

Reduced emissions of warm mix asphalt during construction

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Warm mix asphalt (WMA) is produced at lower temperature and hence its production in the asphalt plant is favourable in terms of energy consumption and CO₂ emission. As a side effect, it is expected that the emissions during pavement construction are reduced too, which is highly relevant for work place hygiene and for environmental protection.

In a field trial five pavement sections were constructed with different warm-mix asphalt types and one hot mix asphalt (HMA) used as a reference. During construction several road workers were equipped with personnel samplers to determine their exposure to pollutants relevant for occupational health. Particular attention was paid to the total particulate matter (TPM) and the polycyclic aromatic hydrocarbons (PAH). In addition, emissions of TPM, PAH, and other organic pollutants (volatile organic compounds, TVOC) were sampled on paver and rolling compactor. To evaluate total emissions of road construction, an integrative pollutant sample was collected downwind of the construction site and mass emissions were estimated using a tracer gas technique with constant dosing at known source strength. In parallel a series of laboratory experiments were carried out on the same material to elucidate differences in the emission behaviour between the warm mix asphalt types and in relation to HMA under more controlled conditions.

The laboratory experiments indicate a 90 % reduction of TPM and 50-70 % lower TVOC values by the use of WMAs in comparison to hot mix asphalt. Emissions of pollutants during road construction, however, were low for all asphalt types, mostly below the occupational health limits and the detection limits of the applied analytical techniques. The study demonstrates the benefit of temperature reduction during road construction for occupational health and the environment.

Keywords: warm mix asphalt (WMA); emissions during construction; occupational health; PAH; TVOC

1 **1. Introduction**

2 Reduction of energy consumption and CO₂ emission is an important issue in asphalt produc-
3 tion. Hot mix asphalt (HMA) is still the actual standard in many countries although new tech-
4 nologies are available to produce asphalt at lower temperature (Kheradmand, Muniandy, Hua,
5 Yunus, & Solouki., 2014). Warm mix asphalt mixtures (WMA) are typically produced at
6 temperatures around 20 to 40°C lower than traditional hot mix asphalt. Different techniques to
7 produce warm mix asphalt are available, e.g. addition of waxes, chemical modifiers, zeolites,
8 foam bitumen, and it is difficult to evaluate the pros and cons of the different WMA types
9 (Rubio, Martínez, Baena, & Moreno, 2012). Therefore, objective criteria are required to sup-
10 port decision-making by road authorities and other customers, including durability, ecological
11 balance and occupational health, for the right choice of WMA pavement type.

12 Temperature reduction is an efficient way to reduce emission of pollutants during road
13 construction as has been demonstrated in previous studies (Rubio, 2012; Hugener, Emmeneg-
14 ger, & Mattrel, 2007; Autelitano, Bianchi, & Giuliani, 2017). However, environmental agen-
15 cies are cautious regarding WMA due to the lack of emission data. Furthermore, chemical ad-
16 ditives of unknown composition are added to some WMA types, which could be released
17 during road construction resulting in an occupational hazard for road workers (Rühl, Musan-
18 ke, Kolmsee, Priess, & Breuer, 2007). Moreover, warm mix techniques are often used in con-
19 junction with recycled asphalt pavement material (RAP). RAP is sometimes heavily contami-
20 nated with tar, known to contain large amounts of toxic polycyclic aromatic hydrocarbons
21 (PAH), which are released to some extent during construction. Measurements taken during
22 actual road construction are required to determine the exposure relevant to occupational
23 health, i.e. the harmful substances directly inhaled by the workers from mixtures and ma-
24 chines (paver, compactor, truck). Few emission measurements during road construction were
25 conducted so far due to the complex sampling and analytical technique (Brandt, &

1 Cordingley, 1992; Hugener, Emmenegger, & Mattrel, 2010; Kriech et. al., 2010) and even
2 less for WMA (Rubio, Moreno, Martínez, Martínez-Echevarría, & Vázquez, 2013; Leon, &
3 Jensen, 2012). One study on WMA with wax was restricted to the analysis of VOC emissions
4 from bitumen in a laboratory setup by headspace gas chromatography-mass spectrometry
5 (Autelitano et. al., 2017). More than 200 volatile compounds have been identified including
6 linear and branched hydrocarbons, cyclic alkanes, aromatic hydrocarbons and heterocyclic
7 (mainly sulphur) compounds. Another laboratory study on pollutant emissions of bitumen and
8 bituminous mixtures investigated the influence of different parameters (Gaudefroy, Viranai-
9 ken, Paranhos, Jullien, & de La Roche, 2010).

10 In the paper presented here, emission and occupational health measurements were
11 conducted during sequential construction of four warm asphalt mixture types and one HMA
12 reference material in a field trial in Switzerland (Hugener et al., 2016).

13 **2. Materials and methods**

14 This study was part of the extensive research project PLANET (Arn, 2017) with different sub-
15 projects dealing with mix design, quality control, mixing plants, life cycle analysis, durability
16 and occupational health. A central part of the project was a field test near Wohlen in the can-
17 ton Bern (Switzerland), where on a rural road with a length of 850 m and a width of 6 m, a
18 binder course AC B 16 S (EN 13108-1, 2006) was constructed on October 6th 2012. Five dif-
19 ferent asphalt types, four WMAs and one HMA used as a reference (Table 1), were imple-
20 mented sequentially in 5 equal sections of 170 m length. All mixes were identical in terms of
21 binder content, penetration value and aggregate size distribution. Three different types of
22 warm mix technologies were applied including foam bitumen, zeolite (WZ) and a chemical
23 additive (WC) (Table 1). Two variations of foam asphalt were produced, one with 50 % RAP
24 (WFR) and one without recycling material (WF).

1 The construction of the five test sections did take place on the same day. The weather
2 was dry with temperatures ranging from 7.4 °C in the morning to 23.6 °C in the late after-
3 noon. There were weak winds (<0.1 m/s) from northeast in the morning, which turned to
4 southwest after the second section was finished. The road was constructed towards the north-
5 east with two asphalt pavers in parallel. First, a conventional hot mix asphalt was placed as a
6 reference with an asphalt temperature around 145°C. Next, two warm-mix asphalts were laid
7 at temperatures between 120 and 130°C. The first contained a surface-active chemical addi-
8 tive for reducing the friction in the mixture (WC), the second a hydrated zeolite (WZ), where
9 the adsorbed water formed temporary bituminous foam. Finally, the last two sections were
10 constructed using foamed asphalt (WF, WFR) at a temperature of 106°C.

Table 1. Temperatures used for WMA and HMA production / construction

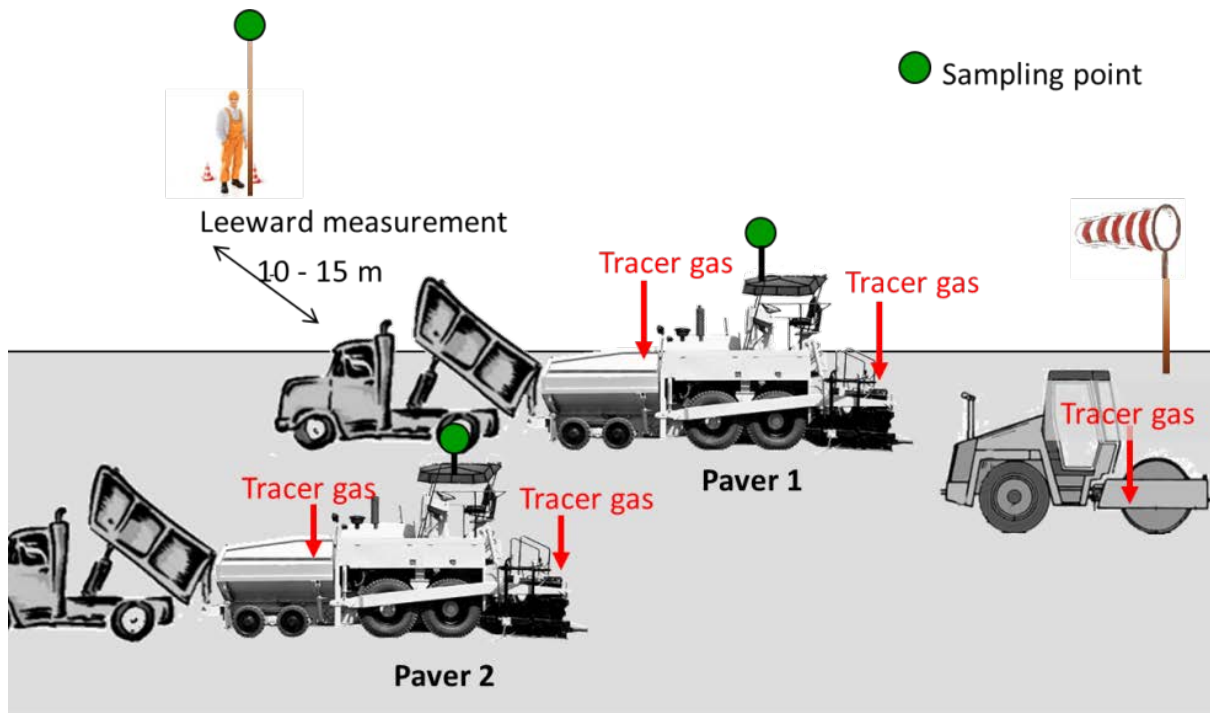
Material	Warm mix technology	Production temperature (°C)	Construction temperature (°C)
HMA (reference)	-	165	145
WC	Chemical additive	135	123
WZ	Zeolite	135	128
WF	Foam bitumen	115	106
WFR	Foam bitumen + 50% RAP	115	107

11

12 **2.1. Emission measurements and sampling**

13 Six road workers with different tasks (2 paver drivers, 1 roller driver, 3 workers with univer-
14 sal tasks) were equipped with personal samplers to measure their occupational health expo-
15 sure. Stationary sampling devices were installed on two pavers with air sampling at roof
16 height of the pavers (Figure 1). An additional sampling was conducted at the lee side of the
17 construction site in order to determine the total emissions of the pavement activities. Dilution
18 of pollutants for the lee side measurements was considered using a tracer gas technique with
19 constant tracer gas dosing (SF₆ or SF₅CF₃) at the main emission sources (both paver and one
20 roller compacter) (Mohn et. al., 2018). Total emissions of a target substance were calculated

1 as the ratio of target and tracer gas multiplied with the mass emissions of the tracer gas. A
2 combination of glass fibre filters and adsorption tubes filled with different resin types were
3 used to collect the pollutants at the three stationary measurement points and in the personal
4 sampler. Total particulate matter (TPM), total volatile organic compounds (TVOC) and the 16
5 EPA-PAHs (List of 16 PAH of the United States Environmental Protection Agency) were col-
6 lected on membrane filters and/or adsorption tubes (Supelco Orbo-43, Merck & Co., US)
7 (Hugener et. al., 2010) with a flow of 2 l/min. TVOC were accumulated on thermal desorption
8 tubes (TDS-Tenax, Dräger Safety AG, Switzerland). Separate samples were collected for each
9 test section with a length of 170 m, which limited the sampling interval to 40 minutes.

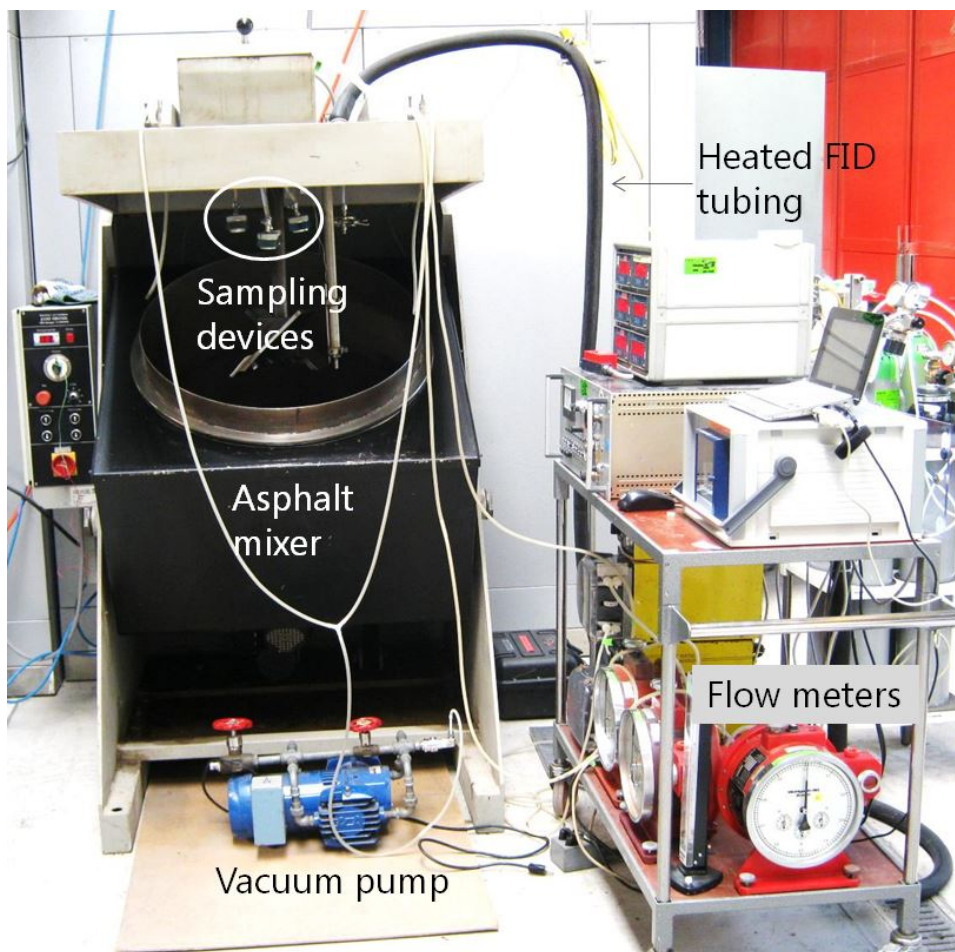


10
11 Figure 1. Schematic for tracer gas dosing and sampling points used during the emission measurements
12

13 Since the comparability of emission data for different WMA types based on field sampling is
14 limited even under defined construction conditions due to the differing topographical and me-
15 teorological conditions, results were complemented by laboratory measurements. This also
16 enabled a higher level of sensitivity for the analysis of pollutants, due to lower dilution ratios
17 and longer sampling intervals, as compared to the real construction site, where sampling in-

1 intervals were limited due to the short test sections. For the lab measurements, four WMA and
2 one HMA subsample were collected at the construction site. In the laboratory the subsamples
3 were pre-heated first in closed containers over 3 hours to approx. 90°C, in order to soften the
4 material, before the samples were transferred in a closed 150 kg laboratory mixer at construc-
5 tion temperature (Figure 2) and sampling was initiated. In this preheating phase some emis-
6 sions were lost, which, however also applies for the transport of the asphalt mixes from the
7 production plant to the construction site. Sampling for TPM, TVOC and EPA-PAHs was con-
8 ducted identically to the field measurements. In addition, the total organic carbon content was
9 measured continually with a flame ionization detector (FID) connected through a heated Tef-
10 lon tubing (200°C) to the asphalt mixer (Figure 2) (EN 12619). FID measurements were not
11 carried out in the field trial as the device was not suitable for battery operation.

12



13

14 Figure 2. Laboratory measurements with 150 kg closed asphalt mixer

1 **2.2. Chemical analysis**

2 Filters and adsorption tubes were extracted separately. Filters were extracted twice with 2 mL
3 toluene in an ultrasonic bath. The solution was filtered and diluted to a volume of exactly 5
4 mL (Solution A). The resin of the PAH adsorption tubes was extracted twice with 2 mL of
5 toluene within a 30 minute period with gentle shaking and diluted to 5 mL (Solution B). From
6 one part of solution A the solvent was evaporated to determine the toluene soluble matter
7 (TSM).

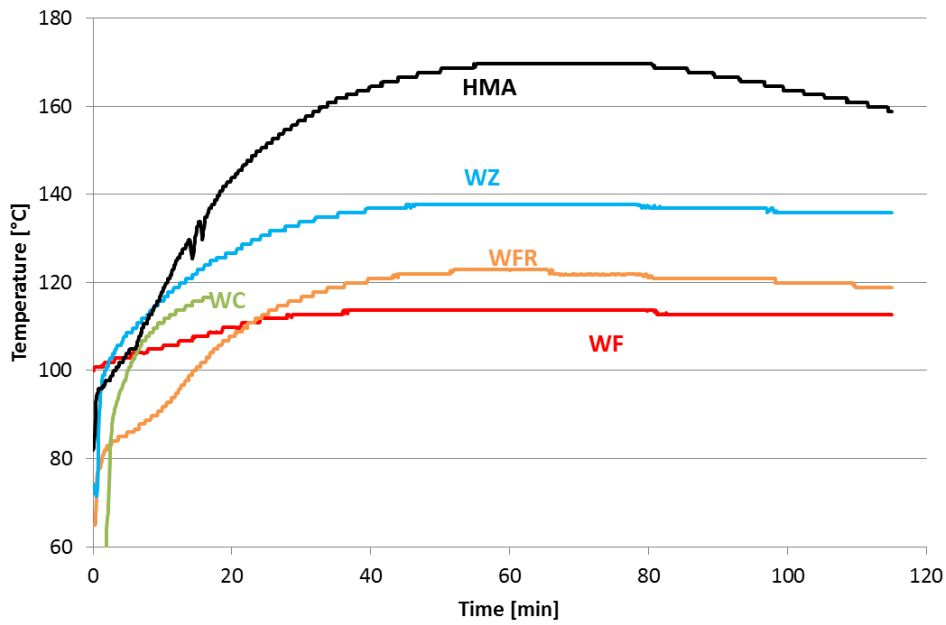
8 To provide a representative sample for the analysis of the 16 EPA-PAH, 500 µl of
9 both solutions (A, B) were combined and an aliquote of 100 µl was doped with an internal
10 standard and analysed by GC-MS in multi-ion detection mode. TVOC analysis including
11 sample preparation was performed according to ISO 16000-6 (2011). The substances enriched
12 on the thermo-desorption tube were desorbed and after cryofocusing analysed by gas chroma-
13 tography using a helium carrier gas and a mass selective detector with electron ionization
14 (EI). Compound identification was performed by means of retention time and comparing the
15 recorded mass spectra with a spectral library.

16 17 **3. Results and discussion**

18 In the beginning of the 2 hours laboratory experiments the material was heated to the con-
19 struction temperature which was reached after 40 to 50 minutes (Figure 3). In the case of the
20 reference HMA, a steep increase of the emissions (FID signal) is observed, which decreased
21 in a similar way after the target temperature of 170°C has been reached (Figure 4). For the
22 WMA this behaviour was significantly less pronounced, emissions stayed at the same level
23 for most of the time. Probably, the preheating phase was too long and depletion of the volatile
24 compounds occurred already in the oven.

25

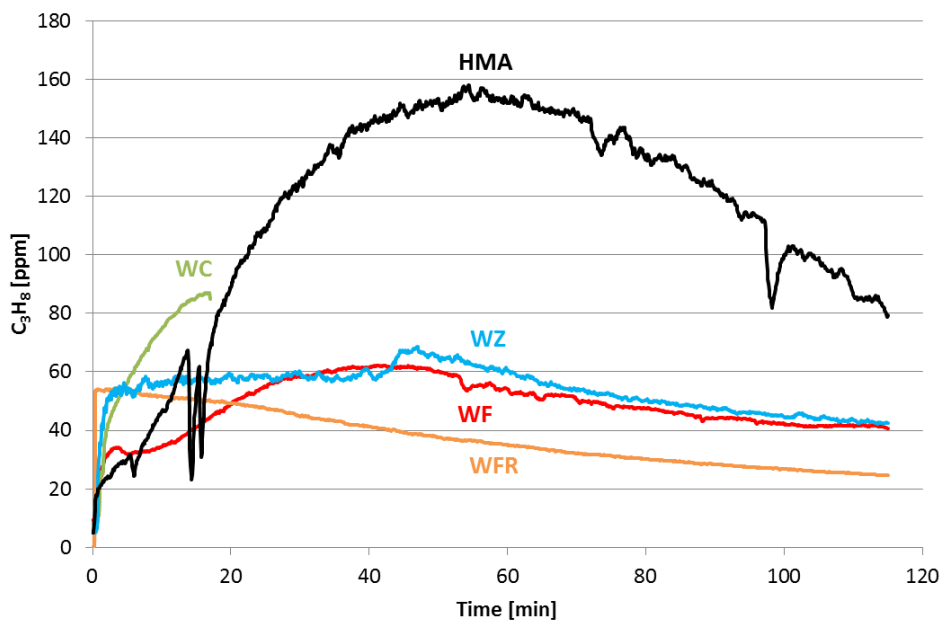
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2

3 Figure 3. Temperature of the asphalt mixtures in the laboratory mixer. In the emission measurements
4 of material WC the data transmission was corrupted after 17 minutes causing erroneous data.

5



6

7 Figure 4. FID-signal representing the organic carbon emissions during the mixing procedure in the la-
8 boratory mixer. Short-term variations in the concentration were caused by temporary openings of the
9 mixer, which resulted in dilution with ambient air and therefore in a drop in the hydrocarbon concen-
10 tration.

11

1 **3.1. Total particulate matter and total soluble matter**

2 The length of each test field of 170 m, equivalent to 40 minutes of construction time, turned
3 out to be too short for the measurement of the low pollution concentrations. Consequently, not
4 enough solid particles were collected on the filters for gravimetric analysis and most of the
5 measured concentrations were below the detection limits. Total emissions for the road-
6 building were estimated based on target gas concentrations at the "leeward measurement site"
7 and a dilution ratio determined from tracer gas concentrations and the volume of tracer gas
8 supplied.

9 In the field measurements, all TPM results, for all sampling locations (e.g. personal
10 samplers) and all asphalt types, including HMA, were below detection limit and significantly
11 below the Swiss regulated maximum workplace concentration (MAK) of 10 mg/m^3 for bitu-
12 men fumes and aerosols (Suva, 2018).

13 Laboratory measurements indicated that emissions of TPM are essentially influenced
14 by the temperature of the mixture. Therefore, TPM emissions of HMA (5.2 mg/m^3) were
15 around ten times higher compared to warm mix asphalt, with TPM emissions close to the de-
16 tection limits of 0.3 mg/m^3 . Between the different WMA technologies no clear difference in
17 TPM and TSM emissions is visible. A better correlation was observed between the maximum
18 asphalt mix temperatures and FID results. In the emission measurements of material WC the
19 data transmission was corrupted after 17 minutes causing erroneous data. Hence, the maxi-
20 mum temperature and the total carbon amount (FID) in Table 2 were estimated from Figure 3
21 and 4 taking into account the data of the first 17 minutes. Assuming a similar evolution of
22 temperature and FID signal like for the other mixtures, the maximum temperature was esti-
23 mated to be between 120 and 130°C . The average total carbon amount of WC is expected to
24 be higher than for WF but lower than for HMA. In contrast to findings of Gaudefroy et. al.

1 (2010), we did not observe an increase in emissions due to water present in the foam bitumen,
2 which might be due to the low water content of around 1%.

Table 2. Total carbon, TPM and TSM concentrations in the laboratory experiments

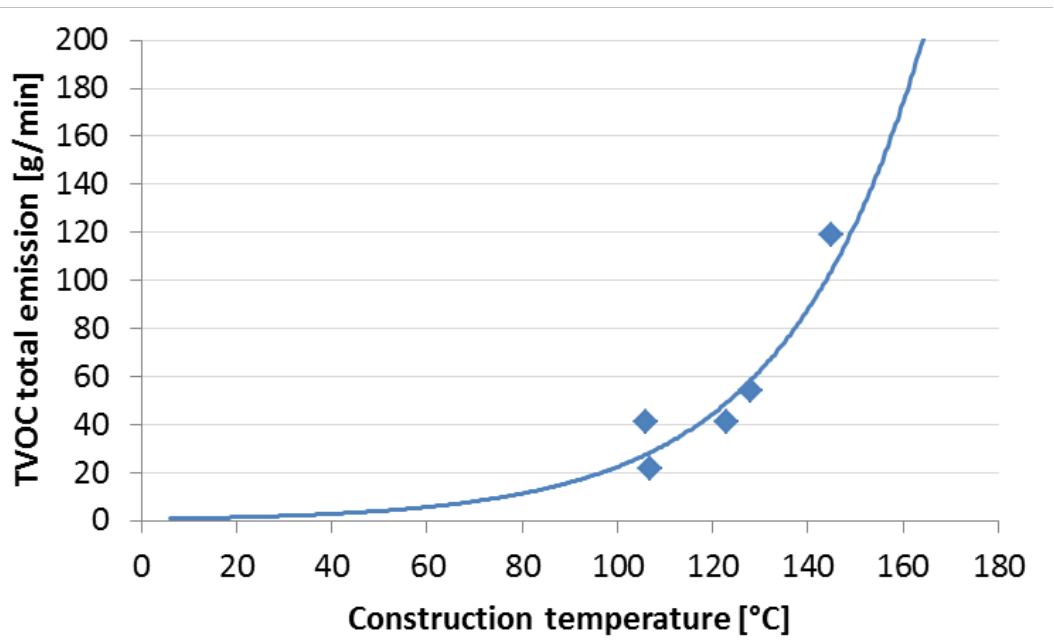
Material	Maximum temperature of the asphalt mix (°C)	Average of the total carbon amount (FID) (ppm)	TPM (mg/m ³)	TSM (mg/m ³)
HMA	164	124	5.2 ± 3.9	3.0 ± 2.4
WC	120-130*	60-90*	-	-
WZ	136	53	0.4 ± 0.1	0.3 ± 0.1
WF	113	34	<0.3	0.2 ± 0.2
WFR	120	50	<0.3	0.3 ± 0.2

3 *estimated values

4 **3.2. Total volatile organic compounds (TVOC)**

5 In the laboratory experiments, TVOC emissions for all WMA types displayed a similar emis-
6 sion profile with no emissions of additives or their derivatives being detected. As expected,
7 the proportion of higher boiling point hydrocarbons n-octane to n-eicosane within total TVOC
8 emissions was dominant, at 80% to 90%. The concentrations of individual substances relevant
9 with respect to toxicity (e.g. benzene, toluene, cyclo-hexanone, benzaldehyde) were at least
10 two times lower than the prevailing MAK values.

11



1 Figure 5. Correlation between TVOC total emissions and construction temperature determined in field
 2 experiments

3 In the field measurements no correlation between TVOC emissions and construction tempera-
 4 ture was detected and concentrations were 100 to 1000 times lower than in the laboratory ex-
 5 periments. This was attributed to differing meteorological conditions (e.g. wind speeds) re-
 6 sponsible for varying dilution ratios and the fast depletion of these volatile compounds.

7 Accounting for target gas dilution using the tracer gas technique it was possible to calculate
 8 total emissions of TVOC, which again showed a clear correlation with the construction tem-
 9 perature (Figure 5). In the laboratory the TVOC concentrations were comparable for HMA,
 10 WC and WZ, but 50% lower for the two foam asphalt types (Table 3), which is in contrast to
 11 the field results (total emission) and FID-measurements in the laboratory, where significantly
 12 higher values for the HMA were recorded.

13

Table 3. TVOC concentrations in laboratory and field experiments given in $\mu\text{g}/\text{m}^3$ and total emission in g/min

Material	Laboratory	Paver 1	Paver 2	Leeward measurement	Total emission
HMA	57800	700	39	3010	119
WC	57000	369	36	1380	54
WZ	47400	1050	386	837	41
WF	16800	1680	896	1570	41
WFR	23800	577	596	1340	22

1

2 **3.3. Polycyclic aromatic hydrocarbons (PAH)**

3 The very low PAH release (Table 4) was apparently caused by a very low PAH content of the
 4 binders in the bituminous material. This was anticipated for the majority of the mixture types,
 5 since they do not contain any recycled material where tar could be present, which was used as
 6 a bitumen substitute in earlier time. Results were unexpected for WFR with 50% RAP addi-
 7 tion, showing only slightly increased PAH concentrations. Analysis of the applied RAP, how-
 8 ever, revealed that it's PAH content of 25 ppm was rather low for Switzerland, where often
 9 PAH contents of 250 ppm and more are observed when tar is present. Similarly, ben-
 10 zo(a)pyrene, frequently used as a lead substance for toxicity, was below the detection limit.

11 In contrast to TPM and TVOC, and in contradiction to the literature (Hugener et. al.,
 12 2007; Gaudefroy et. al., 2010), PAH emissions did not show a temperature effect. This is at-
 13 tributed to the low PAH values which were always below or close to detection limits. For in-
 14 stance, PAH emissions were not elevated for HMA both in the laboratory and in the field
 15 campaign. Only the warm mix asphalt with 50% RAP showed in general higher PAH-
 16 concentrations. Conclusions should be made with reservations, however, as all measured val-
 17 ues were close to the detection limit. Such inconsistent results at very low PAH concentration
 18 levels were also observed by Rubio et. al. (2013) comparing emissions of HMA and half-
 19 warm asphalt mixes.

Table 4. PAH* concentrations in $\mu\text{g}/\text{m}^3$ during construction and in the laboratory experiments

Material	Laboratory	Paver 1	Paver 2	Lee-measurement	Personal sampler 1-4
HMA	13	bdl	36	31	bdl -34
WC	12	bdl	bdl	bdl	bdl
WZ	14	bdl	bdl	bdl	bdl -35
WF	11	53	34	bdl	bdl
WFR	22	49	46	bdl	bdl -35

1 bdl: below detection limits of $31 \mu\text{g}/\text{m}^3$; *16 PAH according the list of the US Environmental Protection Agency
 2 EPA
 3

4. Conclusions

5 In summary, it can be concluded that pollutant emissions during construction of warm-mix
 6 asphalt were low and below the valid MAK values (Suva, 2018), independent of the asphalt
 7 material, and often below the detection limits of the applied measurement techniques (TSM,
 8 TPM, PAH). This can be partly attributed to the short test fields and consequently short sam-
 9 pling times and small sample gas volumes. In addition, laboratory experiments were conduct-
 10 ed to normalize experimental conditions and increase sensitivity, resulting again in rather low
 11 emission values. For future research projects on diffuse emissions of warm mix-asphalt con-
 12 struction it would be beneficial to increase sensitivity of the applied sampling / analytical
 13 techniques. TVOC concentrations during construction in the field were equivalent to indoor
 14 air, which is not surprising, since pollutant emissions were also low in the laboratory experi-
 15 ments and outdoor emissions are significantly diluted with ambient air. Taking into account
 16 pollutant dilution using the tracer gas technique, TVOC emissions (total emission) during
 17 WMA construction were 2-5 times lower as compared to HMA. This was observed as well in
 18 the laboratory for the total carbon amount (FID signal). However, TVOC values of HMA and
 19 the warm mix asphalt samples WC and WZ were similar. The low TVOC of HMA in the la-
 20 boratory is in contrast to the other data and cannot be explained.

1 Surprisingly, PAH emissions were not affected by temperature, but were all below or
2 close to the detection limit. Low emission levels can be rationalized by low PAH contents in
3 the asphalt mixtures, surprisingly also in the RAP, but the absence of a temperature correla-
4 tion is in contrast to results from earlier research projects on asphalt mixtures with higher
5 PAH contents (Hugener et. al., 2010). Results in the presented study might be impaired by in-
6 sufficient analytical sensitivity or loss of pollutants prior to experiments during preheating of
7 asphalt mixtures in the laboratory.

8 Nevertheless, the presented study demonstrates that reduction of construction tempera-
9 ture is an efficient measure to reduce pollutant emissions. TPM concentrations of warm mix
10 asphalt are reduced by a factor of 10 and TVOC emissions are lower as compared to HMA.
11 Concerning the different WMA types no significant difference was observed in terms of
12 harmful emissions.

13

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17

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