burden between treatment and 521 control children, we used Poisson regression models to examine associations between 522 treatment status and the presence of any pathogen that was harbored by at least 5% of 523 children, while controlling for child's age and study site (secondary outcome). The resulting p-524 values were adjusted for multiple comparisons using the Benjamini-Hochberg method. 525 We conducted a Spearman's correlation analysis to investigate how changes in gut 526 taxonomic structure might be driving changes in ARG relative abundance between treatment 527 and control children. Only correlations between ARGs and bacterial families that occurred in at 528 least half of samples were examined. The rho and p-values were calculated using the rcorr 529 function of the Hmisc package (https://github.com/harrelfe/Hmisc/) and resulting p-values were 530 adjusted for multiple comparisons using the Benjamini-Hochberg method.

531 All analyses were conducted in R v. 3.5.0 and R Studio v. 1.1.463. All p-values were 532 two-sided and considered to be statistically significant at the 0.05 level.

533

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535 (<u>pickering@berkeley.edu</u>) and Dr. Tim Julian (tim.Julian@eawag.ch)

### 537 Data Availability

538 All raw reads (human sequences removed) were deposited in the Sequence Read Archive

539 (SRA; https://www.ncbi.nlm.nih.gov/sra) under the project number PRJNA726052. Metadata are

- 540 publicly available at the following link: https://osf.io/wb3pv/.
- 541

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555

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557 The funder had no role in data collection, data analysis, data interpretation, or writing of this 558 report.

559

## 560 <u>Author contribution statement</u>

561 AJP and TRJ conceptualized this follow-up study and obtained funding. AJP, SS, and SPL 562 designed the original trial. AJP, SS, YSC, and JS contributed to data collection in the original

- trial. MCM, LT, LC, and ERF performed stool DNA extractions. YSC and JB contributed to the
- 564 enteric pathogen analysis. CW, AME, MAI, and VFL provided input on methods and data
- 565 interpretation. MLN and VFL contributed to metagenomic data analysis. MLN wrote the first
- 566 draft. All co-authors contributed to writing and editing the manuscript.
- 567

# 568 <u>Tables</u>

Table 1. Characteristics of 130 Bangladeshi children participating in a cluster-randomized automated water chlorination trial at the time of stool sample collection.

571

|  | Control    | Treatment  |
|--|------------|------------|
|  | n=64 (%)   | n=66 (%)   |
| Female   | 40 (63)    | 37 (57)    |
| Age  | <b>、</b>   |            |
| 6-14 months  | 13 (20)    | 14 (21)    |
| 15-30 months   | 25 (41)    | 26 (39)    |
| 31-61 months   | 26 (41)    | 26 (39)    |
| Study site   | <b>、</b>   |            |
| Dhaka Uddan  | 28 (44)    | 28 (42)    |
| Tongi  | 36 (56)    | 38 (58)    |
| Currently receiving any human milk <sup>a,b</sup>    | 32 (50)    | 29 (45)    |
| Experienced fever in past seven days <sup>a</sup>    | 15 (23)    | 9 (14)     |
| Experienced diarrhea in past seven days <sup>a</sup> | 3 (5)      | 3 (5)      |
| Received antibiotics in past two months <sup>a</sup> | 25 (36)    | 23 (35)    |
| Months enrolled in trial (mean, SD)                  | 11.5 (3.0) | 10.8 (3.1) |

<sup>a</sup>Caregiver reported at time of stool collection.

<sup>b</sup>Exclusive breastfeeding was rare at the time of stool sample collection; no children in the

574 control group and only one child in the treatment group (age 14 months) was exclusively 575 breastfed.

575 576

577 Table 2. Detection of 8 bacterial pathogens in the stool of 125 children participating in a cluster-

578 randomized automated water chlorination trial following 10 months of exposure, on average.<sup>a</sup>

|   | Control<br>n=61 (%) | Treatment<br>n=64 (%) | RR (95% CI)       | Adjusted <i>p</i> -<br>value <sup>b</sup> |
|---|---------------------|-----------------------|-------------------|---|
| Campylobacter spp.  | 10 (16)             | 16 (25)               | 1.42 (0.65, 3.25) | 0.81                                      |
| Salmonella spp.   | 15 (25)             | 25 (39)               | 1.56 (0.83, 3.02) | 0.81                                      |
| Enterotoxigenic <i>E. coli</i> (ETEC) LT/ST                   | 22 (36)             | 20 (31)               | 0.87 (0.47, 1.60) | 0.81                                      |
| Shigella spp.   | 19 (31)             | 23 (36)               | 1.17 (0.64, 2.18) | 0.81                                      |
| C. difficile  | 3 (5)               | 3 (5)                 | 0.90 (0.17, 4.90) | 0.90                                      |
| Shiga-like toxin-producing <i>E.</i><br>coli (STEC) stx1/stx2 | 4 (7)               | 0                     |                   |   |
| Vibrio cholerae   | 0                   | 0                     |                   |   |

<sup>a</sup>Raw stool aliquots from 125 of 130 children were available for pathogen analysis.

<sup>b</sup>Adjusted for multiple comparisons using the Benjamini–Hochberg method.

581 *Note*: Relative risk ratios rates (RR) were calculated using Poisson regression models adjusted

582 for child's age and study site. RRs, associated 95% CIs, and adjusted p-values are only

583 presented for pathogens that were detected among at least 5% of samples. 75% of children

harbored at least one bacterial pathogen; 67% of children harbored at least one non-

585 *Escherichia*-related bacterial pathogen.

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6-14 months (n=27)

15-30 months (n=51)

595 Figure 1. Differentially abundant gut taxa among children aged 6-61 months who were cluster randomized to an automated chlorinated water intervention in urban Bangladesh 596 and effects on overall richness and diversity. Panel A) is a differential heat tree depicting the 597 taxonomies of bacterial genera and families that significantly differed in their relative abundance 598 599 between treatment and control children. Genera within differentially abundant families are also 600 depicted. For any given taxonomic level, only taxa that were significantly less abundant (orange) 601 or more abundant (purple) among treatment relative to control children are depicted in color; non-significant taxa are depicted in gray. Panel B) depicts estimated genera richness and Panel 602 603 C) depicts Shannon diversity indices for treatment and control children, stratified by child age. 604 Estimated genera richness differed between treatment and control children aged 15-30 months 605 by the Wilcoxon signed-rank test; Shannon diversity did not significantly differ for any age 606 stratum. For all box plots, center line indicates the median; box limits indicate the upper and 607 lower quartiles; and whiskers indicate 1.5x interquartile range. Panels D) – E) are differential heat trees depicting the taxonomies of bacterial genera and families that significantly differed in 608 609 their relative abundance between treatment and control children aged D) 6-14 months, E) 15-30 610 months, and F) 31-61 months, controlling for study site.

611 *Note*: ns=non-significant. \*indicates p<0.05 by Wilcoxon signed-rank test.



612 613

Figure 2. Effects of automated water chlorination on the taxonomic structure of 614 children's gut microbiomes in urban Bangladesh. Water chlorination impacted the structure of the gut microbiomes of children aged 15-30 months, but not younger or older children. Panel 615 616 A) depicts the average relative abundance of bacterial families with ≥1% mean relative abundance across all samples. Panel B) is a two-dimensional representation of the pairwise 617 618 genomic distances between each sample, as identified by Mash. We observed a marginal 619 association between treatment status and cluster classification by the chi-square test. Panels C-

620 E) depict pairwise distances and resulting clusters when stratified by child age. Treatment status

621 was only associated with cluster classifications among children aged 15-30 months.



623 624

625 Figure 3. Effects of automated water chlorination on the richness, diversity, and relative 626 abundance of antibiotic resistance genes (ARGs) harbored by children's intestinal flora in urban Bangladesh. Panel A) describes ARGs that were differentially abundant between 627 628 treatment and control children, controlling for study site (and child age in the "Overall" panel 629 only). Error bars depict the 95% confidence interval. Positive treatment coefficient values 630 indicate ARGs were more abundant among treatment children relative to controls; negative 631 values indicate ARGs were less abundant. Panel B) depicts the estimated number of unique 632 ARGs (*i.e.*, ARG richness) detected in children's fecal metagenomes, stratified by child age. Estimated ARG richness significantly differed between treatment and control children aged 15-633 634 30 months and 31-61 months using the *betta* function of the R package *breakaway*. Panel C) 635 depicts the Shannon diversity indices for treatment and control children, stratified by child age. 636 There was no statistical association between treatment status and ARG diversity for any of the age stratum by the Wilcoxon signed-rank test. For all box plots, center line indicates the median; 637 box limits indicate the upper and lower quartiles; and whiskers indicate 1.5x interquartile range. 638 639 Panel D) is a heatmap depicting Spearman correlations between ARGs listed in Panel A) that occurred in at least half of samples and bacterial families that occurred in at least half of 640 641 samples. Rho values are only depicted for statistically significant correlations (p<0.05 after 642 adjustment for multiple testing). 643 *Note*: ns=non-significant. \*indicates  $p \le 0.05$ .

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