

1 **Transferability and the effect** **of colour calibration during multi-** 2 **image classification of Arctic vegetation change**

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19 **Acknowledgements**

20 The research was initiated with a WSL internal innovative research grant and supported by a Villum Young Investigator grant
21 (VKR023456), an Aarhus University (AU), Research Foundation grant (AUFF-E-2015-FLS-8-73) and an AU Science and
22 Technology Synergy grant. Fieldwork was possible due to funding from the AU Arctic Research Centre. We are grateful to
23 Rok Kreslin and Peter Peer for providing their code in MATLAB for the DT method; to Constantinos Tsirogiannis for his
24 valuable help in code optimization and to Ditte Grube Barild for geo-referencing the images.

25

This document is the accepted manuscript version of the following article:
Kolyaie, S., Treier, U. A., Watmough, G. R., Madsen, B., Bøcher, P. K., Psomas, A., ...
Normand, S. (2019). Transferability and the effect of colour calibration during multi-image
classification of Arctic vegetation change. *Polar Biology*, 42(7), 1227-1239.
<https://doi.org/10.1007/s00300-019-02491-7>

26 **Abstract**

27 Mapping changes in vegetation cover is essential for understanding the consequences of climate change on Arctic ecosystems.
28 Classification of ultra-high spatial-resolution (UHR, <1cm) imagery can provide estimates of vegetation cover across space
29 and time. The challenge of this approach is to assure comparability of classification across many images taken at different
30 illumination conditions and locations. With warming, vegetation at higher elevation is expected to resemble current vegetation
31 at lower elevation. To investigate the value of classification of UHR imagery for monitoring vegetation change, we collected
32 visible and near infrared images from 108 plots with handheld cameras along an altitudinal gradient in Greenland and
33 examined the classification accuracy of shrub cover on independent images (i.e. classification transferability). We
34 implemented several models to examine if colour calibration improves transferability based on an in-image calibration target.
35 The classifier was trained on different number of images to find the minimum training subset size. With a training set of ~20%
36 of the images the overall accuracy levelled off at about 81% and 68% on the non-calibrated training and validation images,
37 respectively. Colour calibration improved the accuracy on training images (1-4%) while it only improved the classifier
38 transferability significantly for training sets <20%. Linear calibration only based on the target's grey series improved
39 transferability most. Reasonable transferability of Arctic shrub cover classification can be obtained based only on spectral
40 data and about 20% of all images. This is promising for vegetation monitoring through multi-image classification of UHR
41 imagery acquired with hand-held cameras or Unmanned Aerial Systems.

42 **KEY WORDS:** Arctic tundra, climate change, colour calibration, standardization, spectral data, classification transferability

43 **Introduction**

44 The Arctic is warming faster than the rest of the world (Masson-Delmotte et al. 2013). Several studies document recent
45 vegetation changes in response to the increasing temperatures (Tape et al. 2006; Myers-Smith et al. 2011; Elmendorf et al.
46 2012; Myers-Smith et al. 2015; Guay et al. 2015; Nielsen et al. 2017) and model projections highlight the potential for large
47 future changes (Pearson et al. 2013; Normand et al. 2013). Arctic shrub species have been found to increase growth, cover,
48 and height in response to warming, but to varying degrees depending on local environmental conditions (Tape et al. 2006;
49 Elmendorf et al. 2012; Myers-Smith et al. 2015; Nielsen et al. 2017). Increased height and dominance of Arctic shrubs are
50 expected to negatively affect the cover of bryophytes and lichens (Elmendorf et al. 2012), change composition of arthropod
51 communities (Hansen et al. 2016), speed up climate change (Myers-Smith et al. 2015), and lead to profound changes in Arctic
52 ecosystems (Post et al. 2009). Mapping and monitoring changes of Arctic shrub cover is crucial for understanding the spatial
53 magnitude of the potential biodiversity and ecosystem consequences of climate change in the Arctic.

54 Studying changes in vegetation at different spatial and temporal scales is a central challenge in ecology. Fine resolution
55 data are required for studying local changes in vegetation cover (Elmendorf et al. 2012) and for upscaling locally observed
56 patterns across larger areas (e.g., Liu and Treitz 2016). Point framing or visual cover estimation in the field are commonly
57 used methods for providing fine resolution data (Luscier et al. 2006; Liu and Treitz 2016). However, providing these data is
58 either expensive (time/cost; point-frame method) or has reduced reproducibility due to the observer's bias, with an unknown
59 error distribution, which limits the inference of vegetation changes (Neeser et al. 2000; Tichy 2016; Kercher et al. 2003).
60 Vegetation cover estimation using ultra-high spatial-resolution (UHR) images taken by handheld-cameras or Unmanned
61 Aerial Systems is a promising and pragmatic approach that can speed up field data collection (Booth and Cox 2008; Bold et

62 al. 2010; Bricher 2012). Moreover, vegetation cover can be measured by image classification with known accuracy (Lengyel
63 et al. 2008) and the source data can be archived for objectivity and reproducibility of measurements. This can improve our
64 ability for fine-scale vegetation mapping and monitoring (Lengyel et al. 2008; Zlinszky et al. 2015; He et al. 2015) as well as
65 for detailed investigations of vegetation characteristics (Neumann et al. 2015). Here, we seek for an effective and standard
66 processing method to improve field-based observations by UHR images. The goal is to increase comparability of vegetation
67 cover estimates across space and time.

68 Several researchers have used UHR images taken by handheld cameras from vegetation plots and provided
69 measurements that are more reliable compared to field-based observations. They mainly applied object-based image
70 classification to measure ground or vegetation cover (Luscier et al. 2006; Chen et al. 2010; Liu and Treitz 2016). The key
71 aspect for operational use of UHR imagery in ecological field-based studies is the ability to semi-automatically classifying
72 large numbers of images based on reference data collected from the smallest possible number of images. Therefore, we need
73 novel methodologies for the analysis of UHR imagery to obtain vegetation cover and other ecologically relevant parameters
74 efficiently. A pragmatic approach is to train the classifier based on a limited number of images and using that classifier to
75 classify other images (i.e. transferring the classifier). Monitoring vegetation change with this approach can be further
76 challenging because the images are from different locations and different times (hours, days, years). Therefore, vegetation
77 composition, health and life stage, as well as illumination conditions, are likely to vary among images. Vegetation
78 characteristics (species, age, and health condition) and illumination conditions both influence the chromatic outcomes of the
79 vegetation in the images (Jackowski et al. 1997; Villafuerte and Negro 1998; Ritchie et al. 2008; Menesatti et al. 2012; Wang
80 et al. 2013). This leads to high intra-class variation for the classification and make obtaining a representative reference dataset
81 of the images challenging (Gehler and Nowozin 2009).

82 Colour calibration is an important approach to mitigate intra-class variation of reflectance due to the illumination
83 differences among images (Finlayson and Trezzi 2004; Gehler et al. 2008; Wang et al. 2013). Colour calibration has
84 successfully improved image interpretation and analysis for applications in ecology (Villafuerte and Negro 1998),
85 environmental monitoring (Hyman 2010), food science (Quevedo et al. 2010), medicine (Wang et al. 2013), as well as art and
86 museum documentation (Berns et al. 2005). Using colour calibration, the spectral values of the images are converted to
87 standard values using a mathematical model, e.g., polynomial regression models (Wang and Zhang 2010), exponential models
88 (Fischer et al. 2012), or transformation using Delaunay Triangulation (DT) (Kreslin et al. 2014). Defining the models'
89 parameters depends on the relationship between standard values of a calibration target (e.g., Macbeth colour checker,
90 McCamy et al. 1976), placed in the image swath at the image acquisition time, and the values measured from the calibration
91 target on the acquired images. Jackowski et al. (1997) calibrated 20 images based on a Gaussian basis function and the
92 Macbeth colour checker and achieved calibrated images with values closer to the standard values on the calibration target.
93 Polynomial regression models are widely used for colour calibration purposes (Wang and Zhang 2010). Wang and Zhang
94 (2010) calibrated over 300 images for disease diagnosis and showed that a polynomial-based regression provided the best
95 calibration, compared with calibrated images with ridge, support vector, and neural network regressions. Kreslin et al. (2014)
96 tested different colour calibration models on 568 images (each containing a Macbeth colour checker) acquired under different
97 indoor and outdoor illumination conditions. They found that DT-based transformation outperformed other calibration models
98 in producing closer values to the Macbeth colour checker's standard values. The above studies applied colour calibration on
99 imagery including the three visible bands (R, red; G, green; B, blue, hereafter RGB). Near-infrared data are often valuable for

100 vegetation classification and monitoring (Fischer et al. 2012). Using an exponential equation for radiometric calibration of
101 RGB and near-infrared (NIR) images of biological soil crust, Fischer et al. (2012) documented a high linear correlation ($r^2 =$
102 0.91) between estimates of the normalized difference vegetation index (NDVI) from the calibrated images and data obtained
103 from a field spectrometer. Hence, polynomial and DT-based colour calibration models show promise for colour calibration
104 of large RGB image datasets and exponential equations show promise for calibration of NIR images. Nonetheless, while it is
105 documented that colour calibration provides a good standardization of reflectance values across images, the importance of
106 colour calibration for reducing intra-class variation and improving classifier transferability during classification of multiple
107 images remains unknown.

108 Patterns of vegetation composition considerably change along elevational gradients (Engler et al. 2011; Morueta-
109 Holme et al. 2015). With warming, vegetation at higher elevations potentially will become more similar to the current
110 vegetation at lower elevations (Engler et al. 2011; Morueta-Holme et al. 2015). We acquired RGB and NIR images with two
111 handheld cameras from 108 plots distributed across an altitudinal gradient in western Greenland to assess the spatial and
112 temporal classification transferability of UHR imagery for shrub cover quantification. Our overall goal was to examine the
113 effect of colour calibration on the transferability of the classifier and to optimize a multi-image classification framework to
114 automatize monitoring of Arctic vegetation change. Specifically, we addressed the following questions on spectral data: (i)
115 How accurate can we classify images in a multi-image classification framework, (ii) does colour calibration increase
116 classification transferability, and (iii) what is the minimum reference data set for optimising classification transferability.

117 **Materials and methods**

118

119 **Study area and sampling design**

120

121 Digital images of 108 permanent plots (80×80 cm) were sampled from the 21st to 24th of July 2013 in a valley in the inner
122 Nuuk fiord (Latitude: 64.2093; Longitude: -50.2920) (Fig. 1). The plots were distributed stratified random across altitudinal
123 isoclines (at 20, 100, 200, 300, 400, and 500 m a.s.l.). Three groups of six plots were approximately 500 m apart along each
124 isocline and plots were placed 10 m apart within each plot group (for more details on the sampling design see Nabe-Nielsen
125 et al. 2017). Vegetation in the area is composed of a mosaic of several dwarf, low and tall shrub species (*Betula nana*, *Cassiope*
126 *tetragona*, *Dryas integrifolia*, *Empetrum nigrum*, *Ledum groenlandicum*, *Ledum palustre*, *Phyllodoce coerulea*, *Salix glauca*,
127 *Salix arctophila*, *Vaccinium vitis-idaea*, *Vaccinium uliginosum*), graminoids (*Juncaceae*, *Cyperaceae*, *Poaceae*), other herbs,
128 bryophytes, lichens, pteridophytes, and bare ground. In this study, we define shrubs as encompassing dwarf, low, and tall
129 shrubs (cf. Myers-Smith et al. 2015).

130 **Image data**

131

132 We used two handheld single-lens reflex cameras (Canon EOS 550D) to collect the image data. We acquired the visible (VIS)
133 light spectrum with one of the cameras and modified the other one to acquire the near-infrared (NIR) light spectrum (Fig. 1
134 & 3) by replacing the low-pass filter to restrict the cameras sensitivity to wavelengths above 800nm ([http://www.optic-](http://www.optic-makario.de/transmissionskurven/)
135 [makario.de/transmissionskurven/](http://www.optic-makario.de/transmissionskurven/): IR LP2-830nm). Raw images were converted to 8-bit TIFF images by applying the
136 appropriate lens correction model with standard parameters (Adobe Photoshop Camera Raw 6.7). Due to the build in Bayer

137 filter both cameras provided images with 3 bands, hereafter defined for the unmodified (VIS) camera R (red), G (green) and
138 B (blue) bands and for the modified (NIR) camera NIR-R, NIR-G, and NIR-B bands. Four sticks marked each of the corners
139 of the field-plot and allowed for geometric correction. A Macbeth colour checker was placed next to the plot, within the image
140 swath at the image acquisition time. All the images were recorded at about two meters height above the plots from a central
141 nadir position to minimize distortions as much as possible.

142 **Image processing and analyses**

143

144 Our methodology to assess the effect of colour calibration on classifier transferability for Arctic vegetation change studies
145 across the sampled altitudinal gradient had four main parts (Fig. 2): (i) data preparation including geometric correction, (ii)
146 colour calibration, (iii) defining and extracting reference data, and (iv) image classification and accuracy assessment in a
147 multi-image classification framework.

148 **Data preparation and geometric correction**

149 All 108 VIS and 108 NIR images were geo-referenced two times (Fig. 1): (1) relative to the four plot corners (80×80cm) and
150 extracting the plot area image with 2500×2500 pixels, and (2) relative to the cross marks in the corners of the Macbeth colour
151 checker and extracting the *colour checker image* with 570×860 pixels. Both extractions resulted in a ~0.3 mm pixel resolution
152 on the ground. We did the georeferencing in *ArcGIS 10.3.1* (ESRI Redlands, California, USA).

153 **Colour calibration**

154 We implemented colour calibration based on the Macbeth colour checker and 11 different calibration models (Fig. 2). The
155 Macbeth colour checker has been used in several studies using close range photography (McCamy et al. 1976; Jackowski et
156 al. 1997; Kreslin et al. 2014). Reference reflectance values of the 24 colours on the colour checker were based on Ritchie et
157 al. (2008). We extracted the 24 colour values (DN: digital numbers) of each plot area image in six spectral bands (R, G, B,
158 NIR-R, NIR-G, NIR-B), using *Python 2.7.8* (Python Software Foundation, Beaverton, USA). To assess the effect of the Bayer
159 filter on NIR images, we compared the standard deviations of the DN for the NIR-R, NIR-G, and NIR-B bands within
160 sampling grids (i.e., each of the 24 colours). NIR-R had the lowest standard deviation (Online Resource 1) and thus provided
161 the most consistent NIR spectral information. Therefore, only NIR-R was calibrated and used in the classification.

162 First, we did colour calibration on RGB images by implementing 11 different calibration models based on first (1st)
163 and second (2nd) order polynomial regression models, an exponential model, logistic regressions, and the DT-based
164 transformation (Table 1). All calibration models were implemented in *R 3.3.5* (R Development Core Team). Based on DN
165 from the colour checker image, represented as a vector $V: (R_i, G_i, B_i) (i = 1, 2, \dots, 24)$, and the corresponding reference
166 reflectance values as given in Ritchie et al. (2008), represented as $sRGB$, with $S: (sR_i, sG_i, sB_i) (i = 1, 2, \dots, 24)$, the
167 parameters (a) of the calibration algorithms were defined. For example, a simple linear transformation (i.e., 1st order
168 polynomial transformation $x: [R, G, B, 1]$), was formulated as follows:

$$169 \quad sR_i = a_{11}R_i + a_{12}G_i + a_{13}B_i + a_{14}$$

$$170 \quad sG_i = a_{21}R_i + a_{22}G_i + a_{23}B_i + a_{24}$$

$$171 \quad sB_i = a_{31}R_i + a_{32}G_i + a_{33}B_i + a_{34} \quad (1)$$

172	<i>Where:</i>	
173	sR, sG, sB	Reflectance values from Ritchie et al. (2008)
174	R, G, B	Digital numbers extracted from the in-image colour checker
175	$(i = 1, 2, \dots, N)$	Fields on the Macbeth colour checker ($N = 24$)

176 We implemented three 1st order polynomial transformations (M1 – M3, Table 1). Four 2nd order polynomial transforms (M4
 177 – M7) were used to increase the transformation accuracy. For M2 and M4 only corresponding bands were used for the
 178 calibration (Table 1). M5 was highly parametrized and resulted in overfitting and false colours occurred. Furthermore, we
 179 implemented two logistic regression models with sigmoid curves with exponential growth (M8 and M9) and an exponential
 180 regression model (M10) (Table 1). Finally, DT was implemented in R by converting the MATLAB code of Kreslin et al.
 181 (2014).

182 We first applied all the colour calibration algorithms on RGB images to compare the values of the colour checker from the
 183 calibrated images with the reference values and with the values from the non-calibrated images. Four calibration methods
 184 (M2, M8, M9 and DT) resulted in colour values closer to the reference values compared to the non-calibrated images and did
 185 not change the natural colour space (Online Resource 2). We therefore selected these four calibration models and calibrated
 186 the NIR band based on the NIR₈₀₀₋₉₀₀ reflectance values from Ritchie et al. (2008) (Table 1). Since the DT model cannot be
 187 applied to only one band, we here used the sNIR values obtained with M8.

188 Images calibrated with these four models were taken forward for the classification. Moreover, since the spectral reflectance
 189 is almost constant across wavelengths for the grey colours compared to other colours (see Fig.4 in Berns et al. 2005), we also
 190 calibrated all images with only the grey colours of the colour checker with M2, M8 and M9 (hereafter M2_g, M8_g and
 191 M9_g). Hence, in total eight image data sets were used for classification, including the non-calibrated images.

192 **Reference data preparation**

193 For each plot, we created reference polygons for the following four cover classes: shrub, other vegetation cover (i.e.,
 194 graminoids, pteridophytes, lichens, bryophytes, and herbs), other cover classes (i.e., markings), and ground (including bare
 195 ground and stones). All reference polygons were drawn by the same person to reduce observer bias. We extracted spectral
 196 values per pixel of each image within the defined polygons and used them as the reference dataset. On average, 105,328 pixels
 197 of 6.25MP = 1.7% (min = 0.5%, max = 9%) were selected per image, and on average 41,095 (39%) of these pixels were the
 198 shrub class. The process was automated in R 3.3.5, and was done for each of the 108 plots in each of the eight calibrated
 199 image sets. The large reference data set provides a unique opportunity for assessing classification transferability for Arctic
 200 vegetation change studies using images stratified randomly across almost 500 altitudinal meters.

201 **Multi-image classification**

202 The four cover classes were classified using random forest classification with four (R, G, B and NIR) parameters. We
 203 implemented pixel-based classification, as the aim was to investigate the effect of the colour calibration on spectral
 204 classification transferability and not to obtain the most accurate classification of each image. We used the random forest
 205 classifier because of its robustness (Rodriguez-Galiano et al. 2012).

206 The following steps were taken to investigate the classifier's transferability, i.e., to what extent the classifier can be applied
207 to other images beyond a training subset, and to find the optimum size of reference data regarding classification transferability.
208 We trained the random forest classifiers on randomly selected portions of images (5, 10, 15, 20, 25, 30, 35, 40, 45, 50, 60 and
209 70%, hereafter training subset) and subsequently applied the classifier on the remaining images (remaining subset, Fig. 2).
210 Reference data of the images in training subsets were merged and used to train the classifier. The remaining subsets included
211 all the images in each image set except the images of training subsets. Moreover, for training portions $\leq 50\%$, we selected a
212 random subset with the equal number of images as in the specific training subset. This procedure was implemented to examine
213 if the overall trend of accuracy could be captured in smaller datasets. We repeated the classification 10 times for each portion,
214 to assess the classification transferability regardless of the specifications of training subsets (Fig. 2). We repeated the whole
215 procedure for the non-calibrated images as well as the seven selected calibration methods (M2, M2_g, M8, M8_g, M9, M9_g,
216 DT) to assess the effect of each calibration method on the transferability of the classifier.

217 To assess the degree of transferability, we calculated overall accuracy (OA), the Kappa coefficient, and the user's (UA) and
218 the producer's (PA) accuracies per class for each classified image (Foody 2002). Classification accuracies of the remaining
219 subsets compared to the classification accuracies of training subsets were used to assess the classifier's transferability. As a
220 benchmark for classification accuracy, we also implemented single image classification on each non-calibrated image.
221 Accuracy of the single image classification was preformed based on training and testing data obtained from the same image
222 to which the classifier was applied. We assessed if calibration methods significantly improved transferability of the classifier
223 by performing Dunnett's test using the 'DescTools' package with the non-calibrated image set as control group (Signorell et
224 al. 2017).

225 **Results**

226 The average (\pm standard deviation) OA and kappa for single image classification of the four ground cover classes across all
227 non-calibrated images was 93% ($\pm 3\%$) and 88% ($\pm 6\%$), respectively. In the multi-image classification framework, OA on the
228 training subsets decreased to 78% ($\pm 8\%$) with increasing size of the training subset (Fig. 4). OA on the remaining subsets
229 increased with increasing subset size and reached average OA of 68% ($\pm 12\%$) and 72% ($\pm 11\%$) when respectively 20% and
230 70% of the non-calibrated images were used for training (Fig. 4, 5). A similar trend was observed on the testing subsets. In
231 general, classification OAs are levelling off with a training set of ca. 20%, with a similar trend for all the image sets (calibrated
232 and non-calibrated) (Fig. 4, 5). OA variations in the remaining subsets decreased with increasing sizes of the training subsets;
233 with the highest variation for portion=5%. For shrub cover average UA and PA of 62% and 68% where reached when 20%
234 of the non-calibrated images were used for training . For the non-shrub vegetation class these values were 61% and 60%,
235 respectively, while they were only 28% and 39% for the ground cover class.

236 Training subset OA based on all the colour calibrated image-sets were higher than OA on the non-calibrated image-set (1-
237 4%, Fig. 4). However, colour calibration only slightly increased transferability of the classifier for small subset sizes ($\leq 20\%$,
238 Fig. 4) and only M2_g and M8_g had significant positive effect on classification transferability (Fig. 5). M9 and M9_g had a
239 significant negative effect on classification transferability (Fig. 5). These trends were also captured on testing subsets.
240 Similarly, in relation to shrub cover classification, M2_g and M8_g significantly increased shrub cover class UA for remaining
241 subsets when assessed across all portions (Fig. 6). However, all other calibration models, except M2_g and DT, decreased

242 shrub cover class producer's accuracies significantly (Fig. 6). When 20% of the images were used for calibration with M2_g,
243 average UA for the shrub class improved with 1.9% (relative to non-calibrated data), while only PA improved (2.6%) for the
244 non-shrub class. Both UA (2.1%) and PA (7.9%) increased for the ground class. Similar results were obtained with M8_g and
245 a subset size of 20%; here PA for the ground class increased by 9%.

246 **Discussion**

247 In ecological field-based studies, researchers estimate vegetation cover visually or with point framing, for analysis and
248 monitoring of vegetation change at fine scale (Luscier et al. 2006; Liu and Treitz 2016). This method, however, is time
249 consuming, might be biased, and provides only limited data for upscaling (Neeser et al. 2000; Kercher et al. 2003; Rose et al.
250 2015; Tichy 2016). Mitigating these challenges was the main motivation to use UHR imagery for vegetation cover estimation.
251 However, usage of UHR imagery requires classification of a large number of images taken at different locations and times
252 (Gehler and Nowozin 2009; Cimpoi et al. 2014). The main challenge is increased intra-class variability (e.g., due to varying
253 illumination condition, various vegetation characteristics) which might reduce classification accuracy and makes selection of
254 a representative reference dataset difficult (Gehler and Nowozin 2009). We assessed the value of UHR imagery for vegetation
255 cover estimation by testing the classification transferability during a multi-image classification of images taken stratified-
256 random across almost 500 altitudinal meters in Western Greenland. Our findings show reasonable transferability of Arctic
257 shrub cover classification, with average overall accuracy of $68\% \pm 12\%$ on independent images when 20% of the images were
258 used to parametrize the classifier. This relatively good transferability based only on spectral data is promising. It illustrates,
259 that monitoring of vegetation cover with UHR imagery is achievable, not only for images taken under varying field conditions,
260 but also for images covering the range of vegetation class variation, which is expected under future climate change.

261 The training subset size affected classification transferability. The aim was to find the smallest possible random subset of
262 images, assuring reliable training of the classifier and minimizing the time spend on creating reference data. Reference data
263 were created by delimitation of polygons for each of the targeted vegetation classes. As expected, increasing the training
264 subset size improved the transferability of the classifier (i.e., classification accuracy on remaining subsets) (Fig. 4). By using
265 more images for training, classification transferability increases as the classifier recognizes more variation of each class due
266 to different species, shadows, age, and health as well as (mitigated) illumination effects. However, due to increased intra-class
267 variation the overall accuracies of classification on training subsets decreased with the increasing number of training images
268 (Fig. 4). Classification accuracy tends to level off when about 20% of all images are included in the training set and this trend
269 is similar for all image sets (calibrated and non-calibrated). Therefore, we concluded that about one fifth of the images would
270 possibly be the optimum size for a training subset to provide an image classifier that is transferrable to all the images.

271 All calibration models improved classification accuracies on the training subsets (Fig. 4). Even though colour calibration
272 slightly increased transferability of the classifier for small subset sizes ($\leq 20\%$, Fig. 4), only M2_g and M8_g significantly
273 improved transferability of the classifier for small subsets (portions $\leq 15\%$, Fig. 5). However, for bigger sizes of training
274 subsets ($>20\%$), overall accuracies of the calibrated image-sets and non-calibrated image-set were similar. This shows that
275 the classifier possibly captured most of the variation of illumination effects when a random subset of at least 20% of the
276 images was used to train the models. Transferability of the classification for the shrub, non-shrub, and ground class increased
277 more with calibration models that only included the grey scales of the calibration target. Transferability of each of the three

278 classes increased 2-9% with these colour calibrations when only 20% of the images were used for training. This underlines
279 that colour calibration is important for maximizing transferability when small portions of the data are used for training, but
280 also that its importance depends on the cover class of interest.

281 Images calibrated with M9_g had the highest classification accuracy on training subsets, compared to the other image-sets
282 (Fig. 4). However, M9 and M9_g had the lowest classification accuracy on the remaining subsets (Fig. 4, 5). This behaviour
283 might be explained by M9 models having a lower dynamic range compared to other models, due to the model specification.
284 In addition, although the DT calibration model enhanced the images best for visualization purposes (Kreslin et al. 2014), it
285 did not improve the transferability of the pixel-based random forest classifier. These results show that different calibration
286 methods could be useful for different applications. Importantly the increase in accuracy (1-4%) on our training data documents
287 that colour calibration is important when classification is performed on one or few images where reference data is available for
288 all images.

289 Colour calibration is one approach to mitigate intra-class variation in reflectance due to illumination differences among images
290 (Finlayson and Trezzi 2004; Gehler et al. 2008; Wang et al. 2013). Another approach is using spatial signatures (like texture
291 and shape); these measures are less sensitive to illumination variation (Gehler and Nowozin 2009; Johansen et al. 2014). In
292 recent studies using high-resolution imagery, object-based classification methods provided more accurate results than pixel-
293 based classification methods (Whiteside et al. 2011). Because we aimed at a fully objective classification approach, which
294 minimized user decisions and optimized time efficiency, we here applied a pixel-based classification method. However,
295 integrating texture measures in a pixel-based classification is likely to improve the classification accuracy.

296 Mapping and monitoring changes in Arctic shrub cover is crucial for understanding the spatial magnitude of the potential
297 biodiversity and ecosystem consequences of climate change in the Arctic. Efficiently obtaining fine-scale ground truthing
298 information of vegetation cover is especially important in the Arctic due to the short field season and the logistical challenges
299 related to cover large areas during one field campaign. Classification of UHR images show promise for providing comparable
300 estimates of vegetation cover across space and time. Moreover, such remotely sensed data can improve, add and speed up the
301 traditional field-based data collection (Neeser et al. 2000; Luscier et al. 2006; Lengyel et al. 2008; Fischer et al. 2012; Tichy
302 2016) and provide fine-scale ground truth data which in combination with satellite-based remote sensing will enable upscaling
303 of fine scale observations across larger areas (Liu and Treitz 2016).

304 Conclusion

305 The goal of this study was to investigate the effect of training data size and colour calibration on transferability of a pixel-
306 based classification. Here, for shrub cover estimation, a simple linear model (M2) based on the grey series of the calibration
307 target worked better than the other models. A random selection of 20% of all images was the optimal size for the training
308 subset. The transferability of the classifier with an overall classification accuracy of about 70% is promising for the use of
309 UHR imagery to assist field-based ecological studies. These results are useful for automating Arctic vegetation monitoring.
310 Further improvement of classification accuracy might be reached by including spatial signatures in the classification.

311 **Compliance with ethical standards**

312

313 **Conflict of interest**

314 The authors declare that they have no conflict of interest.

315 **Figure legends**

316 **Figure 1** Study area, sampling design and example of geometric correction. (a) Location of the study area in the inner
317 Nuuk Fiord, western Greenland. Vegetation classification based on Karami et al. (2018). (b) Distribution of the 108
318 vegetation plots across altitudinal isoclines within the study area. The vegetation plots are distributed in groups of six
319 plots (inlet). The distance between plot groups was 500 m and distance between each of the six plots 10 m. (c) From each
320 image the plot area (80×80 cm) and the colour checker area were extracted as separate images and (e) geometrically corrected
321 (see text for details).

322 **Figure 2** Framework of the applied data processing steps. VIS: visible light spectrum, NIR: near-infrared light spectrum, and
323 R: Red, G: Green, B: Blue bands of the VIS spectrum. OA: overall classification accuracy, UA: user accuracy, and PA:
324 producer accuracy.

325 **Figure 3** Two examples of single image classification results with the pixel-based random forest classifier on non-calibrated
326 data: Left: NIR-RG (near infrared, red, green), and Center: RGB (red, green and blue) images of the 80x80 geometrically
327 corrected plots. Right: Classified images for shrub, non-shrub, ground, and other cover classes,

328 **Figure 4** Relationship between overall accuracy and the proportion of data used for training. Loess-smoothed overall
329 accuracies (mean ± standard deviation) for training and remaining subsets is plotted against the portion of images used for
330 training the classifier for different calibration models; Delaunay triangulation (DT), 1st order polynomial (linear, M2) and
331 exponential (M8, M9) with all the colours from the Macbeth colour checker or only with grey series (M2_g, M8_g, M9_g).

332 **Figure 5** Effect of the calibration models on classification transferability of all ground classes. Each block shows classification
333 overall accuracies (OA) of different portions of images used for training the classifier. Images were either non-calibrated data
334 (RD) or calibrated with different implementations of calibration models: Delaunay triangulation (DT), 1st order polynomial
335 (linear, M2) and exponential (M8, M9) with all the colours from the Macbeth colour checker or only with grey series (M2_g,
336 M8_g, M9_g). We assessed if calibration methods significantly improved transferability of the classifier by performing a
337 Dunnett's test with the non-calibrated image set as control group.

338 **Figure 6** Effect of the calibration models on classification transferability of the shrub class. Users' and producers' accuracies
339 (UA and PA) computed as averages across all portions of the training images. Images were either non-calibrated data (RD)
340 or calibrated with different implementations of calibration models: Delaunay triangulation (DT), 1st order polynomial (linear,
341 M2) and exponential (M8, M9) with all the colours from the Macbeth colour checker or only with grey series (M2_g, M8_g,
342 M9_g). We assessed if calibration methods significantly improved transferability of the classifier by performing a Dunnett's
343 test with the non-calibrated image set as control group.

344

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