

SORPTION OF NICKEL ON MARL

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Batch sorption experiments using nickel have been carried out on marl, a sedimentary, carbonaceous rock. All experiments were performed with a synthetic water of pH 7.3 and in an atmosphere of N₂/1% CO₂. Over the equilibrium nickel concentration range of 10⁻¹¹-10⁻⁵M, sorption was linear and reversible with R_d of 819 ml g⁻¹. Owing to the linear sorption behavior, R_d was independent of rock/water ratio (r/w = 1/5 - 1/100). The data suggested that at [Ni] lower than 10⁻⁷M an isotope exchange mechanism operated, whereas at higher [Ni], sorption involved ion exchange. Sorption experiments on some of the components of marl indicated that nickel adsorbs mainly on the clay minerals and on the trace of iron hydroxide present; sorption on calcite and quartz was low.

INTRODUCTION

Interaction of nickel with rocks and soils is important both because nickel is a heavy metal pollutant and because the nuclides ⁵⁹Ni and ⁶³Ni may be released from

nuclear waste repositories¹⁻⁴. The mobility of nickel in the environment may be reduced by adsorption on rocks and minerals. This is particularly relevant for waste repository safety analysis, where an essential characteristic of the host rock must be its ability to retard radionuclide migration.

In Switzerland, one particular rock under consideration as a host for a nuclear waste repository is marl, a sedimentary, carbonaceous rock. The present report describes the interaction of nickel with marl. The studies described here, had the specific aim of determining the extent and kinetics of sorption of nickel by marl. In addition, measurements were made to provide an estimate of which components of marl sorbed nickel most strongly.

MATERIALS AND METHODS

Marl was collected from the vicinity of Wellenberg in central Switzerland. The sample consisted of 42% calcite, 18% quartz, 22% illite and illite/smectite mixed layer (illite:smectite = 2.5:1), 7% chlorite, 5% ankerite, 3% kaolinite, 0.3% pyrite, 0.6% iron(III) hydroxide and 0.5% organics. The BET surface area was 9 m². The cation exchange capacity was found by the nickel ethylenediamine method⁵ and was 9.0 meq/100 g. For the sorption experiments, the marl was crushed and ground using a tungsten carbide coated jaw crusher and grinder, respectively. A sample with a particle size <63 µm was obtained by dry sieving.

A synthetic ground water of composition $7.9 \times 10^{-2} \text{ M Na}^+$, $1.2 \times 10^{-3} \text{ M K}^+$, $2.7 \times 10^{-3} \text{ M Mg}^{2+}$, $3.3 \times 10^{-3} \text{ M Ca}^{2+}$, $4.4 \times 10^{-4} \text{ M Sr}^{2+}$, $2.6 \times 10^{-4} \text{ M Si}^{4+}$, $1.4 \times 10^{-4} \text{ M F}^-$,

$8.1 \times 10^{-2} \text{ M Cl}^-$, $4.5 \times 10^{-3} \text{ M SO}_4^{2-}$, $4.4 \times 10^{-3} \text{ M}$ alkalinity and pH 7.3, was used for the sorption experiments. This composition is considered to be typical of that in situ. Before carrying out the sorption experiments, the marl and the marl water were equilibrated until a constant water composition was reached. This was done by holding marl suspensions in dialysis tubing (bags) in containers of synthetic water and renewing the water on a daily basis until chemical analysis showed no further change in composition (7 renewals). This equilibrated water and marl were then used for the adsorption/desorption experiments. Major ions were determined by inductively coupled plasma spectrometry (ICP), trace cations by inductively coupled plasma mass spectrometry (ICP-MS), anions by ion chromatography and alkalinity by acid titration.

The concentration of natural, stable nickel in synthetic marl water was $3.4 \times 10^{-8} \text{ M}$ and in equilibrated marl water, this concentration rose to 10^{-7} M , indicating some leaching of nickel from marl during the equilibration step.

Nickel solutions were prepared with ^{63}Ni tracer and inactive $\text{Ni}(\text{NO}_3)_2 \cdot 6\text{H}_2\text{O}$ in equilibrated marl water. The initial nickel concentrations ranged from $5.7 \times 10^{-9} \text{ M}$ to 10^{-3} M . All the solutions were stable (i.e. no detectable colloid or precipitate formation) over at least 7 weeks. ^{63}Ni was measured using liquid scintillation counting.

The speciation of nickel in synthetic marl water was calculated using the geochemical code Phreeque⁶ and the following stability constants^{1,7}, $\lg K(\text{Ni}(\text{OH})_2) = 9.0$, $\lg K(\text{Ni}(\text{OH})_3^-) = 12.0$, $\lg K(\text{NiCO}_3) = 6.87$, $\lg K(\text{Ni}(\text{CO}_3)_2^{2-}) = 10.1$, $\lg K(\text{NiHCO}_3^+) = 2.14$ and

$\lg K(\text{NiOH}^+) = 4.14$. These calculations indicated that in synthetic marl water, the predominant nickel species over the pH range of 7-9, is NiCO_3 . At the pH of the synthetic water (7.3), nickel was present as 89% NiCO_3 and 11% Ni^{2+} .

All adsorption/desorption experiments were carried out in duplicate and with continuous agitation in an atmosphere of $\text{N}_2/1\% \text{CO}_2$ (appropriate to in situ conditions) in a glove box. A complete series of experiments with a rock/water ratio of 1/10 was carried out with initial $[\text{Ni}]$ from 5.7×10^{-9} – 10^{-3} M. The effect of rock/water ratio (1/5, 1/50, 1/100) was checked with $[\text{Ni}] = 1.1 \times 10^{-7}$ M. The marl suspensions were held in dialysis bags (to facilitate phase separation) suspended in polyethylene flasks of spiked, equilibrated water. Blanks (without marl) indicated that uptake of nickel on the walls of the vessel at pH 7.3 was 3% and more than 15% on the dialysis bags. Nickel uptake on the vessel walls and dialysis bags in the presence of marl was determined by acid leaching (M HCl) at the conclusion of the experiments. In the presence of marl Ni uptake on the vessel walls and the bags was no more than 2%; this was taken into consideration in the calculations of sorption coefficients (R_d).

The kinetics of sorption were followed using a bulk suspension (20 g/200 ml) and a nickel concentration of 1.1×10^{-7} M. The suspension was sampled at intervals; the total volume of all aliquots taken during the experiments was less than 3% of the initial volume. Adsorption coefficients were calculated from the difference between the count rates of the initial and final solution. Once the adsorption coefficients from the kinetic experiments reached the steady state, all other suspensions were sampled and measured.

TABLE 1

Sorption of nickel on some components of marl
(rock/water ratio = 1/10, initial $[Ni] = 1.1 \times 10^{-7} M$,
pH = 7.3)

Components	BET surface area, $m^2 g^{-1}$	R_d , $ml g^{-1}$
Calcite (Merck)	<1	5.7
Quartz (Merck)	<1	1.6
Goethite (FeOOH) (Bayer)	70	560
Kaolinite (Ward Ltd)	19.5	1913
Montmorillonite (Wyoming)	31.5	1109
Illite (Le Puy en Velay, France)	130	226

For the desorption step, the original solution was replaced by tracer free, equilibrated, synthetic marl water. The kinetics of desorption from the bulk sample were followed until a steady state was reached after which all other samples were measured.

The uptake of nickel on the components of marl was carried out with pure materials, not with extracted components because earlier work had indicated that extraction techniques could alter the properties (especially the surface area) of the solids. The minerals used and their sources are listed in Table 1. The samples were equilibrated with synthetic marl water until the water composition was invariant (24-48 h) and the solutions then spiked with ^{63}Ni . The suspensions (r/w = 1/10,

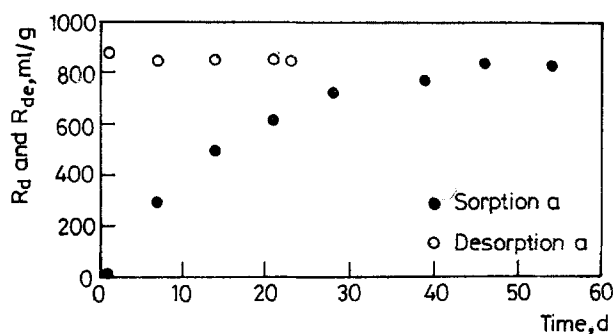


Fig. 1. Kinetics of sorption/desorption of nickel on marl expressed as sorption(R_d) and desorption(R_{de}) coefficients versus time

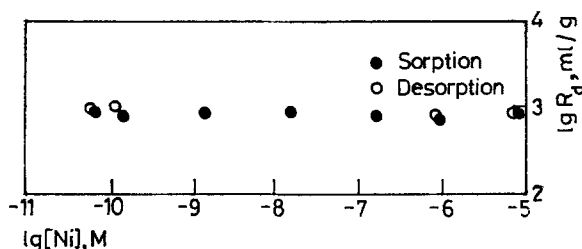


Fig. 2. Change in R_d and R_{de} of nickel on marl with equilibrium nickel concentration in solution

$[Ni] = 1.1 \times 10^{-7} M$) were shaken in the glove box until equilibrium was reached (2 weeks), filtered ($0.45 \mu m$) and the activity of nickel in the filtrate measured.

RESULTS AND DISCUSSION

Figure 1 shows that the uptake of nickel by marl is slow; R_d reached a constant value only after 8 weeks. Desorption, however, was rapid and was complete in one week.

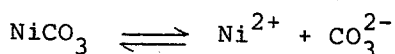
The changes in sorption/desorption coefficients as a function of equilibrium nickel concentration are shown in Figure 2. Adsorption is reversible and linear (i.e.

concentration independent) over the equilibrium nickel concentration range considered (10^{-11} - 10^{-5} M) with an $R_d = 819 \text{ ml g}^{-1}$. The data were fitted to the Freundlich equation; the Freundlich exponent, N , was close to one (0.988) which is consistent with linear adsorption.

The concentration of natural, stable nickel in the equilibrated, synthetic water is 10^{-7} M. It is expected that when the concentration of nickel tracer is below the total $[\text{Ni}]$, an isotope exchange mechanism would operate.

The bulk of nickel in the system was adsorbed by the marl, i.e. both NiCO_3 and Ni^{2+} species appear to have been taken up. This assumption, however, would imply different sorption sites and two different sorption mechanisms; this is not in accord with the observed linear sorption and fast desorption kinetics.

An alternative possibility is that Ni^{2+} is the predominant adsorbing species and the decrease in $[\text{Ni}^{2+}]$ following sorption will shift the equilibrium,



to the right, thus producing a further supply of Ni^{2+} for adsorption. The dissociation step may occur either in solution or, more probably, at the solid/liquid interface. This dissociation reaction may be rate-limiting and hence responsible for the slow adsorption kinetics. On the other hand, adsorbed Ni^{2+} could be readily removed by an ion exchange process and this could account for the fast desorption step.

Sorption of nickel on marl was independent of rock/water ratio ($r/w = 1/5$ to $1/100$) over the entire $[\text{Ni}]$ investigated. This is because, although the equilibrium $[\text{Ni}]$ varied with the rock/water ratio, R_d remained invariant as a result of the linear adsorption (Fig. 2).

This behavior of nickel contrasts with that observed for (non-linear) cesium sorption on marl, where R_d was strongly dependent on the rock/water ratio⁸.

The data in Table 1 indicate that nickel adsorbs most extensively on the clay and iron hydroxide component of marl. As the iron hydroxide makes up only 0.6% of the marl, its contribution to nickel sorption can be expected to be low, unless this phase is present as a coating on the clay particles. The samples of calcite and quartz had far lower surface areas than the clays tested, which may partly account for the low sorption of nickel. However, preliminary tests on the extracted material suggested that in marl these components have a similar low surface area, hence, their contribution to nickel sorption is expected to be low.

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