

Supporting Information

Evaluation of the role of superoxide as chain carrier of ozone decomposition to hydroxyl radicals during ozonation

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S1. Composition of synthetic solutions

Table S1. Composition of synthetic solutions used in this study (buffered with 10 mM phosphate at pH = 8) with the concentrations and the first-order scavenging rate constants. The second-order rate constants for the reaction of methanol, acetate, and *tert*-butanol with $\cdot\text{OH}$ are $9.7 \times 10^8 \text{ M}^{-1} \text{ s}^{-1}$, $7.9 \times 10^7 \text{ M}^{-1} \text{ s}^{-1}$, and $5 \times 10^8 \text{ M}^{-1} \text{ s}^{-1}$, respectively. (Buxton et al., 1988; Wolfenden and Willson, 1982)

Solution	Methanol		Acetate		<i>tert</i> -Butanol		Overall $\cdot\text{OH}$ scavenging rate (s^{-1})
	[C] (M)	$k \cdot_{\text{OH}}[\text{C}]$ (s^{-1})	[C] (M)	$k \cdot_{\text{OH}}[\text{C}]$ (s^{-1})	[C] (M)	$k \cdot_{\text{OH}}[\text{C}]$ (s^{-1})	
1	1.24×10^{-4}	1.2×10^5	0	0	0	0	1.2×10^5
2	0.93×10^{-4}	0.9×10^5	0.38×10^{-3}	0.3×10^5	0	0	1.2×10^5
3	0.62×10^{-4}	0.6×10^5	0.76×10^{-3}	0.6×10^5	0	0	1.2×10^5
4	0.31×10^{-4}	0.3×10^5	1.14×10^{-3}	0.9×10^5	0	0	1.2×10^5
5	0	0	1.52×10^{-3}	1.2×10^5	0	0	1.2×10^5
6	0.93×10^{-4}	0.9×10^5	0	0	0.6×10^{-4}	0.3×10^5	1.2×10^5
7	0.62×10^{-4}	0.6×10^5	0	0	1.2×10^{-4}	0.6×10^5	1.2×10^5
8	0.31×10^{-4}	0.3×10^5	0	0	1.8×10^{-4}	0.9×10^5	1.2×10^5
9	0	0	0	0	2.4×10^{-4}	1.2×10^5	1.2×10^5

S2. Determination of the second-order rate constant for the reaction of $O_2^{\bullet-}$ with acetate

The second-order rate constant for the reaction of $O_2^{\bullet-}$ with acetate ($k_{O_2^{\bullet-}, ACT}$) was measured by competition kinetics in a xanthine-xanthine oxidase (XOD) system (Guo et al., 2021b; Pasternack and Halliwell, 1979). The system (3 mL) contained xanthine (400 μ M), XOD (0.2 U) and DETAPAC (1 mM). The reference compound was nitrotetrazolium blue chloride (NBT), which reacts with $O_2^{\bullet-}$ with a second-order rate constant of $6 \times 10^4 \text{ M}^{-1} \text{ s}^{-1}$. The reaction of $O_2^{\bullet-}$ with NBT^{2+} yields the stable and colored formazan, for which the concentrations was measured by spectrophotometry (Hach DR6000, USA) at $\lambda=560 \text{ nm}$ ($\epsilon = 13800 \text{ M}^{-1}\text{cm}^{-1}$).

In the experiments where the test compound (M) was added together with the NBT^{2+} to the xanthine and XOD system, the generated $O_2^{\bullet-}$ were mainly consumed in the parallel reactions with NBT^{2+} and M



where M and M' are the test compound and the product from the reaction of $O_2^{\bullet-}$ with M, respectively; $k_{O_2^{\bullet-}, M}$ is the second-order rate constant for the reaction of $O_2^{\bullet-}$ with M.

Assuming a steady-state approximation to $O_2^{\bullet-}$, we can obtain

$$\frac{d[O_2^{\bullet-}]}{dt} = 0 = r - k_{O_2^{\bullet-}, NBT^{2+}} [NBT^{2+}][O_2^{\bullet-}] - k_{O_2^{\bullet-}, M} [M][O_2^{\bullet-}] \quad (S3)$$

where r is the rate of $O_2^{\bullet-}$ production from the enzymatically catalyzed reaction of xanthine with XOD. By rearranging Eq. S3, the concentration of $O_2^{\bullet-}$ is given as

$$[O_2^{\bullet-}] = \frac{r}{k_{O_2^{\bullet-}, NBT^{2+}} [NBT^{2+}] + k_{O_2^{\bullet-}, M} [M]} \quad (S4)$$

Meanwhile, the rate of formazan production (v) from Eq. S1 can be expressed as

$$v = \frac{d[\text{formazan}]}{dt} = k_{O_2^{\bullet-}, NBT^{2+}} [NBT^{2+}][O_2^{\bullet-}] \quad (S5)$$

44 By substituting Eq. S4 into Eq. S5, we obtain

$$45 \quad v = \frac{d[\text{formazan}]}{dt} = r \cdot \frac{k_{\text{O}_2^{\bullet-}, \text{NBT}^{2+}} [\text{NBT}^{2+}]}{k_{\text{O}_2^{\bullet-}, \text{NBT}^{2+}} [\text{NBT}^{2+}] + k_{\text{O}_2^{\bullet-}, \text{M}} [\text{M}]} \quad (\text{S6})$$

46 In the experiment where only the reference compound (NBT^{2+}) was added in the system
47 (no test compound was added), Eq. S3 changes to

$$48 \quad \frac{d[\text{O}_2^{\bullet-}]}{dt} = 0 = r - k_{\text{O}_2^{\bullet-}, \text{NBT}^{2+}} [\text{NBT}^{2+}] [\text{O}_2^{\bullet-}] \quad (\text{S7})$$

49 and the rate of formazan production (V) in the absence of M is given as

$$50 \quad V = \frac{d[\text{formazan}]}{dt} = k_{\text{O}_2^{\bullet-}, \text{NBT}^{2+}} [\text{NBT}^{2+}] [\text{O}_2^{\bullet-}] \quad (\text{S8})$$

51 Combining Eq. S7 and S8, we obtain

$$52 \quad r = V \quad (\text{S9})$$

53 Finally, by substituting Eq. S9 into Eq. S6 and rearranging the obtained equation, we can
54 get

$$55 \quad \frac{V}{v} = \frac{k_{\text{O}_2^{\bullet-}, \text{M}} [\text{M}]}{k_{\text{O}_2^{\bullet-}, \text{NBT}^{2+}} [\text{NBT}^{2+}]} + 1 \quad (\text{S10})$$

56 Therefore, by following the production rate of formazan in the absence and presence of
57 M (V and v), the second-order rate constant for the reaction of $\text{O}_2^{\bullet-}$ with the test compound
58 can be estimated from the slope of linear regression of Eq. S10. (Guo et al., 2021b; Pasternack
59 and Halliwell, 1979). As shown in Fig. S1, the slope of the regression line represents the $k_{\text{O}_2^{\bullet-}}$
60 ratio between target compound and NBT^{2+} (Eq. (S10)). Based on the $k_{\text{O}_2^{\bullet-}, \text{NBT}^{2+}}$, the second-
61 order rate constant for the reaction of $\text{O}_2^{\bullet-}$ with acetate was determined as $3.2 \times 10^4 \text{ M}^{-1} \text{ s}^{-1}$.

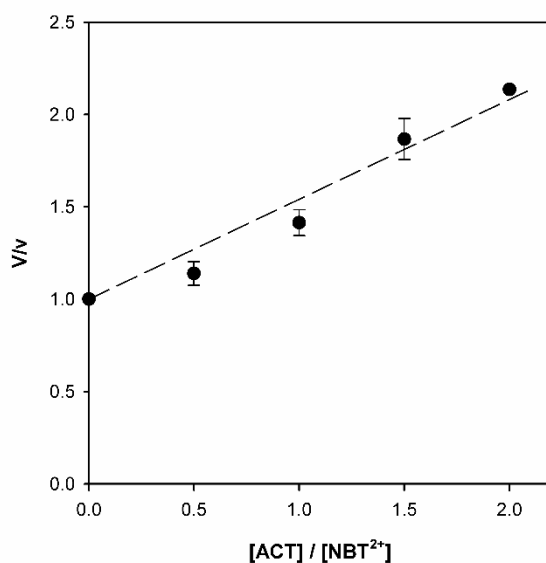


Fig. S1. Determination of the second-order rate constant for the reaction of acetate with $\text{O}_2^{\bullet-}$. Ratio of production rate of the reference compound (NBT^{2+}) as a function of molar ratios of acetate and NBT^{2+} . Reaction conditions: $[\text{NBT}^{2+}] = 1 \text{ mM}$, $[\text{Acetate}] = 0.5, 1.0, 1.5, \text{ and } 2 \text{ mM}$.

S3. Reaction of CCl_4 with $\text{O}_2^{\bullet-}$

The second-order rate constant for the reaction of CCl_4 with $\text{O}_2^{\bullet-}$ has been determined to be $1.1 \times 10^9 \text{ M}^{-1} \text{ s}^{-1}$ in a previous study using the same method as described in Text S1 (Guo et al., 2021a). In the present study, the second-order rate constant for the reaction of $\text{O}_2^{\bullet-}$ with CCl_4 was further verified using the competition kinetic method with 2,5-dichloro-*p*-benzoquinone ($k_{\text{O}_2^{\bullet-}} = 1.1 \times 10^9 \text{ M}^{-1} \text{ s}^{-1}$ (Bielski et al., 1985)) as the reference compound. As shown in Fig. S2a, the concentration of CCl_4 decreased only slightly slower than that of 2,5-dichloro-*p*-benzoquinone in the xanthine-xanthine oxidase system. After 20 min, the concentration of CCl_4 was abated by $\sim 20\%$ in the system. Although the decreases of CCl_4 concentrations are relatively small under the tested reaction conditions, they can still provide a valid estimation of the second-order rate constant. Based on the linear regression between the natural logarithm of the relative residual concentrations of CCl_4 and chlorobenzoquinone (Fig. S2a inset), the second-order rate constant for the reaction of $\text{O}_2^{\bullet-}$ with CCl_4 was determined as

$8.4 \times 10^8 \text{ M}^{-1} \text{ s}^{-1}$, which is very close to the previously reported value (within the experimental errors). This result confirms that CCl_4 can react rapidly with $\text{O}_2^{\bullet-}$, and the previously reported rate constant is reliable. Note that loss of CCl_4 through evaporation was evaluated in pure water and found to be negligible under the tested conditions (Fig. S2b).

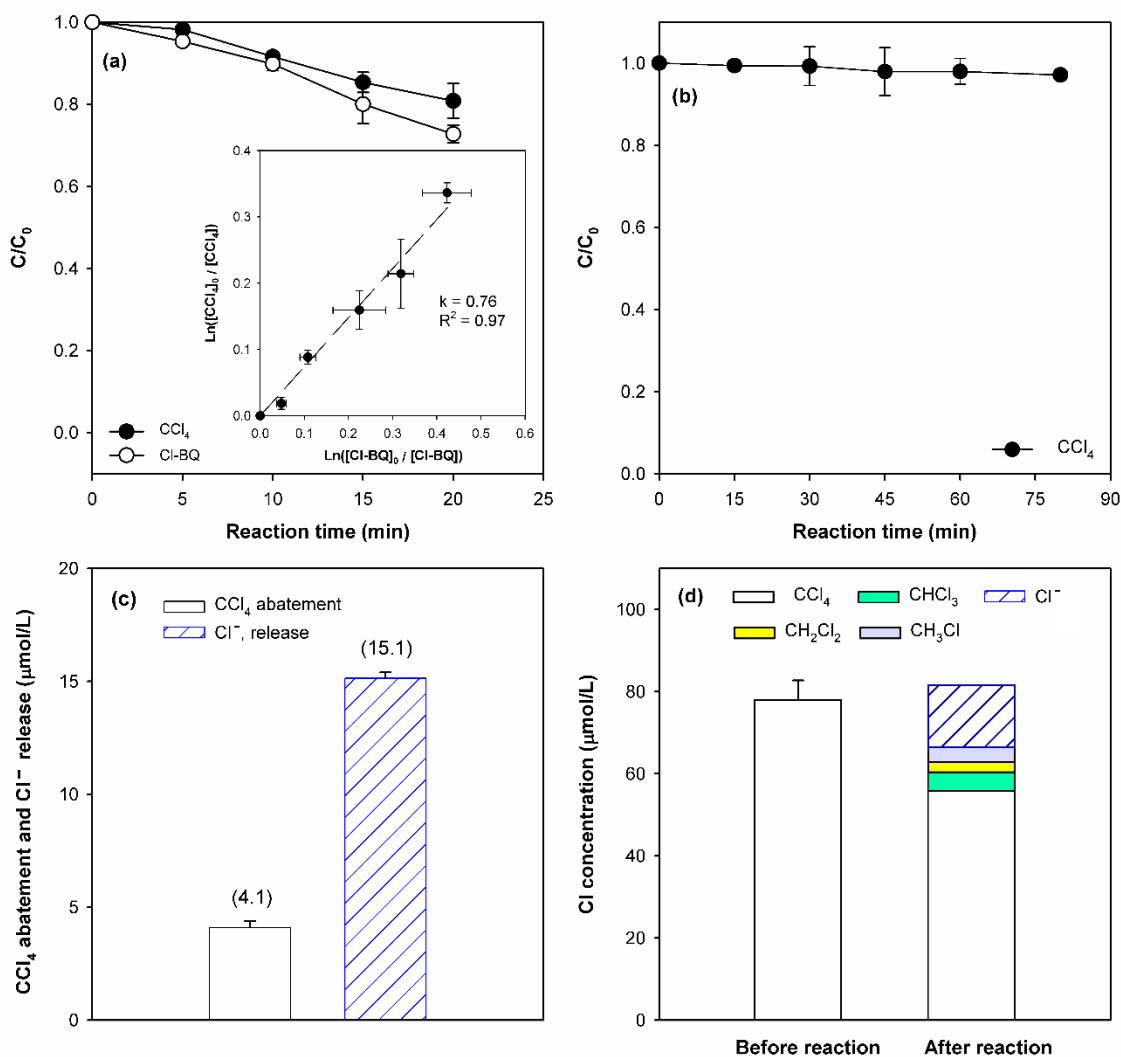


Fig. S2. Determination of the kinetics of the reaction between $\text{O}_2^{\bullet-}$ and CCl_4 . (a) Decrease of CCl_4 and 2,5-dichloro-*p*-benzoquinone concentrations in the xanthine-xanthine oxidase system; (b) blank experiment to test the stability of CCl_4 concentrations in water; (c) degraded CCl_4 and released Cl^- during the xanthine-xanthine oxidase process; (d) chlorine concentration balance before and after the xanthine-xanthine oxidase process. Experimental conditions: (a) $[\text{CCl}_4] = [\text{2,5-dichloro-}p\text{-benzoquinone}]_0 = 200 \mu\text{g/L}$, $[\text{xanthine}] = 400 \mu\text{M}$, $[\text{XOD}] = 0.3 \text{ U}$,

pH = 8.0 (phosphate buffered); (b) $[\text{CCl}_4]_0 = 200 \text{ } \mu\text{g/L}$; (c, d) $[\text{CCl}_4]_0 = 20 \text{ } \mu\text{M}$ ($\sim 3 \text{ mg/L}$),
[xanthine] = $800 \text{ } \mu\text{M}$, [XOD] = 0.8 U , pH = 8.0 (phosphate buffered).

To examine dechlorination of CCl_4 by $\text{O}_2^{\bullet-}$, $20 \text{ } \mu\text{M}$ CCl_4 was added in the xanthine-xanthine oxidase system. After the reaction was completed, approximately $4.1 \text{ } \mu\text{M}$ CCl_4 was degraded, while $15.1 \text{ } \mu\text{M}$ Cl^- was detected in the system. These data suggest that on average, 3.68 moles of Cl^- are released per mole of CCl_4 degraded during the reaction. Meanwhile, dechlorination transformation products of CCl_4 (CHCl_3 , CH_2Cl_2 , and CH_3Cl) were detected in the water, yielding a good overall molar Cl balance (Fig. S2c). The results observed herein are in agreement with the previous findings that $\text{O}_2^{\bullet-}$ is a strong nucleophile and can degrade chlorinated organics through a nucleophilic substitution mechanism (Hayyan et al., 2016; Mitchell et al., 2014; Smith et al., 2004).

S4. Ozone decomposition in the presence of CCl_4

Fig. S3 shows that with increasing CCl_4 concentrations, the rate of O_3 depletion decreased considerably during ozonation. This change can be mainly attributed to an enhanced suppression of the $\text{O}_2^{\bullet-}$ -promoted O_3 decomposition at higher CCl_4 concentrations. For the highest concentration of CCl_4 (6.3 mM), the observed pseudo-first order rate of O_3 depletion was $1.9 \times 10^{-3} \text{ s}^{-1}$, which is about three times the rate of O_3 depletion caused by the reaction with OH^- at pH 9 ($k = 7 \times 10^{-4} \text{ s}^{-1}$). This difference suggests that the $\text{O}_2^{\bullet-}$ -promoted O_3 decomposition is not fully suppressed at the applied CCl_4 concentration, or there are some impurities in the synthetic solutions that may initiate the O_3 decomposition (e.g., chemicals and buffers added in the synthetic solutions may contain some impurities that can react with O_3).

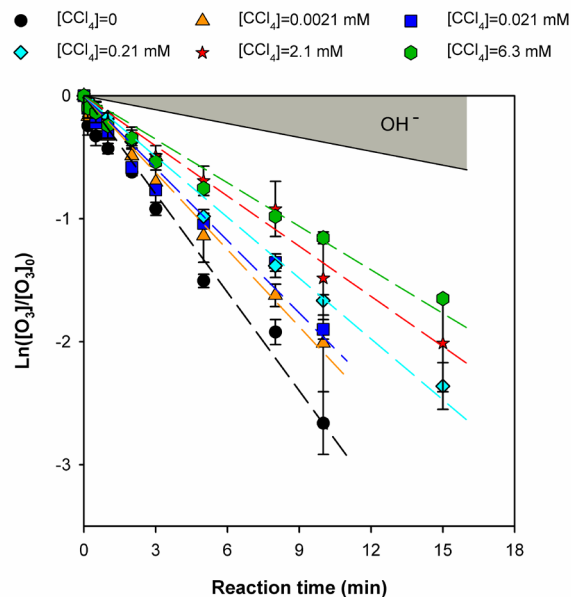


Fig. S3. Ozone decomposition kinetics in synthetic solutions containing varying concentrations of CCl_4 . Reaction condition: O_3 dose = 1.0 mg/L (0.021 mM), $[\text{CCl}_4]$ = 0.0021–6.3 mM, pH = 9 (buffered with 10 mM borate). The line of O_3 depletion by the reaction with OH^- is simulated using a second-order rate constant of $70 \text{ M}^{-1} \text{ s}^{-1}$ (Merényi et al., 2010).

S5. Methanol and acetate as a chain promoter and inhibitor, respectively

Fig. S4 shows the evolution of O_3 , $\cdot\text{OH}$, and $\text{O}_2^{\cdot-}$ concentrations during ozonation of synthetic solutions with constant total scavenging rate and various scavenging ratios of promotor (methanol)/inhibitor (acetate) (see Table 1).

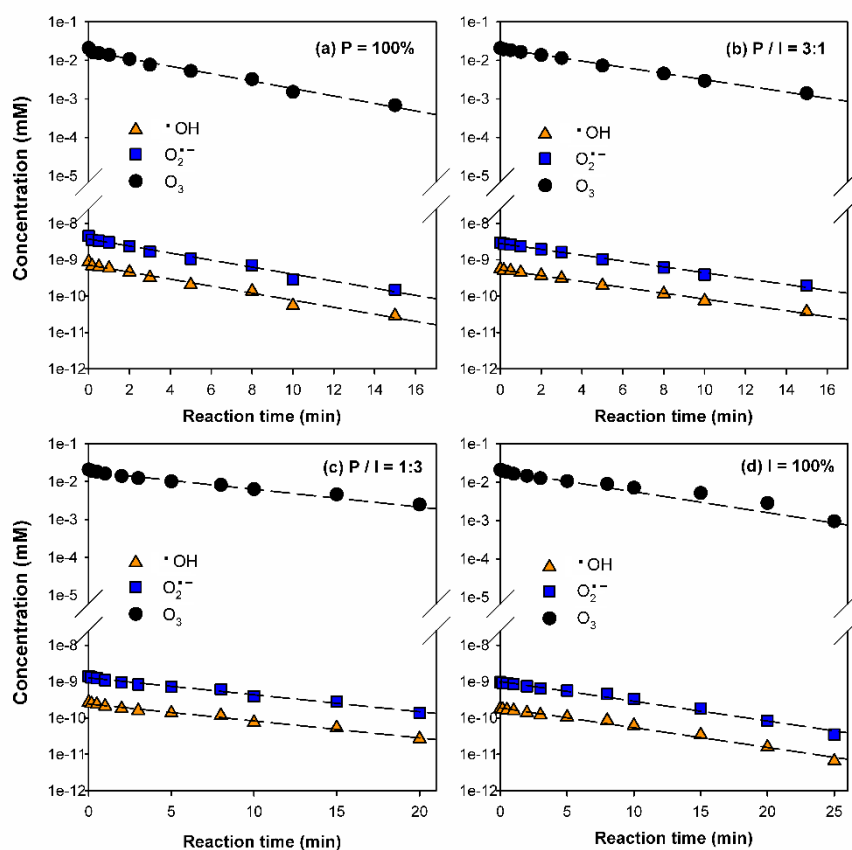


Fig. S4. Evolution of O_3 , $\cdot OH$, and $O_2^{\cdot -}$ concentrations for various promotor (P):inhibitor (I) scavenging ratios (P/I). (a) $P = 100\%$, (b) $P/I = 3:1$, (c) $P/I = 1:3$ and (d) $I = 100\%$ during ozonation. Reaction conditions: O_3 dose = 0.021 mM, pH ~ 8.0 (phosphate buffer, 10 mM), total scavenging rate = $1.2 \times 10^5 \text{ s}^{-1}$, $P=100\%$: $[MeOH] = 0.124 \text{ mM}$; $P/I = 3$: $[MeOH] = 0.093 \text{ mM}$, $[Acetate] = 0.38 \text{ mM}$; $P/I = 1/3$: $[MeOH] = 0.031 \text{ mM}$, $[Acetate] = 1.14 \text{ mM}$; $I=100\%$: $[Acetate] = 1.52 \text{ mM}$.

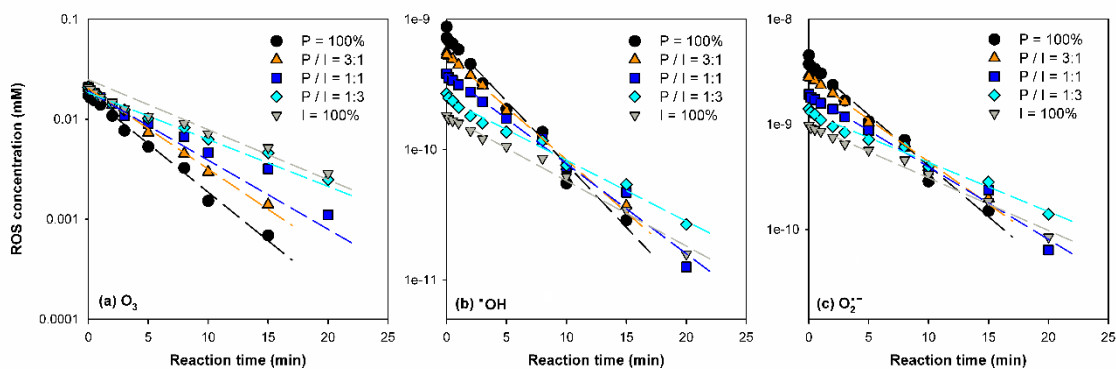


Fig. S5. Effects of the P/I ratios on the evolution of (a) O_3 , (b) $\cdot OH$, and (c) $O_2^{\cdot -}$ concentrations during ozonation of MeOH- and/or acetate-containing solutions. Reaction condition: O_3 dose = 0.021 mM, $[MeOH] = 0\text{--}0.124$ mM, $[acetate] = 0\text{--}1.14$ mM, pH ~ 8.0 (phosphate buffer, 10 mM), total scavenging rate = $1.2 \times 10^5 \text{ s}^{-1}$, P=100%: $[MeOH] = 0.124$ mM; P/I = 3: $[MeOH] = 0.093$ mM, $[acetate] = 0.38$ mM; P/I = 1: $[MeOH] = 0.062$ mM, $[acetate] = 0.76$ mM; P/I = 1/3: $[acetate] = 0.031$ mM, $[acetate] = 1.14$ mM; I=100%: $[acetate] = 1.52$ mM.

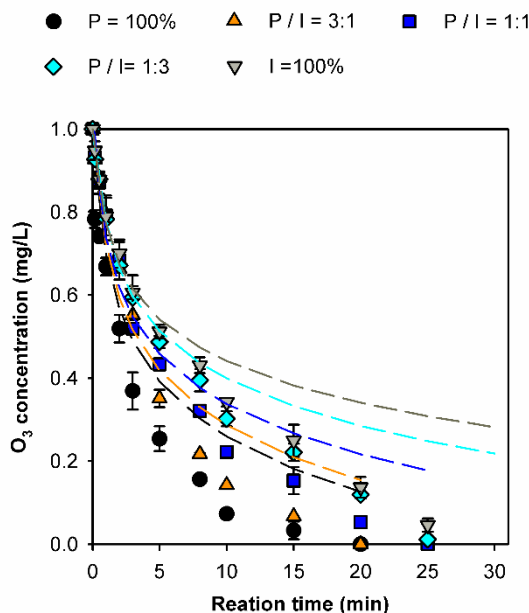


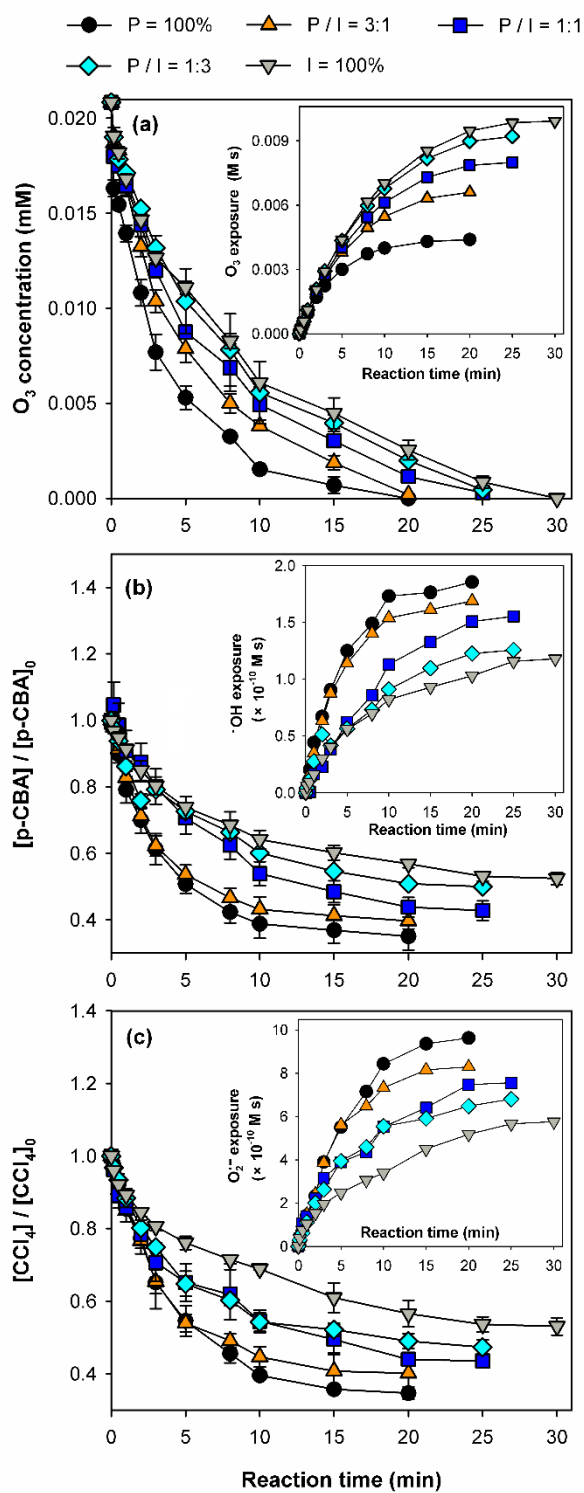
Fig. S6. Modelling of ozone decomposition in MeOH- and /or acetate-containing solutions by Eq. 15 ($k_{\cdot OH, O_3} = 1.1 \times 10^8 \text{ M}^{-1} \text{ s}^{-1}$). The symbols are experimental and the dashed lines modelling results. Reaction condition: O_3 dose = 0.021mM, pH ~ 8.0 (phosphate buffer, 10

147 mM), total scavenging rate = $1.2 \times 10^5 \text{ s}^{-1}$, P =100%: [MeOH] = 0.124 mM, P/I = 3: [MeOH]
148 = 0.093 mM, [acetate] = 0.38 mM; P/I = 1: [MeOH] = 0.062 mM, [acetate] = 0.76 mM; P/I =
149 1/3: [MeOH] = 0.031 mM, [acetate] = 1.14 mM; I =100%: [acetate] = 1.52 mM.

150

151 **S6. Methanol and *tert*-butanol as a chain promoter and inhibitor, respectively**

152 The results obtained with *tert*-butanol as an inhibitor (Fig. S6) were generally very similar
153 to those obtained with acetate (Fig. 1).



154

155 **Fig. S7.** Ozonation of synthetic solutions with constant total scavenging rate and various
 156 promotor (methanol):inhibitor (*tert*-butanol) scavenging ratios (P/I): Effect of the P/I ratios on
 157 (a) ozone decrease, (b) *p*CBA abatement and (c) CCl_4 abatement during ozonation of MeOH-
 158 and/or TBA-containing solutions. The insets in Fig. S6a, b and c show the O_3 exposures, $\cdot OH$

exposures, and $O_2^{\bullet -}$ exposures, respectively. Reaction conditions: O_3 dose = 0.021mM, pH \sim 8.0 (phosphate buffer, 10 mM), total scavenging rate = $1.2 \times 10^5 s^{-1}$, P =100%: [MeOH] = 0.124 mM; P/I = 3: [MeOH] = 0.093 mM, [TBA] = 0.06 mM; P/I = 1: [MeOH] = 0.062 mM, [TBA] = 0.12 mM; P/I = 1/3: [MeOH] = 0.031 mM, [TBA] = 0.18 mM; I =100%: [TBA] = 0.24 mM.

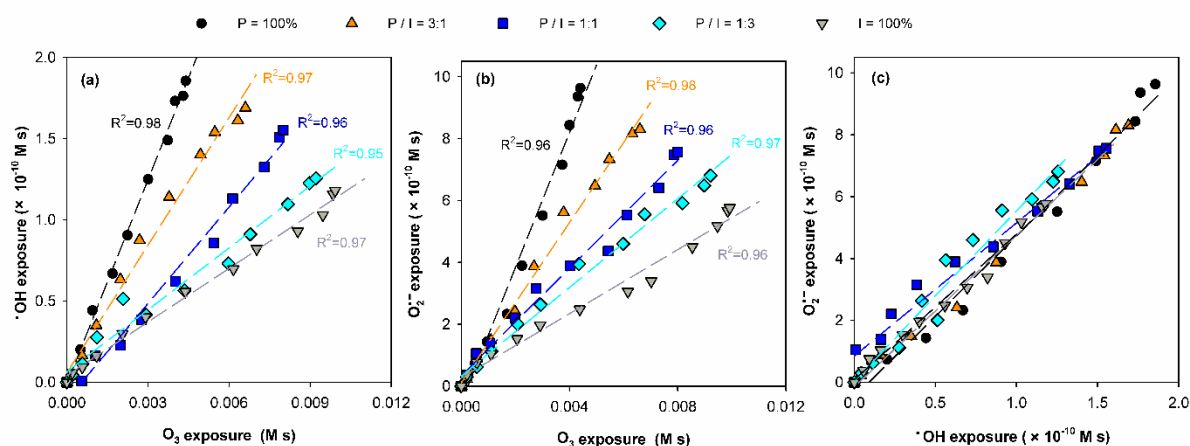


Fig. S8. Ozonation of synthetic solutions with constant total scavenging rate and various promotor (methanol):inhibitor (*tert*-butanol) scavenging ratios (P/I): (a) R_{ct} , (b) R_{SO} , and (c) R_{SH} as a function of the P/I ratios. Reaction conditions: O_3 dose = 0.021mM, pH \sim 8.0 (phosphate buffer, 10 mM), total scavenging rate = $1.2 \times 10^5 s^{-1}$. P =100%: [MeOH] = 0.124 mM; P/I = 3: [MeOH] = 0.093 mM, [TBA] = 0.06 mM; P/I = 1: [MeOH] = 0.062 mM, [TBA] = 0.12 mM; P/I = 1/3: [MeOH] = 0.031 mM, [TBA] = 0.18 mM; I =100%: [TBA] = 0.24 mM.

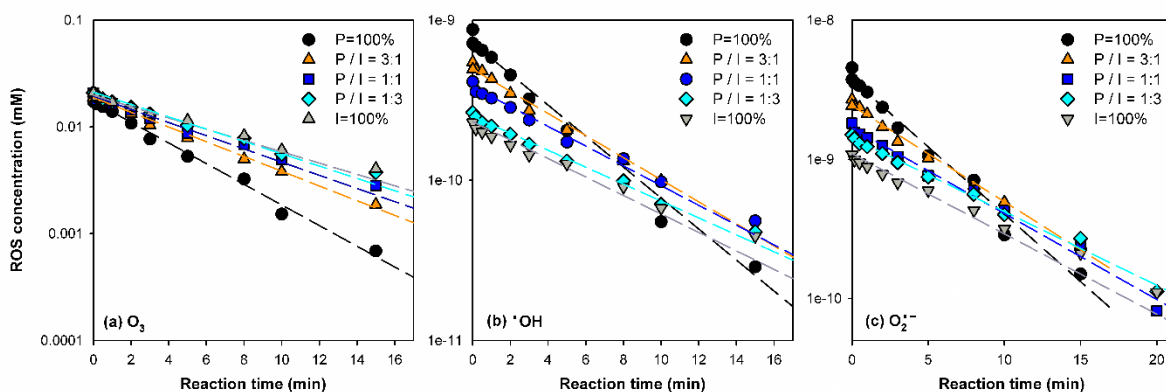


Fig. S9. Ozonation of synthetic solutions with constant total scavenging rate and various promotor (methanol):inhibitor (*tert*-butanol) scavenging ratios (P/I): Effects of the P/I ratios on the evolution of (a) O_3 , (b) $\cdot OH$, and (c) $O_2^{\cdot -}$ concentrations. Reaction conditions: O_3 dose = 0.021 mM, pH ~8.0 (phosphate buffer, 10 mM), total scavenging rate = $1.2 \times 10^5 s^{-1}$. P=100%: [MeOH] = 0.124 mM; P/I = 3: [MeOH] = 0.093 mM, [TBA] = 0.06 mM; P/I = 1: [MeOH] = 0.062 mM, [TBA] = 0.12 mM; P/I = 1/3: [MeOH] = 0.031 mM, [TBA] = 0.18 mM; I=100%: [TBA] = 0.24 mM.

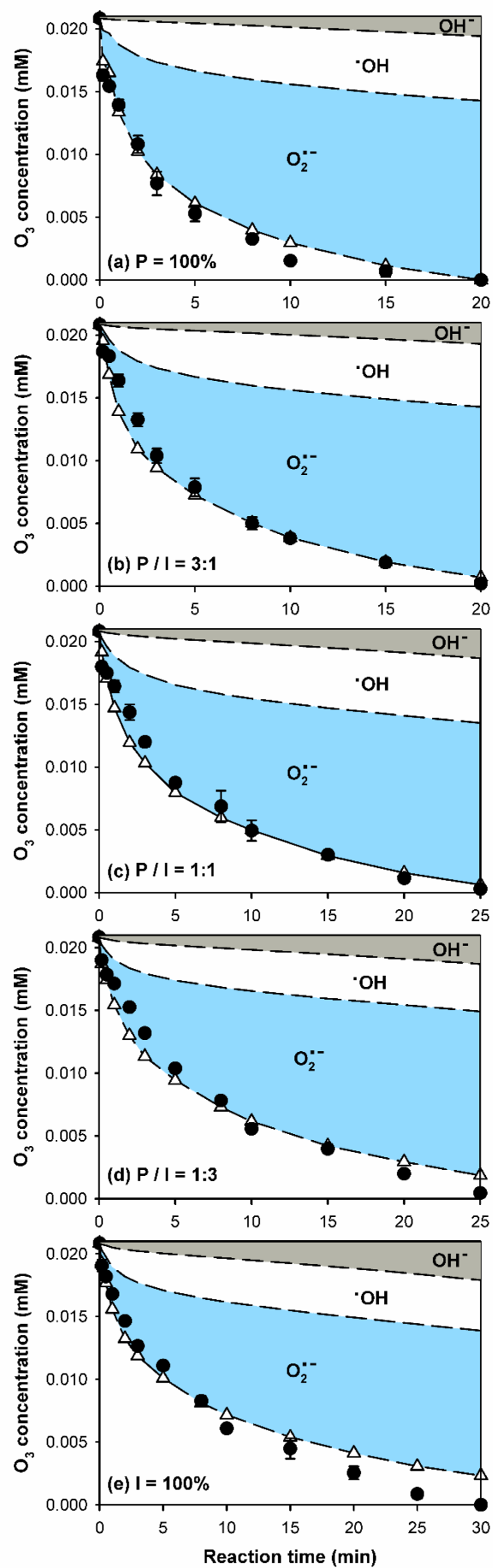


Fig. S10. Modelling of ozone decomposition in MeOH/TBA-containing solutions using Eq. 15 ($k_{\cdot\text{OH},\text{O}_3} = 3.0 \times 10^9 \text{ M}^{-1} \text{ s}^{-1}$). The filled circles are experimental results, the open triangles are modelling results calculated using Eq. 15, and the dash lines are drawn based on the relative contribution of OH^- , $\cdot\text{OH}$, and $\text{O}_2^{\cdot-}$, respectively, to the O_3 decay calculated by Eqs. 16-18. Reaction conditions: O_3 dose = 0.021mM, pH ~8.0 (phosphate buffer, 10 mM), total scavenging rate = $1.2 \times 10^5 \text{ s}^{-1}$. P=100%: $[\text{MeOH}] = 0.124 \text{ mM}$; P/I = 3: $[\text{MeOH}] = 0.093 \text{ mM}$, $[\text{TBA}] = 0.06 \text{ mM}$; P/I = 1: $[\text{MeOH}] = 0.062 \text{ mM}$, $[\text{TBA}] = 0.12 \text{ mM}$; P/I = 1/3: $[\text{MeOH}] = 0.031 \text{ mM}$, $[\text{TBA}] = 0.18 \text{ mM}$; I=100%: $[\text{TBA}] = 0.24 \text{ mM}$.

S7. Ozonation of natural waters

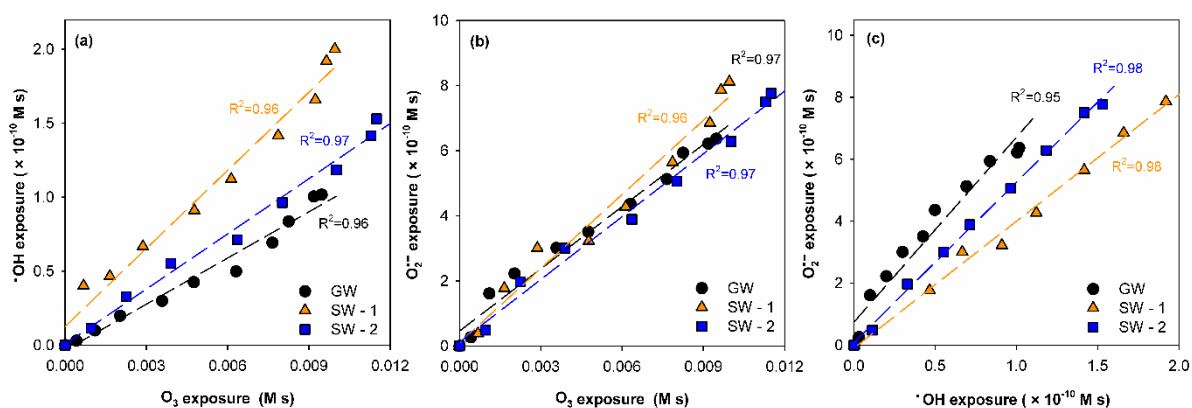


Fig. S11. Ozonation of a selected groundwater (GW) and two surface waters (SW-1, SW-2): (a) R_{ct} , (b) R_{SO} , and (c) R_{SH} . Reaction conditions: Specific O_3 dose = 1.0 mg O_3 /mg DOC. For water quality data of the three real waters see Table 2.

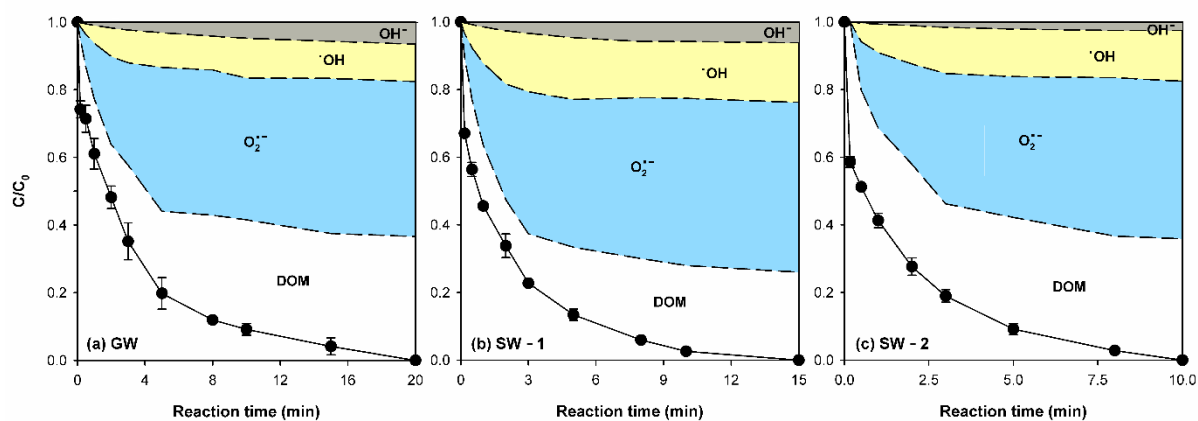


Fig. S12. Relative contribution of OH^- , $\cdot\text{OH}$, $\text{O}_2^{\cdot-}$, and DOM to the relative ozone depletion (C/C_0) in the selected (a) groundwater (GW) and (b, c) surface waters (SW-1, SW-2). The solid circles are experimental data and the dash lines are drawn based on the relative contributions of OH^- , $\cdot\text{OH}$, $\text{O}_2^{\cdot-}$, and DOM to O_3 decay calculated using Eqs. 21-24. Reaction conditions: Specific O_3 dose = 1.0 mg O_3 /mg DOC. Refer to Table 2 for water quality data of the three real waters.

References

- Bielski, B.H., Cabelli, D.E., Arudi, R.L. and Ross, A.B. 1985. Reactivity of HO_2/O_2^- radicals in aqueous solution. *J Phys Chem Ref Data* 14(4), 1041-1100.
- Buxton, G.V., Greenstock, C.L., Helman, W.P. and Ross, A.B. 1988. Critical Review of rate constants for reactions of hydrated electrons, hydrogen atoms and hydroxyl radicals ($\cdot\text{OH}/\cdot\text{O}^-$ in Aqueous Solution. *J Phys Chem Ref Data* 17(2), 513-886.
- Guo, Y., Zhan, J., Yu, G. and Wang, Y. 2021a. Evaluation of the concentration and contribution of superoxide radical for micropollutant abatement during ozonation. *Water Res* 194, 116927.

215 Guo, Y., Zhang, Y., Yu, G. and Wang, Y. 2021b. Revisiting the role of reactive oxygen
 216 species for pollutant abatement during catalytic ozonation: The probe approach versus
 217 the scavenger approach. *Applied Catalysis B: Environmental* 280, 119418.
 218 Hayyan, M., Hashim, M.A. and AlNashef, I.M. 2016. Superoxide Ion: Generation and
 219 Chemical Implications. *Chem Rev* 116(5), 3029-3085.
 220 Merényi, G., Lind, J., Naumov, S. and von Sonntag, C. 2010. The Reaction of Ozone with
 221 the Hydroxide Ion: Mechanistic Considerations Based on Thermokinetic and Quantum
 222 Chemical Calculations and the Role of HO₄⁻ in Superoxide Dismutation. *Chemistry –*
 223 *A European Journal* 16(4), 1372-1377.
 224 Mitchell, S.M., Ahmad, M., Teel, A.L. and Watts, R.J. 2014. Degradation of
 225 Perfluorooctanoic Acid by Reactive Species Generated through Catalyzed H₂O₂
 226 Propagation Reactions. *Environmental Science & Technology Letters* 1(1), 117-121.
 227 Pasternack, R.F. and Halliwell, B. 1979. Superoxide dismutase activities of an iron
 228 porphyrin and other iron complexes. *J Am Chem Soc* 101(4), 1026-1031.
 229 Smith, B.A., Teel, A.L. and Watts, R.J. 2004. Identification of the reactive oxygen species
 230 responsible for carbon tetrachloride degradation in modified Fenton's systems. *Environ*
 231 *Sci Technol* 38(20), 5465-5469.
 232 Wolfenden, B.S. and Willson, R.L. 1982. Radical-cations as reference chromogens in
 233 kinetic studies of one-electron transfer reactions: pulse radiolysis studies of 2,2' -
 234 azinobis-(3-ethylbenzthiazoline-6-sulphonate). *Journal of the Chemical Society, Perkin*
 235 *Transactions* 2 (7), 805-812.